



Australian Government
**Department of Climate Change
and Energy Efficiency**

AUSTRALIAN NATIONAL GREENHOUSE ACCOUNTS



National Inventory Report 2009

Volume 1

The Australian Government Submission to the
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thinkchange



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EXECUTIVE SUMMARY

ES.1 BACKGROUND INFORMATION ON GREENHOUSE GAS INVENTORIES

This is Australia's *National Inventory Report 2009* submitted under the United Nations Framework Convention on Climate Change (UNFCCC) and the Kyoto Protocol (KP). The *National Inventory Report 2009* contains national greenhouse gas emission estimates for the period 1990-2009 compiled under the rules for reporting applicable to the UNFCCC. It also contains estimates for 2008 and 2009 compiled using the reporting rules applicable to the Kyoto Protocol.

The report has been prepared in accordance with the UNFCCC *Reporting Guidelines on Annual Inventories* agreed by the Conference of Parties at its eighth (decision 18/CP.8), ninth (decision 13/CP.9) and eleventh sessions (decision 14/CP.11), and set out in document FCCC/SBSTA/2006/9 and the supplementary reporting requirements under Article 7 of the Kyoto Protocol (Decision 15/CMP.1).

The estimates of emissions contained in this report have been compiled using methods which conform to the international guidelines adopted by the UNFCCC (IPCC 1997, 2000 and 2003). The methodologies used to estimate Australia's inventory have been improved over time and will continue to be refined as new information emerges and as international practice evolves. The impact on greenhouse gas emission estimates of refinements to methodologies adopted for this inventory has been reported in chapter 10 (Volume 3).

The responsibility for Australia's greenhouse emissions reporting has been assigned to the Department of Climate Change and Energy Efficiency (DCCEE). The DCCEE undertakes all aspects of activity data co-ordination, emissions estimation; quality control, the preparation of the reports and their submission to the UNFCCC on behalf of the Australian Government.

In addition to this report, the DCCEE publishes a range of supporting emission estimates that, together, constitute the *Australian National Greenhouse Accounts*, including:

- the *National Greenhouse Gas Inventory* which provides a summary of the Kyoto Protocol account;
- an overview of the *State and Territory Greenhouse Gas Inventories*;
- the *National Inventory by Economic Sector*, comprising emission estimates by economic sector rather than by IPCC sectors as in this report; and
- *Quarterly Update of Australia's National Greenhouse Gas Inventory*, which provides the most recent estimates of Australia's National Inventory.

These documents are available on the DCCEE website at www.climatechange.gov.au/en/climate-change/emissions.aspx. They provide additional information with respect to Australia's emissions on both a regional and industry basis.

ES.2 SUMMARY OF THE NATIONAL EMISSION AND REMOVAL RELATED TRENDS

ES.2.1 Greenhouse Gas Inventory – UNFCCC

Australia's total greenhouse gas emissions, excluding the *land use, land use change and forestry* (LULUCF) sector, were estimated to be 545.8 million tonnes (Mt), CO₂-equivalent (CO₂-e) in 2009. This represents a decrease of 5.0 Mt or 0.9 % on net emissions recorded in 2008 and an increase of 30.5 % (127.5 Mt) above 1990 levels (Table ES.1 and Figure ES.1).

Australia's total greenhouse gas emissions, including net emissions from the LULUCF sector, were 599.8 Mt CO₂-e in 2009. Overall, total emissions have increased by 138.3 Mt or 30.0% on net emissions recorded in 1990. While in 2009 the LULUCF sector was a net source of 54.0 Mt, the LULUCF sector may, from year to year, change from a net source to a net sink (ranging from a source of 342.5 Mt to a sink of 195.3 Mt over the 1990-2009 period).

Table ES.1: Australia's net greenhouse gas emissions by sector under the UNFCCC

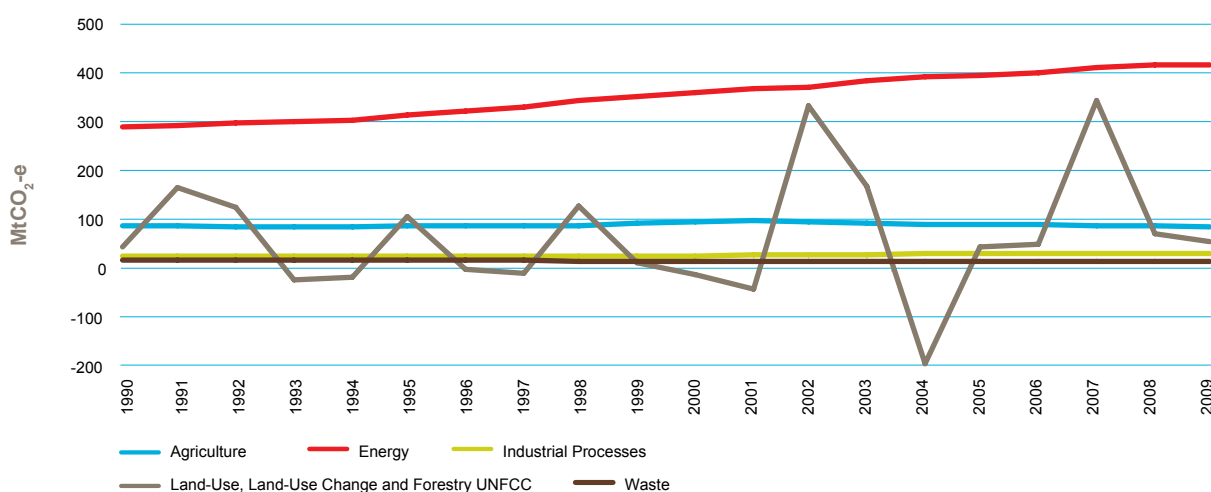
Sector and Subsector	Emissions Mt CO ₂ -e			Per cent change
	1990	2008	2009	1990-2009
1 All energy (combustion + fugitive)	289.0	417.6	417.4	44.4
<i>Stationary energy</i>	194.6	293.4	294.1	51.1
<i>Transport</i>	62.1	84.2	83.6	34.6
<i>Fugitive emissions from fuel</i>	32.3	39.9	39.7	22.9
2 Industrial Processes	24.5	31.2	29.6	21.0
3 Solvent and other product use ^(a)	IE	IE	IE	IE
4 Agriculture	86.8	87.9	84.7	-2.4
6 Waste	18.0	14.1	14.1	-21.9
Total net emissions (excluding LULUCF)	418.3	550.8	545.8	30.5
5 Land use, land use change and forestry	43.1	69.5	54.0	25.1
Total net emissions (including LULUCF)^(b)	461.5	620.3	599.8	30.0

(a) Included elsewhere – emissions are included in industrial processes for confidentiality reasons.

(b) These totals and trend do not reflect Australia's emissions and removal under the Kyoto Protocol accounting framework.

The high variability in emissions and removals in this sector is the result of the requirement to report all emissions and removals on managed land combined with the use of Tier 3 methods that reflect the impacts of both management and climate on emissions/removals. As a result, the trends in the LULUCF sector emissions are driven primarily by inter-annual climate variability and natural disturbance which tend to mask other, underlying patterns in the sector which are directly associated with human activities.

Figure ES.1: Australia's net greenhouse gas emissions by sector from 1990 to 2009



ES.2.1.1 Major Inventory Developments and Recalculations

The inventory is developed in accordance with the National Inventory Improvement Plan. Some of the major method improvements that have resulted in recalculations to the national inventory are listed below:

- Input data to support the preparation of the accounts for important elements of the energy, industrial processes and waste sectors are collected using the National Greenhouse and Energy Reporting System (NGERS). The first reports submitted by companies covered the Australian financial year 2008-09 and this data has been widely used in the preparation of the accounts for the first time. A decision making framework, with respect to the use of facility specific emission factors, was developed to ensure consistent application of facility level data.

- A major study was undertaken that involved directly measuring emissions from 347 in-service vehicles. The study produced updated emission factors and deterioration rates for light-duty petrol vehicles manufactured since 1994.
- Revision to land use change and forestry emission estimates due to corrections identified through enhanced quality control processes.

Overall the net effect of these recalculations was a reduction of 3.0 Mt CO₂-e (0.7%) in 1990 and a net increase of 2.3 Mt CO₂-e (0.4%) in 2008.

ES.2.2 Kyoto Protocol

Australia's initial assigned amount under the Kyoto Protocol is equivalent to 591.5 Mt CO₂-e for each year of the commitment period (2008-2012). The initial assigned amount is fixed and is the value that Australia's Kyoto Protocol emissions account will be compared with for the first commitment period.

- Under the KP accounting rules Parties must report emissions from the Article 3.1 (Annex A) sectors, namely the energy, industrial processes, solvents, agriculture and waste sectors. However, only a subset of the LULUCF sector emissions and removals are included in the account. These include deforestation, afforestation and reforestation activities that have occurred since 1990 (Article 3.3 activities which are mandatory) and forest, cropland or grazing land management and revegetation (Article 3.4 activities which can be elected by the Party). Australia has not elected to account for any Article 3.4 activities for the first commitment period.
- The total net emissions associated with the KP account for Article 3.1 sectors and Article 3.3 LULUCF activities were 579.6 Mt CO₂-e in 2008 and 564.5 Mt CO₂-e in 2009. There have been no transfers of KP units from international sources under Articles 6, 12 and 17.

Table ES.2: Emissions and removals associated with Articles 3.1 and 3.3 of the Kyoto Protocol, 2008 and 2009

Sector and Subsector	2008 emissions Mt CO ₂ -e	2009 emissions Mt CO ₂ -e
Article 3.1 – Annex A sectors		
1 Energy	417.6	417.4
2 Industrial Processes	31.2	29.6
3 Solvent and other product use ^(a)	IE	IE
4 Agriculture	87.9	84.7
6 Waste	14.1	14.1
Inventory Total	550.8	545.8
Article 3.3 – KP LULUCF activities ^(b)		
Afforestation/Reforestation ^(c)	-23.7	-22.6
Deforestation	52.5	41.3
Kyoto Protocol Total	579.6	564.5

(a) Included elsewhere – emissions are included in *industrial processes* for confidentiality reasons

(b) Australia has elected to account for Article 3.3 activities on an annual basis

(c) Accounting quantity in accordance with paragraph 4 of annex to decision 16/CMP.1

While this report is primarily focused on estimates for the 1990-2009 period preliminary estimates for 2010 have been published in the Quarterly update. The preliminary estimates for emissions from Annex A sectors in 2010 is 545 Mt CO₂-e. A significant reduction in emissions from electricity generation (6 Mt CO₂-e) is estimated in 2010. This estimated reduction is due to fuel switching and slightly lower demand, in part driven by unusually high temperatures in the winter months. This decrease was offset by increases in emissions from industrial processes and energy sectors other than electricity generation. These sectors continued to return to higher activity levels after the effects of the economic slowdown in 2009. While data are not yet available, deforestation emissions in 2010 are projected to be 49 Mt,

reflecting an expected rebound in emissions from 2009. Deforestation activity is difficult to predict but appears to be influenced by commodity prices, recent weather patterns and the impact of State government clearing restrictions.

A full report on the holdings and transactions of Kyoto units from the national registry and the calculation of the KP-LULUCF accounting quantity are provided in chapter 12.

ES.3 OVERVIEW OF SOURCE AND SINK CATEGORY EMISSION ESTIMATES AND TRENDS

ES.3.1 Greenhouse Gas Inventory

The *energy* sector was the largest source of greenhouse gas emissions in 2009 comprising 76.5% (417.4 Mt) of total net emissions (excluding LULUCF). *Energy* emissions increased by 44.4% between 1990 and 2009 and decreased by 0.1% between 2008 and 2009 (Table ES.2).

For the *energy* subsectors in 2009:

- stationary energy was the main contributor to total net emissions (53.9%) and increased by 0.2% between 2008 and 2009;
- transport emissions (15.3% of total net emissions) decreased by 0.7% between 2008 and 2009;
- fugitive emissions from fossil fuels (7.3% total net emissions) decreased by 0.7% between 2008 and 2009.

Industrial processes made up 5.4% (29.6 Mt) of the total net emissions (excluding LULUCF) for 2009 and decreased 5.0% between 2008 and 2009.

Agriculture emissions made up 15.5% (84.7 Mt) of total net emissions (excluding LULUCF) in 2009 and decreased by 3.6% between 2008 and 2009.

The *waste* sector contributed 2.6% (14.1 Mt) of the total net emissions (excluding LULUCF) in 2009 and decreased by 0.3% between 2008 and 2009.

The *land use, land use change and forestry* sector was a net source of 54.0 Mt in 2009 and decreased by 22.3% between 2008 and 2009.

A full overview of emission estimates by source and sink is given in chapter 2. More detailed information on the emission results for individual sectors has been reported in the introduction to chapters 3 – 6.

ES.3.2 KP-LULUCF Activities

In 2009 *afforestation/reforestation* activities on unharvested land units contributed a net sink of 22.6 Mt CO₂-e while harvested land units contributed a net source of 7.6 Mt CO₂-e. *Deforestation* activities were a net source of 41.3 Mt CO₂-e in 2009.

Australia has elected to account for the Article 3.3 activities on an annual basis. In accordance with the harvested forest sub-rule (paragraph 4 of the annex to decision 16/CMP.1) in calculating the *afforestation/reforestation* accounting quantity for 2008 and 2009 the emissions on harvested land units have been set to zero where the emissions are greater than the removals on those lands units.

ES.4 OTHER INFORMATION (EG INDIRECT GREENHOUSE GASES)

The National Inventory Report also presents estimates of emissions of indirect greenhouse gases, including NO_x, CO, NMVOC and SO₂. For these gases, no Global Warming Potentials are available to enable their conversion to carbon dioxide equivalent emissions. Consequently these emissions sources are not able to be aggregated, nor are they covered by the reporting provisions of the Kyoto Protocol. Estimates of Australia's emissions from NO_x, CO, NMVOC and SO₂ are presented in Chapter 2.

ACKNOWLEDGEMENTS

The Department of Climate Change and Energy Efficiency wishes to acknowledge the many individuals and organisations that contributed to the development of the national methods over the years. Although the list of all who have provided input is too long to include here, the Department of Climate Change and Energy Efficiency would like to thank the following who have contributed significantly to the preparation of emission estimates for this inventory:

Sector	Contributor	Organisation
Energy	Mr Steven Oliver	DCCEE
	Mr Glen Whitehead	DCCEE
	Mr Nicholas Giles	DCCEE
Industrial processes and solvents	Mr Mark Hunstone	DCCEE
	Ms Nicola Gabay	DCCEE
	Dr Jeff Deslandes	EnerGreen Consulting
	Dr Dean Thomson	Net Balance Pty Ltd
Agriculture	Dr Mick Meyer	CSIRO Marine & Atmospheric Research
	Ms Penny Reyenga	DCCEE
Land Use, Land Use Change and Forestry	Mr Rob Waterworth	DCCEE
	Ms Nikki Fitzgerald	DCCEE
	Mr Rob De Light	DCCEE
	Ms Amanda Twomey	DCCEE
		Australian Bureau of Agricultural and Resource Economics and Sciences Australian Bureau of Meteorology Australian National University CSIRO Land and Water CSIRO Mathematics, Informatics and Statistics CSIRO Sustainable Ecosystems GeoScience Australia
Waste	Mr Mark Hunstone	DCCEE
	Ms Nicola Gabay	DCCEE
	Dr Dean Thomson	Net Balance Pty Ltd

National Greenhouse Gas Inventory Committee

Rob Sturgiss	Australian Government
Ian Galbally	CSIRO
Julian Smith	Victoria
Jessica Lerch	Tasmania
Jarrold Abrahams	Western Australia
Matt Riley	New South Wales
Rod Goodbun	Queensland
David Finlay	South Australia
Janice Warren	Northern Territory
Paul Sutton	Australian Capital Territory

GENERAL NOTES

Units

The units mainly used in this inventory are joules (J), grams (g), tonnes (t), metres (m) and litres (l), together with their multiples. Standard metric prefixes used in this inventory are:

kilo (k) = 10^3 (thousand)

mega (M) = 10^6 (million)

giga (G) = 10^9

tera (T) = 10^{12}

peta (P) = 10^{15}

Emissions are generally expressed in gigagrams (Gg) in the Inventory tables, as called for under international guidelines, and in megatonnes (Mt) in the text of the inventory report:

gigagram (Gg) = 1,000 tonnes = 1 kilotonne (kt)

megatonne (Mt) = 1,000,000 tonnes = 1,000 Gg

Gases

CF ₄	perfluoromethane (a perfluorocarbon)
C ₂ F ₆	perfluoroethane (a perfluorocarbon)
CH ₄	Methane
CO	carbon monoxide
CO ₂	carbon dioxide
HFCs	hydrofluorocarbons
N ₂ O	nitrous oxide
NMVOC	non-methane volatile organic compounds
NO _x	oxides of nitrogen
PFCs	perfluorocarbons
SF ₆	sulphur hexafluoride
SO ₂	sulphur dioxide

Global Warming Potentials

CO ₂ = 1	HFC-23 = 11,700
CH ₄ = 21	HFC-125 = 2,800
N ₂ O = 310	HFC-134a = 1,300
CF ₄ = 6,500	HFC-143a = 3,800
C ₂ F ₆ = 9,200	SF ₆ = 23,900

Conversion Factors

From element basis to molecular mass

C → CO₂: $\times 44/12 = 3.67$

C → CH₄: $\times 16/12 = 1.33$

N → N₂O: $\times 44/28 = 1.57$

From molecular mass to element basis

CO₂ → C: $\times 12/44 = 0.27$

CH₄ → C: $\times 12/16 = 0.75$

N₂O → N: $\times 28/44 = 0.64$

Indicators

In the tables, the following standard indicators are used:

NO	(not occurring) when the activity or process does not occur in Australia
NA	(not applicable) when the activity occurs in Australia but the nature of the process does not result in emissions or removals
NE	(not estimated) where it is known that the activity occurs in Australia but there are no data or methodology available to derive an estimate of emissions
IE	(included elsewhere) where emissions or removals are estimated but included elsewhere in the inventory
C	(confidential) where reporting at a disaggregated level could lead to the disclosure of confidential information

ABBREVIATIONS

ABARES	Australian Bureau of Agricultural and Resource Economics and Sciences
ABARE	Australian Bureau of Agricultural and Resource Economics
ABS	Australia Bureau of Statistics
ADO	Automotive Diesel Oil
ADRs	Australian Design Rules
AEMO	Australian Energy Market Operator
AGEIS	Australia Greenhouse Emissions Information System
ANZSIC	Australia New Zealand Standard Industrial Classification
APPEA	Australian Petroleum Production and Exploration Association
AVHRR	Advanced Very High Resolution Radiometer
BoM	Bureau of Meteorology
BTX	Benzene, Toluene, Xylene
BRS	Bureau of Rural Science
COBIT	Control Objectives for Information and related Technology
CSIRO	Commonwealth Scientific and Industrial Research Organisation
CRF	Common reporting format
DCCEE	Department of Climate Change and Energy Efficiency
DSEWPC	Department of Sustainability, Environment, Water, Population and Communities
DRET	Department of Resources, Energy and Tourism
ERT	Expert Review Team
ESAA	Energy Supply Association of Australia
ETS	Emissions Trading Scheme
GEDO	Greenhouse and Energy Data Officer
IDF	Industrial Diesel Fuel
IEF	Implied Emission Factor
IPCC	Intergovernmental Panel on Climate Change
KP	Kyoto Protocol
LPG	Liquid Petroleum Gas
LULUCF	Land use, land use change and forestry
NCAS	National Carbon Accounting System
NFI	National Forest Inventory
NG	Natural Gas
NGERs	National Greenhouse and Energy Reporting System
NIR	National Inventory Report
QA/QC	Quality assurance/Quality control
QDME	Queensland Department of Mines and Energy
QDNRME	Queensland Department of Natural Resources, Mines and Energy
UNFCCC	United Nations Framework Convention on Climate Change
VKT	Vehicle Kilometres Travelled

PART 1 – Annual Inventory Submission

1. INTRODUCTION AND INVENTORY CONTEXT

1.1 BACKGROUND INFORMATION ON GREENHOUSE GAS INVENTORIES

1.1.1 Inventory reporting

The United Nations Framework Convention on Climate Change (UNFCCC) was ratified by Australia in 1992 and entered into force in March of 1994. One of the principal commitments made by the ratifying Parties under the Convention was to develop, publish and regularly update national emission inventories of greenhouse gases (GHGs).

Australia's *National Inventory Report 2009* provides estimates of Australia's net greenhouse gas emissions for the period 1990-2009. This report and associated common reporting format (CRF) tables¹ are submitted to the UNFCCC to fulfil Australia's reporting obligations under the Convention.

Australia's ratification of the Kyoto Protocol came into force in March 2008 and this inventory submission also constitutes Australia's annual submission under the Kyoto Protocol.

The report has been prepared in accordance with the UNFCCC *Reporting Guidelines on Annual Inventories*

agreed by the Conference of Parties at its eighth (decision 18/CP.8), ninth (decision 13/CP.9) and eleventh sessions (decision 14/CP.11), and set out in document FCCC/SBSTA/2006/9 and the supplementary reporting requirements under Article 7 of the Kyoto Protocol (Decision 15/CMP.1).

The emission estimates provided in this report have been compiled in accordance with the Intergovernmental Panel on Climate Change (IPCC) *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC 1997), *IPCC Good Practice Guidance and Uncertainty Management in National Greenhouse Gas Inventories* (IPCC 2000); and the *Good Practice Guidance on Land Use, Land Use Change and Forestry* (IPCC 2003). The aim is to ensure that the estimates of emissions are accurate, transparent, consistent through time and comparable with those produced in the inventories of other countries.

1.1.2 Gases

The *National Inventory Report* covers sources of greenhouse gas emissions and removals by sinks resulting from human (anthropogenic) activities for the major greenhouse gases – carbon dioxide (CO₂), methane (CH₄), nitrous oxide (N₂O), perfluorocarbons (PFCs), hydrofluorocarbons (HFCs) and sulphur hexafluoride (SF₆). Also covered in ancillary fashion for reporting under the Convention are the indirect greenhouse gases-carbon monoxide (CO), oxides of nitrogen (NO_x), and non-methane volatile organic compounds (NMVOCs). Sulphur dioxide (SO₂), an aerosol precursor, is also included because emissions of this gas influence global warming.

The *National Inventory Report* presents emissions for each of the major greenhouse gases as carbon dioxide equivalents (CO₂-e) using the 100-year global warming potentials (GWPs) contained in the 1995 IPCC *Second Assessment Report* (IPCC 1996)². As greenhouse gases vary in their radiative activity and in their atmospheric residence time, converting emissions into CO₂-e allows the integrated effect of emissions of the various gases to be compared.

1 Australia's CRF tables are available on the web at <http://ageis.climatechange.gov.au/QueryCRFTable.aspx>

2 GWPs used are, 1 for CO₂, 21 for CH₄, 310 for N₂O, 6,500 for the PFC perfluoromethane (CF₄), 9,200 for the PFC perfluoroethane (C₂F₆), and 23,900 for SF₆. GWPs are not available for the indirect greenhouse gases and in accordance with the UNFCCC reporting guidelines are reported but are not included in the inventory total.

1.1.3 Sectors

Emissions and removals have been grouped under six sectors that have been defined by the IPCC. These represent the main human activities that contribute to the release or capture of greenhouse gases into or from the atmosphere:

- Energy
- Industrial processes
- Solvent and other product use
- Agriculture
- Land use, land use change and forestry (LULUCF)
- Waste.

Under the Kyoto Protocol only a subset of the LULUCF sector emissions and removals are open to being included in the account. These include deforestation, afforestation and reforestation activities that have occurred since 1990 (Article 3.3 activities which are mandatory) and forest, cropland or grazing land management and revegetation (Article 3.4 activities which can be elected by the Party). Australia has not elected to account for any Article 3.4 activities for the first commitment period.

1.1.4 Reporting year

The Australian inventory is reported for Australian fiscal years as key data sources such as the National Greenhouse and Energy Reporting System (NGERS) and national energy and agricultural statistics obtained from national statistical agencies (the Australian Bureau of Agricultural and Resource Economics and the Australian Bureau of Statistics) are published on this basis. The year 2009 refers to the Australian fiscal year from 1 July 2008 to 30 June 2009, and a similar format is used for other years. Time series consistency is maintained. The estimates of emissions and removals in the *land use, land use change and forestry* sector, where inventory-specific monitoring systems have been put in place, are produced on a calendar year basis. The use of fiscal year data is consistent with *Good Practice* as the use of these data conforms to the normal practice of Australia's national statistical agencies and leads to more accurate emissions estimates.

1.1.5 Structure of the National Inventory Report

The structure of this *National Inventory Report* has been organised to conform to the UNFCCC *Reporting Guidelines on Annual Inventories* (FCCC/SBSTA/2006/9) and the supplementary reporting requirements under Article 7 of the Kyoto Protocol (Decision 15/CMP.1).

The *National Inventory Report* provides estimates of Australia's total net emissions in 2009 and identifies trends in emissions between 1990 and 2009 for each of the sectors and for the main greenhouse gases. It also provides, *inter alia*, comprehensive information on estimation methodologies and data quality; details of recalculations of emissions estimates and background on the national system and the inventory preparation processes in order to facilitate international review and comparisons with the inventories of other countries.

National System

In accordance with Article 5, paragraph 1 of the Kyoto Protocol, Australia has put in place a national system for the estimation of anthropogenic emissions by sources and removals by sinks of all greenhouse gases not controlled by the Montreal protocol. The guidelines for national systems (annex to Decision 19/CMP.1) detail the characteristics of a national inventory system (Table 1.1). Sections 1.2 to 1.8 of the *National Inventory Report* describe the main components of Australia's national system.

Supplementary Kyoto Protocol reporting requirements

Chapters 11 to 15 of the *National Inventory Report* contain the supplementary Kyoto Protocol reporting information on emissions and removals from the LULUCF Article 3.3 activities, Kyoto units, minimisation of adverse impacts in accordance with Article 3.14 and changes to the national system and registry.

1.1.6 National Greenhouse Accounts

In addition to this report the Department of Climate Change and Energy Efficiency (DCCEE) publishes a range of supporting emission estimates that, together, constitute the *Australian National Greenhouse Accounts*. In addition to the *National Inventory Report*, the DCCEE also prepares:

- the *National Greenhouse Gas Inventory*, which provides a summary of the Kyoto Protocol account;
- an overview of the *State and Territory Greenhouse Gas Inventories*; and
- the *National Inventory by Economic Sector*, comprising emission estimates by economic sector rather than by IPCC sectors as in this report.

These documents are available on the DCCEE website www.climatechange.gov.au/en/climate-change/emissions.aspx. They provide additional information with respect to Australia's emissions on both a regional and industry basis.

Table 1.1: Reporting of national system characteristics against the guidelines for national systems (annex to Decision 19/CMP.1)

General Functions		
Paragraph number	Description of national inventory system characteristic	Section cross reference
10a	Establish and maintain institutional, legal and procedural arrangements	1.2
10b	Ensure sufficient capacity for timely function of NS functions	1.2
10a	Technical competence of staff	1.2
Specific Functions		
Paragraph number	Description of national inventory system characteristic	Section cross reference
12a	Designate a single national entity	1.2
12b	Make available postal and electronic addresses of national entity	1.2
12c	Information on actors, institutional, legal and procedural arrangements	1.2
12d	Elaborate a QA/QC plan	1.6
12e	Establish process for official consideration	1.2
13	Improve quality of the inventory	1.6, 10
14a	Identify key source categories	1.5
14b	Prepare estimates in accordance with methods described by the IPCC	1.5
14c	Collect sufficient activity data to support the methods	1.3
14d	Estimate inventory uncertainty	1.7 Annex 7
14e	Information on recalculations	10
14g	Information on general inventory QC (tier 1) procedures in accordance with the QA/QC plan	1.6
15a	Information on specific QC (tier 2) procedures	1.6
15b	Information on QA procedures including provision for basic review of the inventory by personnel not involved in the inventory development	1.2, 1.6
15c	Information on provision for more extensive review for key source categories	1.6
15d	Information on how 15(b) and 15(c) relate to evaluation of inventory planning process in order to meet quality objectives	1.3.1
16a	Information on how information is archived	1.3.2
16b	Information on what information is archived	1.3.2

1.2 INSTITUTIONAL ARRANGEMENTS

1.2.1 Single national entity

In accordance with the guidelines for national systems (Decision 19/CMP.1 Annex paragraph 12(a)), the responsibility for Australia's national inventory has been assigned to a single agency – the Department of Climate Change and Energy Efficiency under the Administrative Arrangements Orders of the Australian Government.

The DCCEE is responsible for all aspects of activity data co-ordination, emissions estimation; quality control, improvement planning, preparation of the reports and their submission to the UNFCCC on behalf of the Australian Government.

The designated representative with overall responsibility for the national inventory is:

The Director
National Inventory Team
Department of Climate Change and Energy Efficiency
Australian Government
GPO Box 854
Canberra ACT 2601
AUSTRALIA
nationalgreenhouseaccounts@climatechange.gov.au

1.2.2 Capacity for timely performance of the general and specific functions of the national system

The guidelines for national systems (Decision 19/CMP.1 Annex paragraph 10(b)) require that there is sufficient capacity for the timely performance of national inventory system functions. The production of high quality and timely greenhouse gas inventories is a resource-intensive process. To meet these objectives of quality and timeliness Australia has invested significant financial and human resources through the development of capital assets, training of DCCEE staff and the contracting of expert consultants as needed.

IT software systems

Estimation of emissions is conducted by the DCCEE utilising the Australian Greenhouse Emissions Information System (AGEIS) and, for the *land use, land use change and forestry* sector, the National Carbon Accounting System (NCAS) (see Figures 1.1 and 1.2).

The AGEIS has been designed to meet the requirements for national inventory systems and is an integral part of the inventory preparation and publishing processes. In particular, it fully integrates quality control procedures into the compilation process as well as centralising emissions estimation, inventory compilation and reporting; and data storage activities. The AGEIS provides high transparency levels for the inventory, with emissions data for the set of *National Greenhouse Accounts* publicly accessible through an interactive web interface <http://ageis.climatechange.gov.au/>.

The AGEIS is continuing to be expanded and refined to support the range of National Greenhouse Accounts in accordance with the *AGEIS Strategic Plan*. Recent investment includes commencement of the integration of the National Greenhouse and Energy Reporting (Measurement) Determination 2008 into the AGEIS; construction of a range of QC tools to enhance and automate existing actions; facilitation of more timely production of emissions estimates; and the implementation of carbon balance constraints across the inventory.

While the AGEIS is used for final preparation of the *National Greenhouse Accounts*, the inventory uses the National Carbon Accounting System (NCAS) estimates of the emissions and removals from the *land use, land use change and forestry* sector and KP-LULUCF activities. Since 1998 the NCAS has been progressively developed to provide a complete greenhouse gas accounting capability for

agriculture, forestry and land use change (including all carbon pools, gases, lands and land use activities). The eventual capacity will be a full spatial enumeration with emissions and removals calculated using a process-based, mass balance, carbon and nitrogen cycling ecosystem model. The progressive development of the NCAS is set around priorities according to the scale of emissions from either the land use activity or carbon pool. To date the fully spatially explicit process-based ecosystems modelling capability of the NCAS has been completed for the conversion of forests to other land uses (e.g. cropping and grazing), conversion of lands to forest, croplands remaining croplands and the grassland component of grasslands remaining grasslands.

Figure 1.1: Department of Climate Change and Energy Efficiency inventory asset structures and relationships

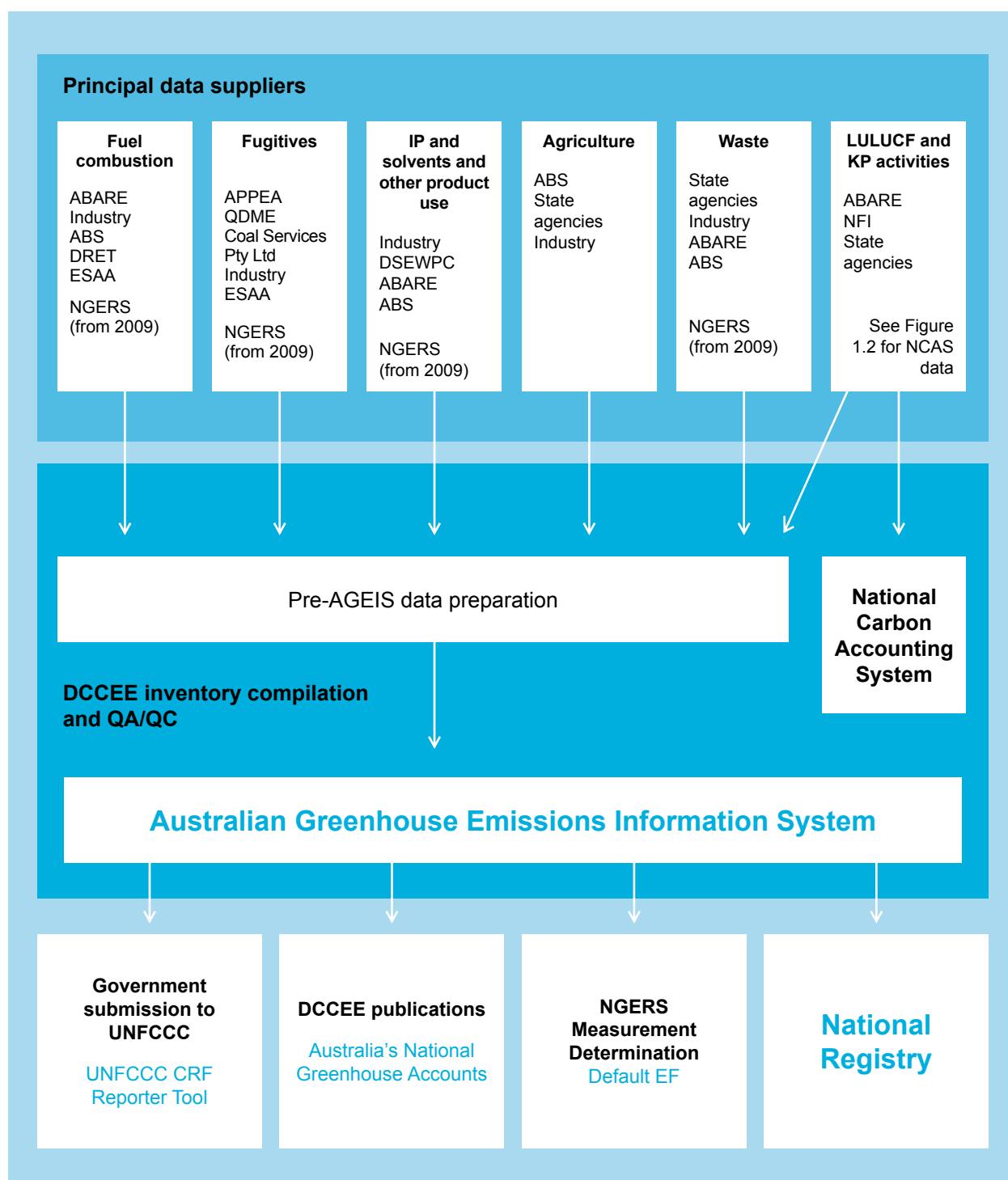
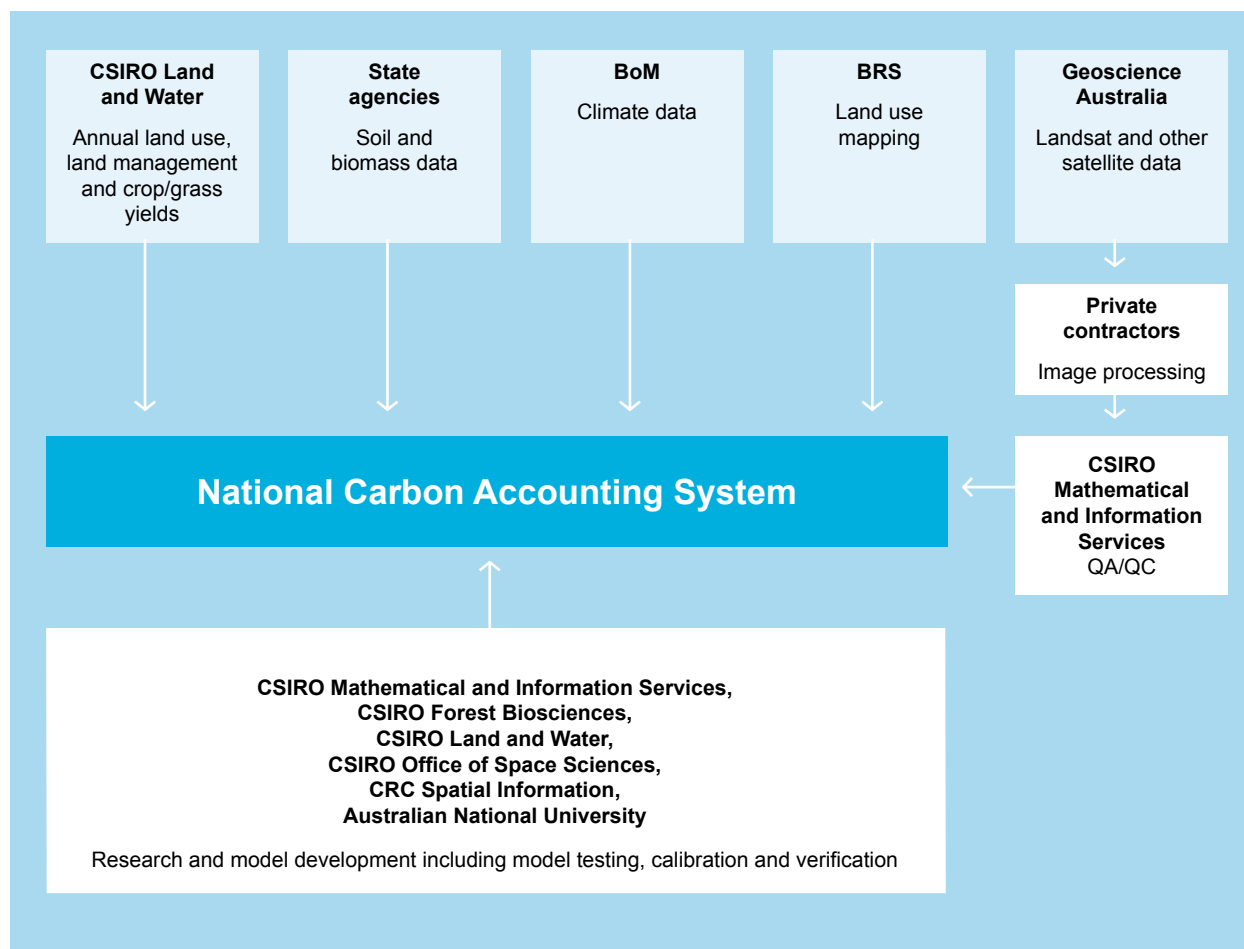


Figure 1.2: National Carbon Accounting System institutional arrangements



1.2.3 Technical competence of staff

DCCEE staff and external consultants have extensive experience in inventory preparation. The DCCEE aims to maximise the number of staff that have undergone the UNFCCC reviewer training and participated in UNFCCC Expert Review processes. All senior staff are qualified reviewers and have been accepted onto the UNFCCC Roster of Experts. The DCCEE has strongly supported the participation of Australian experts in the UNFCCC review process as this experience helps staff identify areas for improvement in Australia's inventory preparation. Where particular technical expertise is not available within the DCCEE, expert consultants are engaged to undertake analysis and review work.

1.2.4 Process for Official Consideration and Approval of the Inventory

The *National Inventory Report* is prepared by the Department of Climate Change and Energy Efficiency and is subject to the oversight of the Department's National Inventory Systems Executive Committee. The draft *National Inventory Report* is considered by the National Greenhouse Gas Inventory Committee, which comprises representatives of the Australian, State and Territory governments and the Commonwealth Scientific and Industrial Research Organisation (CSIRO) and has been in place since the early 1990s. This Committee is the principal mechanism for formal external review of the report prior to its release. The *National Inventory Report* is also circulated prior to submission to other Australian government departments and agencies and relevant state experts through the National Greenhouse Gas Inventory Committee.

Release of each year's new inventory and submission to the UNFCCC is approved by the Secretary of the Department of Climate Change and Energy Efficiency.

1.3 BRIEF DESCRIPTION OF THE PROCESS OF INVENTORY PREPARATION

1.3.1 Inventory Preparation

Key steps in the annual inventory preparation process (with indicative dates in parentheses) are determined by the needs of the system and output and quality objectives. The timing is determined by the UNFCCC submission timelines and data availability. Steps 1-17 provide an overview of the general inventory cycle. The cycle commences with a review of emission estimation methods, the allocation of tasks and selection of external consultants and the preparation of the AGEIS for the compilation of the forthcoming inventory. The cycle is completed by external independent review provided by the UNFCCC Expert Review Teams.

Planning and methodology improvement

1. Preparation of *Evaluation of Outcomes* document for the previous year.
2. Preparation of *QA/QC* and *Inventory Improvement plans*, taking into account DCCEE review of methodologies and activity data; UNFCCC expert review recommendations and the *Evaluation of Outcomes* document.
3. Development of investment and maintenance plan for the AGEIS, incorporating the *QA/QC plan* (June).
4. Methodology development and review. Incorporation into AGEIS (June – October)

Data collection and entry

5. Activity data collection conducted annually by the DCCEE supplemented by data collections by external consultants. Heavily reliant on NGERS data and published data from Australia's economic statistics agencies and subject to quality control checks.
6. Activity data entry into the AGEIS input database – by the DCCEE through predefined data entry templates (August – December).

Implementation of quality control measures

7. Activity data verification and quality control – the DCCEE uses the AGEIS to systematically report a range of diagnostic statistics on the activity data to facilitate identification and correction of anomalous entries to ensure time-series consistency and consistency across sectoral emissions estimates.
8. A designated analyst (known as a Supervisory user) investigates anomalies and records an assessment of the quality of the activity data in the system.
9. The data quality is checked and internally audited by a designated analyst (known as the Database Operations Manager DOM) to provide quality control. Only when the DOM is satisfied is the input data transferred to the core database where emissions estimation are undertaken.

Emission estimation

10. The AGEIS is used to generate emission estimates for all inventory years using time-series consistent methodologies.

Emission and report review

11. Emissions estimates verification – the DCCEE analysts repeat the range of tests on emissions estimates generated by the AGEIS to ensure time-series consistency and consistency across sectoral emissions estimates and accuracy of recalculations.
12. Completion of quality control measure tests to ensure estimates meet quality criteria.
13. The compiled inventory is circulated to Australian Government departments and the National Greenhouse Gas Inventory Committee of State and Territory government representatives for comment prior to public release (February).

Report publication

14. Automated population of reporting (CRF) tables (February).
15. The inventory is available for public release.
16. Release of Australia's National Greenhouse Accounts and the AGEIS database of emission estimates and background data at <http://www.climatechange.gov.au/en/climate-change/emissions.aspx> (April).
17. UNFCCC Expert Review of the National Inventory Report and CRF Tables (August-November).

1.3.2 Data collection, processing and storage

Data collection

Data collection to support the preparation of the *Accounts* is managed centrally by the DCCEE utilising a mix of approaches to ensure the reliable flow of data from other agencies to support inventory preparation.

National Greenhouse and Energy Reporting System (NGERS)

Input data to support the preparation of the *Accounts* for important elements of the *energy, industrial processes* and *waste* sectors are collected using the *National Greenhouse and Energy Reporting System* (NGERS).

The legislative framework for this mandatory national greenhouse and energy reporting system was established through The *National Greenhouse and Energy Reporting Act 2007*. An explicit objective of the Act is to collect information to support the development of the national inventory.

Under the NGERS, companies whose energy production, energy use, or greenhouse gas emissions (from the *energy, industrial processes* and *waste* sectors) meet certain thresholds must report facility-level data to the NGER regulator (the Greenhouse and Energy Data Officer), who currently is embedded in the DCCEE. NGERS provides data for activity data inputs, such as fuel combustion, emission factors at facility level and, in some cases, directly measured emissions.

The first reports submitted by companies covered the Australian financial year 2008-09 and this data has been used in the preparation of this inventory.

The rules for the estimation of activity data, emission factors and emissions by companies are well specified and set out in the *National Greenhouse and Energy Reporting (Measurement) Determination 2008*.

The estimation methodologies used for company and facility emissions are estimated within the *National Greenhouse Accounts* framework ensuring consistency among the relevant accounts – national, state and territory, industry, company and facility-level inventories. Integration of the estimation methods and data is critical for ensuring that changes in emissions at facility level are captured efficiently and accurately in the national inventory. The default methods used by companies are derived from the national inventory methods while the default emission factors have been derived using the AGEIS.

The *Greenhouse and Energy Data Officer* in the DCCEE manages the process of input data collection from companies and the dissemination of this data to relevant agencies. The Online Comprehensive Reporting tool (OSCAR) is used for the collection of the input data from companies.

Other data sources

NGERS data sources are used for the *energy, industrial processes* and *waste* sectors where possible, supplemented by the use of other published data sources only where necessary. The collection process for other data is well-integrated with the objectives of other programmes with a strong reliance on data collected and published by Australia's principal economic statistics agencies: the Australian Bureau of Statistics (ABS) and the Australian Bureau of Agricultural and Resource Economics and Sciences (ABARES). ABARES have collected energy statistics for over 35 years and use this data to meet Australia's reporting commitments to the International Energy Agency. The ABS is the national statistical

agency with legislative backing for its collection powers. It is the source of agricultural activity and some commodity and energy related data.

The DCCEE employs consultants to collect minor data sources directly from companies for the *industrial processes* sector and to process the satellite imagery to determine land cover change for the *land use, land use change and forestry* sector. Satellite imagery is sourced from Geosciences Australia (Australia's principal satellite ground station and data processing facility) via a memorandum of understanding. Data to support estimates of HFCs are sourced from compulsory reporting by importers under licensing arrangements under the *Ozone Protection and Synthetic Greenhouse Gas Management Act 2003*. The collection of solid waste data from State and Territory Government agencies is supported by an exchange of letters between Australian and State Government agencies.

Data processing

As described in section 1.2.2 the estimation of emissions is conducted by the DCCEE utilising the AGEIS and, for the *land use, land use change and forestry* sector, the NCAS.

Data Storage

The Australian documentation systems aim to both manage and retain all data used in the estimation of emissions and to provide a means for knowledge management to ensure continuity and security of the National Inventory Systems.

The AGEIS is at the heart of Australia's documentation systems as it allows efficient electronic data management and archiving of the significant quantities of data needed to generate an emissions inventory. AGEIS data management functions include:

- archival and storage within the AGEIS database of the emissions estimates of past submissions;
- archival and storage within the AGEIS of past activity data, emission factor and other parameters and models;
- archival and storage of data source descriptions; methodology descriptions and source reference material; and
- integrated access to the documentation of data sources; methodology description and source reference material.

The aims of these systems include giving inventory staff ready access to all related materials that underpin the emissions estimates and to provide the means for replication of emission estimates from past submissions.

The AGEIS functions are supported by some additional and important elements of the documentation system:

- documentation of the inventory's emission estimation methodologies in the NIR;
- maintenance of a National Inventory Library of source material documents.

1.4 BRIEF GENERAL DESCRIPTION OF METHODOLOGIES AND DATA SOURCES

1.4.1 Estimation methods

Greenhouse gas emissions are generated from a large number of processes and from a range of often diffuse sources. Emissions are not usually monitored directly but are generally estimated through the application of models and methodologies that link emissions to data on observable activities. The Australian methodology for estimating greenhouse gas emissions and sinks uses a combination of country-specific and IPCC methodologies and emission factors. These methods are consistent with the *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC 1997), the IPCC

Good Practice Guidance and Uncertainty Management in National Greenhouse Gas Inventories (IPCC 2000) and *Good Practice Guidance on Land Use, Land Use Change and Forestry* (IPCC 2003) and are comparable with international practice.

In general, *Australia's National Greenhouse Accounts* have been moving towards a mix of tier 2 and tier 3 estimation methods that incorporate:

- facility-specific emission estimation processes;
- characterisations of the capital and technology types at the point of emission;
- dynamic relationships that link current emissions outcomes with the activity levels of previous years; and
- spatial differences in emissions processes across Australia.

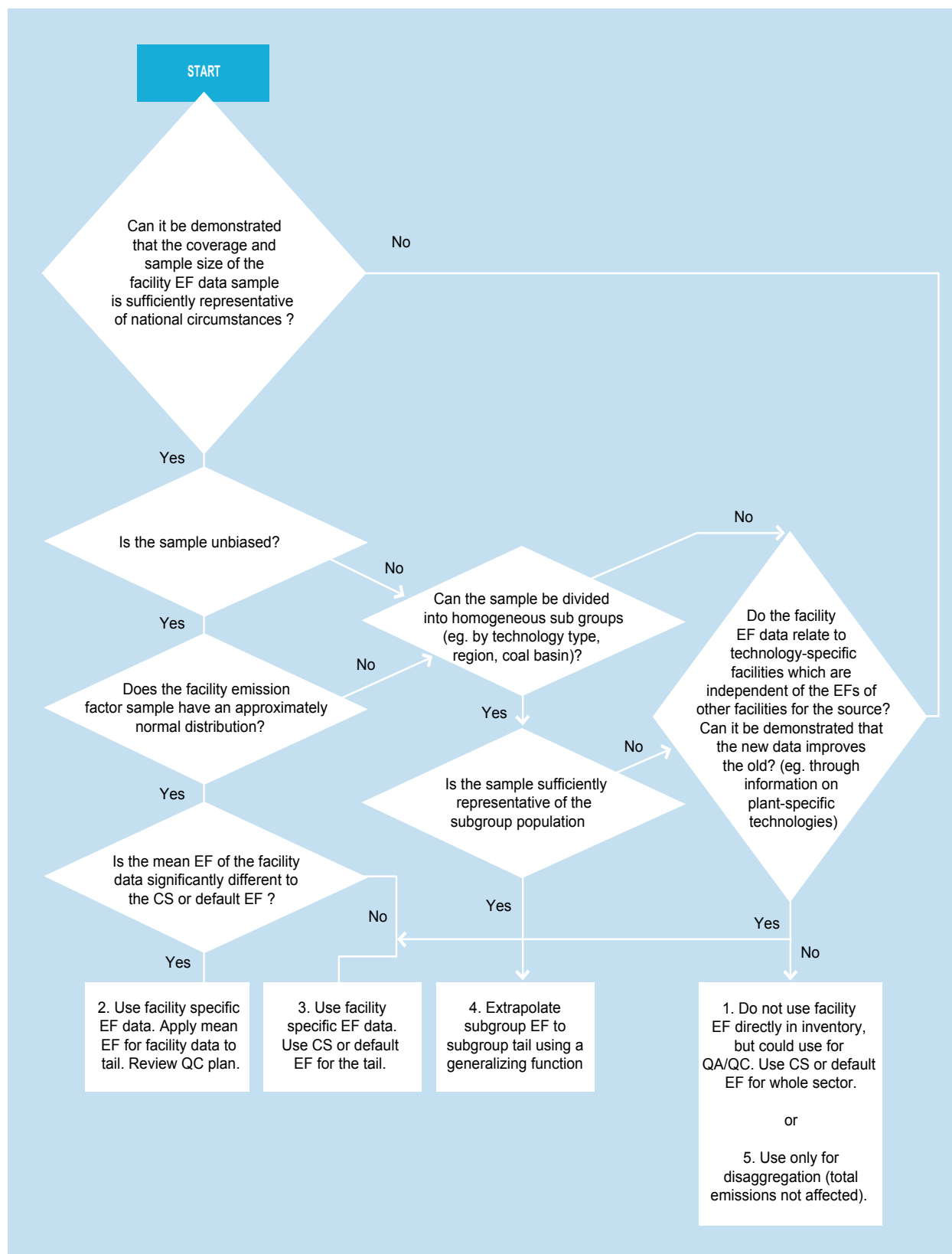
The additional complexity in the methodology allows emissions to be estimated more accurately. Detailed descriptions of methods chosen are set out in the relevant chapters.

The commencement of data collection under the *NGER Act 2007* provides the foundation for a progression of national inventory estimation approaches towards tier 3 or facility-specific estimation approaches within the energy, industrial process and waste sectors.

Tier 3 approaches have been in place for fuel combustion in the electricity industry for a number of years and from fugitive emissions from underground coal mining sources since the most recent inventory. These existing tier 3 approaches have been supported by the use of the new NGERS data. For a range of additional categories a mix of tier 2/3 approaches will be implemented over time as methods for facility-specific measurement of emissions or key data inputs are adopted by reporters and as key pre-conditions for implementation of the new methods are met. These circumstances include that the data must comply with prescribed data standards (in this case set out in the NGERS Measurement Determination); that there is a timely and comprehensive data collection system in place; and that the resulting emission estimates for the source pass the inventory quality criteria set out in the QA-QC plan (for example, in relation to completeness and international comparability).

Consistent decision making with respect to the use of facility specific emission factors has been ensured through the application of a decision tree as set out in Figure 1.3.

Figure 1.3: Consistent decision making in method selection



In particular, tier 3 methods incorporating facility-specific emission factor data obtained from NGERS have been utilised where the following conditions have been met:

- the sample size of the available NGERS data is sufficiently large;
- there is no evidence of bias in the distribution of the NGERS emission factor data ; and
- the distribution of the NGERS data is approximately normally distributed.

For the balance of a source where there are facilities for which no facility-specific data are available, a country-specific factor is applied.

Tier 3 methods incorporating NGERS facility-specific data are also able to be used in two other cases where large samples displaying characteristics of an approximately normal distribution cannot be obtained.

The first additional case relates to the situation where, within one source, a number of homogenous sub-samples can be discerned. In particular, this is the case for fugitive emissions from coal mining where fugitive emissions depend on a number of variables including the geological history of a particular coal mining basin. Fugitive emissions from the coal mining source have been broken into sub samples of homogenous groups based on coal mining basins and where data for the tail of the source – where facilities have unknown characteristics – have been determined by the extrapolation of information from the relatively homogenous sub sample.

The second additional case relates to the situation where facility data are heavily technology dependent and where the data for each facility are likely to be independent of one another. In particular, this is the case in the industrial wastewater category where knowledge of the technology deployed at one facility does not affect the likelihood of a certain technology being deployed at another facility where no facility data is available. In these cases, it is possible to utilise the facility data where available and it may not be appropriate to extrapolate information from the NGERS sample to the remainder of a particular source. Consequently, in these cases, the original tier 2 emission factor has been retained for the tail of the source where NGERS data has not yet been collected.

In Table 1.2, the categories where NGERS data on emission factors has been utilised in this inventory are identified (as ‘implemented’ in the fourth column of the table). Categories where NGERS data will be considered for use in the inventory in future submissions as it becomes available are identified as ‘potentially’ in the fourth column.

Table 1.2: Summary of planned uses of NGERS data for Australia’s national inventory, by IPCC sector

Category	Facility – level activity data	Tier 2/3	Verification test for tier 2 parameters	Completeness/ sectoral improvement	Improved uncertainty estimates
1 Energy					
1.A.1a Electricity (coal)	Implemented	Implemented	Yes	No	Yes
1.A.1a Electricity (gas)	Implemented	Implemented	Yes	No	Yes
1.A.1a Electricity (liquid)	Implemented	Potentially	Potentially	No	Yes
1.A.1b Petroleum refining	Implemented	Potentially	Potentially	Yes	Yes
1.A.1c Coke production	Potentially	Potentially	Potentially	No	Yes
1.A.2 Manufacturing	Potentially	Potentially	Potentially	No	Yes
1.A.3 Transport	No	No	No	No	No
1.A.4 Other sectors	No	No	Potentially	No	No
1.A.5 Other	No	No	Potentially	No	No

Category		Facility – level activity data	Tier 2/3	Verification test for tier 2 parameters	Completeness/ sectoral improvement	Improved uncertainty estimates
1.B.1	Coal Mining	Implemented	Implemented	Potentially	Yes	Yes
1.B.2	Oil & Gas	Partially implemented	Potentially	Potentially	No	Yes
2.A.1	Cement	Implemented	Potentially	Potentially	No	Yes
2.A.2	Lime	Implemented	Potentially	Potentially	No	Yes
2.A.3	Limestone and Dolomite use	Implemented	Potentially	Potentially	Yes	Yes
2.A.4	Soda ash production and use	Implemented	Implemented	NA	No	Yes
2.B.1	Ammonia	Implemented	Potentially	Potentially	No	Yes
2.B.2	Nitric acid	Implemented	Implemented	NA	No	Yes
2.B.5	Synthetic rutile and titanium dioxide	Implemented	Potentially	Potentially	No	Yes
2.C.1	Iron and steel	Potentially	Potentially	Potentially	No	Yes
2.C.2	Ferro-alloy metals	Implemented	Potentially	Potentially	No	Yes
2.C.3	Aluminium	Implemented	Potentially	Potentially	No	Yes
2.C.4	Other metals	Implemented	Potentially	Potentially	No	Yes
2.E	HFC production	No	No	No	No	No
2.F	HFC consumption	No	No	No	No	No
2.F	SF ₆ consumption	Implemented	Potentially	Potentially	Yes	No
3	Solvents	No	No	No	No	No
4	Agriculture	No	No	No	No	No
5	LULUCF	No	No	No	No	No
6 Waste						
6.A	Solid waste	No	Potentially	Potentially	No	Yes
6.B.1	Domestic and Commercial Wastewater	No	Potentially	Potentially	No	Yes
6.B.2	Industrial Wastewater	Partially implemented	Potentially	Potentially	Yes	Yes
6.C	Waste incineration	Yes	Potentially	Yes	Yes	Yes

1.4.2 Data sources

The inventory is prepared using a mix of sources for activity data including DCCEE collections and published data from national statistical agencies. The principal data sources are set out in Table 1.3.

Table 1.3: Principal data sources for the estimation of Australia's inventory

Category (UNFCCC sector)	Principal data sources	Principal collection mechanism
Energy sector (1A1, 1A2, 1A4, 1A5)	Australian Bureau of Agricultural and Resource Economics and Sciences (ABARES), NGERS	Published, DCEE Mandatory data reporting system
Energy sector (1A3)	ABARES, Australian Bureau of Statistics	Published
Energy sector (1B)	NGERS, Coal Services Pty Ltd, Australian Petroleum Exploration Association	DCCEE Mandatory data reporting system, published
Industrial processes (2) and solvents (3)	NGERS Australian Government Department of Sustainability, the Environment, Water, Population and Communities	DCCEE Mandatory data reporting system Mandatory reporting of HFCs under import licensing arrangements
Agriculture (4)	Australian Bureau of Statistics	Published
Land use, land use change and forestry (5)	Geosciences Australia ABARES	Memorandum of Understanding Published
Waste (6)	State and Territory government waste agencies	Exchange of letters between government agencies

National Greenhouse and Energy Reporting System (NGERS)

The NGERS is an integral element of the national inventory system. The rules for estimation of data and emissions at the facility level by companies are set out in the *National Greenhouse and Energy Reporting Measurement Determination 2008*, which is made under subsection 10 (3) of the Act.

The structure of the Determination is designed to facilitate the integration of corporate and facility level data provided under the Act with international data standards on greenhouse emissions.

The scope of the Determination is given by the following categories of emission sources:

The emission sources are:

- **Fuel combustion:** emissions from the combustion of fuel for energy (see Chapter 2 of the Determination);
- **Fugitive emissions** from the extraction, production, flaring, processing and distribution of fossil fuels (see Chapter 3 of the Determination);
- **Industrial process emissions** where a mineral, chemical or metal product is formed using a chemical reaction that generates greenhouse gases as a by-product (see Chapter 4 of the Determination); and
- **Waste** emissions from waste disposal – either in landfill, as management of wastewater or from waste incineration (see Chapter 5 of the Determination).

The scope of the Determination does not include land based emissions covered by the IPCC categories 'Agriculture' and 'Land Use, Land Use Change and Forestry'. Emissions from fuel combustion for land based industries are, nonetheless, covered by the Determination.

Four estimation methods are provided for under NGERS ranging from low cost simple default methods to higher order methods requiring sampling and analysis of inputs or direct monitoring of emissions.

In general, as with the EU ETS, reporters may choose the estimation method appropriate to their own circumstances. Some important exceptions relate to reporters in the electricity generation, underground coal mining and aluminium industries which are required to use method 2 or higher (see below) for key components of their emission estimations. These restrictions cover around 60 per cent of emissions reported under NGERS.

NGERS Method 1: the National Greenhouse Accounts default method. Method 1 specifies the use of designated emission factors in the estimation of emissions. These emission factors are national average factors determined by the Department of Climate Change and Energy Efficiency using the Australian Greenhouse Emissions Information System (AGEIS). Although significantly updated, this method is very similar in approach to that used by many corporations for over a decade to report emission estimates under the Greenhouse Challenge Plus program.

The national inventory only utilises activity data collected from companies that report using this method as no new information is collected in relation to emission factors or in relation to other key facility-specific parameters.

NGERS Method 2: a facility-specific method using industry sampling and Australian or international standards listed in the Determination or equivalent for analysis of fuels and raw materials to provide more accurate estimates of emissions at facility level. Method 2 enables corporations to undertake additional measurements – for example, the qualities of fuels consumed at a particular facility – in order to gain more accurate estimates for emissions for that particular facility. Method 2 draws on the large body of Australian and international documentary standards prepared by standards organisations in order to provide the benchmarks for procedures for the analysis of, typically, the critical chemical properties of the fuels being combusted. Method 2 was based on existing technical guidelines used by reporters under the Generator Efficiency Standards program which had been in place since 1998.

The national inventory may utilise activity data and emission factor or other key facility-specific parameters collected by companies using this method depending on the analysis of the quality of the data and in accordance with the decision tree set out in section 1.4.1.

NGERS Method 3: a facility-specific method using Australian or international standards listed in the Determination or equivalent standards for both sampling and analysis of fuels and raw materials: Method 3 is very similar to method 2, except that it requires reporters to comply with Australian or equivalent documentary standards for sampling (of fuels or raw materials) as well as documentary standards for the analysis of fuels.

NGERS Method 4: direct monitoring of emission systems, either on a continuous or periodic basis: Method 4 provides for a different approach to the estimation of emissions. Rather than providing for the analysis of the chemical properties of inputs (or in some case, products), method 4 aims to directly monitor greenhouse emissions arising from an activity. This approach can provide a higher level of accuracy in certain circumstances, depending on the type of emission process: however, it is more likely to be more data intensive than other approaches.

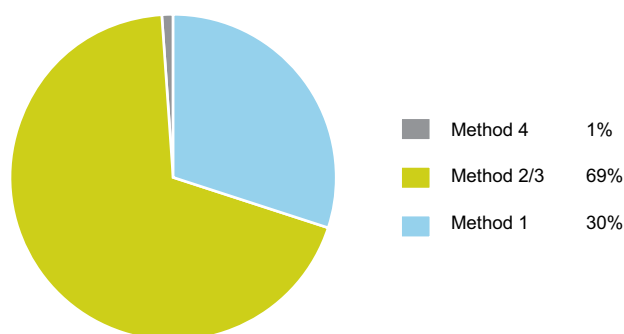
As for methods 2 and 3, there is a substantial body of documented procedures on monitoring practices and state and territory government regulatory experience that provide the principal sources of guidance for the establishment of such systems.

The national inventory may utilise emissions data generated using NGERS method 4 depending on the analysis of the quality of the data and in accordance with the decision tree set out in section 1.4.1.

Implementation of the NGER Measurement Determination

In the first year of implementation of NGERS (2008-09), sixty nine per cent of carbon dioxide emissions were estimated using method 2 or 3 – ie using analysis of carbon content of fuels or other inputs. By comparison, thirty per cent of carbon dioxide emissions were estimated using method 1. Around 1 per cent of carbon dioxide emissions were estimated using method 4. These outcomes reflect the choices made by the NGER system and by companies within the system and reflect the significance of the source and the likely variability in the carbon content of the source. For example, over 90 per cent of emissions from the combustion of coal were estimated using a higher order method. However, method 1 continued to be used principally for petroleum products which tend to be homogenous in character and where payoff from additional measurement effort is often limited. Choices made by companies for gas lay somewhere between coal and petroleum products.

Figure 1.4: Carbon dioxide emissions: method selected by NGERS reporters

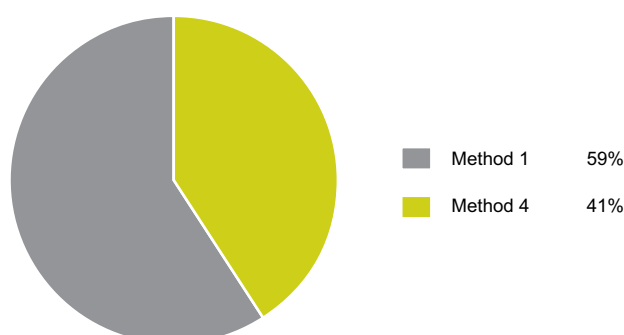


There is a different story, however, when choices made about estimation methods used for methane are considered. Around 41% of methane emissions were estimated using direct monitoring of emissions while fifty – nine per cent of methane emissions were estimated using method 1.

As for carbon dioxide, the choices of the system, and of companies within the system, have resulted in the use of actual measurements from facilities to determine emissions for major sources of methane. This outcome relates principally to the choices made by underground coal mines to use directly monitored estimates.

For minor sources of methane and where measurement is difficult, such as methane from combustion of fuels, method 1 has been used by reporting companies under NGERS.

Figure 1.5: Methane emissions: method selected by NGERS reporters



The particular use of this NGERS data within the national inventory for each category is explained within their respective chapters of this document

Activity data

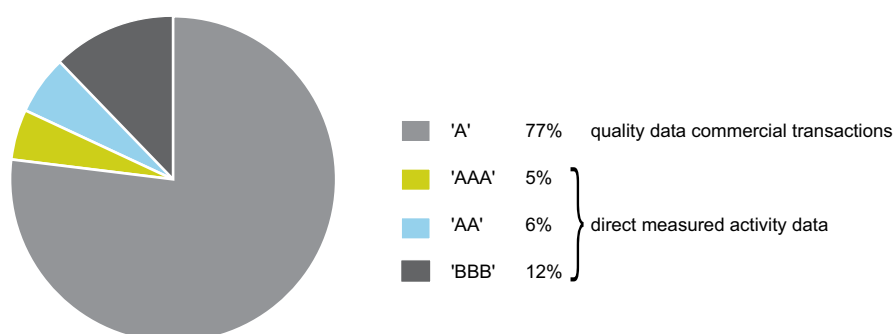
The NGERS system generates activity data on fuel consumption and key activity data inputs in the industrial process and waste sectors for NGERS reporters.

The NGERS system aims to maximise the amount of activity data collected from companies that is used by companies for other, regulated purposes including commercial activity and taxation. This approach both reduces the regulatory burden on companies and ensures consistency across national datasets but also formalises the role of the national measurement systems in the national inventory system.

Activity data is rated 'A' if it is estimated using information used to support commercial transactions such as estimates of the amount of fuel purchased. Activity data is rated 'AAA' if companies make an estimate of fuel consumed, rather than fuel purchased, using the same tolerance levels for measurement error that govern commercial transactions. Activity data is rated 'AA' if companies estimate fuel consumed based on information on the amount of fuel purchased and change in stock at the facility. In some small amount of cases fuel use is not subject to either commercial or taxation activity (ie, where a facility both extracts and utilises fuel). In these cases, the quality of the data must be signified by a quality rating (e.g. "BBB" data). All quality data is publishable by the NGERs regulator.

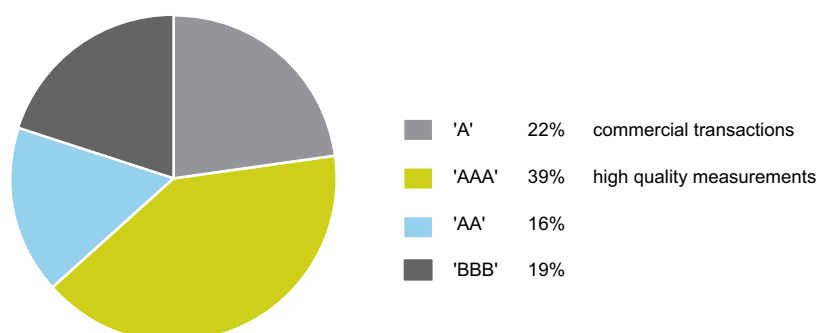
The choices made by companies with respect to the quality of their activity data inputs are presented in Figure 1.6. Over 75 per cent of reported activity data points under NGERs is derived from commercial transactions and requires no new measurements to be undertaken by the company in order to meet NGERs reporting requirements.

Figure 1.6: Activity data: method selected by NGERs reporters by percentage of data points



However, in terms of carbon dioxide emissions, companies have tended to choose to use actual measurements of activity to underpin emissions estimates. Thirty-nine percent of emissions were estimated using 'AAA' activity data inputs – ie estimates of fuel measured at the point of combustion at an accuracy level consistent with standards required to support commercial activity.

Figure 1.7: Activity data: method selected by NGERs reporters by percentage of emissions



It follows that companies have generally used existing commercial data for relatively minor emission sources. While commercial data accounted for 77 per cent of the data points used in emission estimation processes, these data points only related to 22 per cent of the estimated emissions.

Use of commercial activity data occurs primarily for gas and petroleum products – often minor sources or where uncertainties associated with the use of data on fuels purchased as a proxy for fuels consumed are considered low. It appears that for major emissions sources, Australian companies have chosen to use the most accurate data requiring explicit measurement effort while for minor emission sources they have chosen to use low cost, albeit slightly less accurate data.

NGERS data is supplemented where necessary by alternative data sources. Currently national data for the *energy* sector are principally obtained by voluntary data collection surveys undertaken by the Australian Bureau of Agricultural and Resource Economics and Sciences. Agriculture data is obtained by agricultural censuses and surveys conducted by the Australian Bureau of Statistics while waste data is principally obtained under State Government legislation.

1.5 BRIEF DESCRIPTION OF KEY SOURCE CATEGORIES

The *IPCC Good Practice Guidance* (2000, 2003) introduces the concept of ‘key categories’ for prioritising the inventory development process.

A key category has a significant influence on a country’s total inventory of direct greenhouse gases in terms of absolute level of emissions, the trend in emissions, or both. Australia has identified the key categories for the inventory using the tier 1 level and trend assessments as recommended in the *IPCC Good Practice 2003* report and adopted by COP decision 13/CP9. This approach identifies sources that contribute to 95% of the total emissions or 95% of the trend of the inventory in absolute terms.

When the LULUCF sector is included in the analysis, Australia has identified *public electricity (solid fuel)*, *grassland remaining grassland*, and *land converted to grassland* as the most significant of the key categories (i.e. contributing more than 10% of the level and/or trend) in 2009. When the LULUCF sector is excluded from the analysis the most significant key categories in 2009 are *public electricity (solid fuel)*, *road transportation (liquid fuels)* and *enteric fermentation (sheep)*. More details are provided in Annex 1.

The concept of key categories is also used for choosing the good practice estimation methods for emissions and removals due to activities under Articles 3.3 and 3.4 of the Kyoto Protocol. The KP-LULUCF key categories have been identified as outlined in the *IPCC Good Practice for LULUCF* (IPCC 2003).

For the Article 3.3 activities Australia has identified both deforestation and afforestation/reforestation as key categories.

1.6 INFORMATION ON THE QA/QC PLAN

This section outlines the major elements of the QA/QC plan. Australia’s QA/QC plan is documented in full in *National Inventory Systems: Quality Control-Quality Assurance plan*.

The IPCC defines QC as being a system of routine technical activities to measure and control the quality of the inventory as it is being developed. A basic QC system should provide routine and consistent checks to ensure data integrity, correctness, and completeness, identify and address errors and omissions, and document and archive inventory material and record all QC activities.

Quality assurance comprises a planned system of review procedures conducted by personnel not directly involved in the inventory compilation and development process.

The QA/QC processes deployed by DCCEE aim to conform to *IPCC Good Practice Guidance* (2000, 2003) and to contribute to the production of inventories which are accurate, and in which uncertainties are reduced to the extent practicable, and in which the inventories are transparent, documented, consistent over time, complete, and comparable. The QA/QC plan identifies key risks to the achievement of these objectives and sets out the mitigation strategies employed to ensure that the quality objectives for emission estimates are attained.

Key risks to the attainment of the defined quality objectives are identified at each level of inventory preparation including the measurement of data at the facility level; the collation of activity and other input data by DCCEE and other agencies; and the process of emissions estimation.

Principal mitigation strategies are discussed in sections 1.6.1 to 1.6.3. A detailed summary of the quality control measures employed in the preparation of Australia’s inventory is presented in Annex 6.

Systems have been established to monitor the outcomes of the mitigation strategies and control measures – principally managed through the AGEIS system – see sections 1.6.1 and 1.6.2. Each year an evaluation of the data collected under the monitoring systems is undertaken and documented in the *National Inventory Systems: Evaluation of Outcomes* document. Following consideration of the *Evaluation of Outcomes* document, improvements to the inventory are then effected through the *National Inventory Systems: Inventory Improvement Plan*.

1.6.1 Quality Control procedures implemented

The principal data source for this inventory is the National Greenhouse and Energy Reporting system. The quality control system for this data is critical for the quality for the inventory as a whole.

Use of Standards

A key mitigation strategy to manage risks associated with measurement error is to ensure that rules for emissions estimation are well specified. Rules for the estimation of emissions by companies have been developed to conform with the National Greenhouse Accounts framework and aims to ensure that consistent estimation methods are deployed at the national, state and territory, industry, company and facility level. This consistency is critical to ensure policy efficiency but also engenders confidence in the company estimates because the methods used are also consistent with IPCC guidelines.

The *NGER Measurement Determination* is supplemented by the referencing of standards for sampling and analysis of key data inputs. For the estimation of facility-specific emission factors, for example, NGERs methods reference relevant Australian, ISO, and equivalent international standards (EU, US) for sampling and analysis of relevant fuel qualities and characteristics (such as carbon content). These standards provide, *inter alia*, sample handling protocols and tolerance levels for precision (repeatability and reproducibility) as well as for the management of bias.

Where possible, the NGERs has been designed to utilise the data systems that operate to support other regulatory functions such as commercial or taxation activities. In particular, measurement of commercial activity data in Australia is regulated by the *National Measurement Act 1960* and *National Measurement Regulations 1999* and, for utilities, by State Government regulations. These legislative instruments underpin the quality of all activity data subject to commercial operation that are used in the *National Greenhouse Accounts*. For example, the *National Measurement Regulations 1999* specify maximum tolerances for measurement error for any amount of solid fuel subject to commercial activity.

Certain data sources are also governed by the regulations of the taxation system. For example, data on liquid fuels are governed by the requirements of the *Excise Tax Act 1901* which places strict tolerance limits on measurement error. To an important extent, the quality of commercial and taxation data in Australia underpins the quality of emissions data reported under NGERs.

Compliance

In order to facilitate accurate reporting of information, the Greenhouse and Energy Data Officer (GEDO) has resources devoted to ‘outreach’ whereby the departmental officials liaise with reporting companies to assist them in the preparation of reports. A validation unit is also deployed by the GEDO to assist with the initial inspection of reported data, checking for transcription errors and liaising with companies about possible resubmission of estimates.

The NGER Act also provides for a risk-based system for the independent verification of NGERs data. Under the Act, the GEDO in the DCCEE has the authority to order a corporation to conduct an external audit on aspects of the corporation’s compliance with the Act or with the regulations. Sections 73 and 74 of the Act define the circumstances under which a greenhouse and energy audit may be initiated and allow for the appointment of Registered Greenhouse and Energy Auditors to undertake audit engagements.

The National Greenhouse and Energy Reporting (Audit) Determination 2009 sets out the requirements for preparing, conducting and reporting on greenhouse and energy audits. Greenhouse and energy audits may only be conducted by a greenhouse and energy auditor who has been registered under section 75A of the Act. The purpose of greenhouse and energy audits is to determine the extent to which entities that are required to register and report under the Act have, or have not, complied with its requirements.

The Act empowers the Greenhouse and Energy Data Officer to initiate a greenhouse and energy audit where:

- there are reasonable grounds to suspect that an entity that is required to register and report under the Act has contravened, is contravening, or is proposing to contravene either the Act or the Regulations; or
- it is determined that, for another reason, an audit of an entity's compliance with one or more aspects of the Act or the Regulations is necessary.

The Officer may also apply to a magistrate for a warrant in order to search the premises of a Corporation for anything on the premises that may relate to compliance with the Act. Significant penalties may apply to Chief Executive Officers for contravention of the Act.

An important focus of greenhouse and energy audits in the first two years of reporting will be to assist corporations to understand their obligations and develop efficient reporting processes that meet the needs of government and business.

Given the risk of a mandatory audit ordered by the GEDO, and the threat of significant penalty, many companies have utilized external auditors voluntarily to audit their reports prior to submission to the GEDO in 2009 and 2010.

Transparency is a key principle for the operation of the system. Data for individual companies in relation to total emissions are publicly disclosed by the GEDO through publication four months after submission to ensure additional scrutiny of reported data.

Time series consistency with audited data

For the preparation of the national inventory, data collected under NGERS has been checked for time series consistency with facility data available for previous years. In some cases, this data has been collected previously for the inventory – for example fuel combustion in the electricity generation sector. However, more generally, NGERS data has been compared with data the DCCEE has collected at facility level for 2007 and 2008. This data has become available as the DCCEE has sought submissions from targeted companies specifying their emission estimates for 2007 and 2008 in order to provide the basis for the provision of compensation arrangements to apply for various climate mitigation related policies.

Rules for these schemes have been issued and include a requirement that these data are audited prior to submission to DCCEE. Under the Emissions Intensive Trade Exposed scheme (EITES), NGERS methods have been used for the estimation of emissions. Under the Electricity Sector Adjustment Scheme (ESAS), the regulations also require companies to reconcile any data provided with data previously submitted to the DCCEE for the purpose of preparing the national inventory. These audit and reconciliation processes ensure the quality of the data available for 2007 and 2008 and, through time series consistency checks, the data from NGERS used in 2009.

All EITES emissions data have been made available for use in the preparation of the national inventory. Use of this data is indicated in the respective chapters of the National Inventory Report. Where EITES data has been used, facility level data has been implemented for the years 2007, 2008 and 2009.

Other datasets

Where the inventory utilises official national statistics, the quality control of this data is managed by the source agencies. The ABS publishes assessments of data quality and quantitative estimates of sampling errors for transport and agricultural activity data. National level energy activity data are produced by the ABARES through its annual Australian Energy Statistics. The ABARES data was reviewed and 'benchmarked' by the ABS in its role of national statistics co-ordinator (ABS 4648.0.55.001 Detailed Energy Statistics Australia). With respect to electricity, explicit reconciliations of energy data are undertaken consultants comparing data collected under NGERS and the estimates produced by ABARES, the Energy Supply Association of Australia and the Australian Energy Market Operator (AEMO) (which are all undertaken for slightly differing reasons and with slight differences in coverage).

Tier 1 quality control checks – emissions estimation

Emissions estimation is conducted through the use of the AGEIS software (apart from the LULUCF sector). Management of the AGEIS is conducted in accordance with the Control Objectives for Information and related Technology (COBIT) framework. The AGEIS was subject to a performance audit by the Australian National Audit Office in 2009 (see section 1.6.2).

For this inventory and associated time-series, there are approximate 2.5 million data inputs in the non-LULUCF sectors. To facilitate the management of such a large amount of data AGEIS was specifically developed to play a central role in the quality control of the national inventory. Key Tier 1 QC controls have been systematically built into the operation of the AGEIS. Auditable checks are undertaken *inter alia* to reduce the risks of errors associated with the input of activity data, missing data, recalculations and the time series consistency of generated emission estimates.

Input data and implied emission factors are also checked for recalculations and time-series consistency prior to submission using AGEIS and the CRF reporter tool. The allocation of roles and responsibilities of staff provide for the separation of data handling and data approval roles within the DCCEE to improve accountability.

Extensive internal verification of emission estimates as well as external acceptance testing of system integrity and functionality was undertaken during the development of the AGEIS. Emissions estimated by the AGEIS were compared with those previously reported using traditional spreadsheets to ensure emissions were calculated correctly, that parameter and emission units were correctly recorded and that data was correctly aggregated from lower to higher reporting levels. Implementation of new estimation methodologies are undertaken using a dual estimation approach which ensures that AGEIS emission estimates are verified independently.

Australia's QA-QC Plan is designed to align with the requirements of IPCC Guidelines. The set of tier 1 QC procedures for the inventory compilation process specified in the IPCC *Good Practice Guidance* (2000, 2003) along with the relevant control measure reference in Australia's QA-QC Plan are identified in Table 1.4.

Table 1.4: Implementation of tier 1 quality control checks

Tier 1 QC activity: Checks ^(a)	Control Measure ^(b)	Implementation / Comment
Assumptions and criteria for the selection of activity data and EFs documented	3.E.1	Documented in the NIR
Transcription errors in data input and reference	2.A.1-3, 2.B.2.	Errors checked for using internal AGEIS data verification checks. AGEIS fully integrated with the UNFCCC CRF Reporter Tool removing risk of errors in CRF tables. Error checks are also implemented during the pre-processing of input data Bibliographical data references checked for correct citation.
Emissions are calculated correctly	3.A, 3.B, 3.C	Extensive testing during AGEIS development phase and when new methods introduced. Selected dual estimation process using traditional spreadsheets.
Parameter and emission units are correctly recorded and that appropriate conversion factors are used	3.A, 3.B, 3.C	Extensive testing during AGEIS development phase and when new methods introduced. Selected dual estimation process using traditional spreadsheets.

Tier 1 QC activity: Checks ^(a)	Control Measure ^(b)	Implementation / Comment
Integrity of database files	3.A1-3	Extensive verification/external acceptance testing during the AGEIS development phase. Selected dual estimation process using traditional spreadsheets. Database system and operation documentation updated and archived.
Consistency in data between source categories	3.A.1-3	Parameters (activity data, constants, EF) which are common to multiple sources are entered into global or general data tables so data is only entered once into database.
Movement of inventory data among processing steps is correct	3.A.1-3	Extensive testing during AGEIS development phase and when new methods introduced. Standard reconciliation reports are run to ensure correct aggregation of emission estimates
Uncertainties in emissions and removals are estimated or calculated correctly		Independent review by CSIRO completed
Time series consistency/ Methodological and data changes resulting in recalculations	3.C, 3.D	Where changes are made to methods or activity data the full time-series of emissions is recalculated, the AGEIS ensures consistent use of methods across time-series.
Completeness	2.B.1-2, 3.B.1-4	Checked through CRF Reporter Tool. Mass balance checks undertaken for fuel, carbonates, biomass and synthetic gases
Trend	3.D.1-2	Activity data, emissions and implied emissions factors are compared with previous years estimates and across time series through the AGEIS and CRF Reporter Tool
Review of internal documentation	3.E 1-3	All activity data, emission factors and algorithms are archived within AGEIS. Past inventories may be reproduced using AGEIS. Electronic and hard copies of each year's NIR and methodology are kept in a safe. All bibliographical data references are archived within the AGEIS and in a hardcopy library.

(a) Source: IPCC 2003, Table 5.5.1, page 5.51 (b) References refer to numbering in Australia's QA/QC Plan (see Annex 6)

Table 1.5: Results of reconciliation quality control objectives

Test	Objective (% difference)	Result
CM 3.B.2 (i) Reconciliation between national inventory and sum of State and Territory inventories	<0.001	Achieved
CM 3.B.2 (ii) Reconciliation between national inventory and national inventory by economic sector	<0.001	Achieved
CM 3.B.1 (i) Carbon balance: Reconciliation of data submitted into the AGEIS and national inventory: fossil fuel consumption	<0.001	Achieved
CM 3.B.1 (ii) Carbon balance: Reconciliation of data submitted into the AGEIS and national inventory: carbonates consumption	<0.001	Achieved
CM 3.B.1 (iii) Carbon balance: Reconciliation of data submitted into the AGEIS and national inventory: biomass consumption	<0.001	Achieved
CM 2.B.1 (i) Completeness: Reconciliation of data submitted into AGEIS and reference data: fossil fuels consumption	<1	Achieved
CM 2.B.1 (ii) Completeness: Reconciliation of data submitted into AGEIS and reference data: carbonates consumption	<10	Achieved
CM 2.B.1 (iii) Completeness: Reconciliation of data submitted into AGEIS and reference data: biomass consumption	<1	Achieved
CM 2.B.1 (vi) Completeness: Reconciliation of data submitted into AGEIS and reference data: synthetic gas consumption	<1	Achieved

Tier 2 quality control checks

Category-specific QC (Tier 2) checks are conducted for all sectors to test for completeness, international comparability and verification of country-specific parameters.

Completeness and accuracy are tested through the operation of mass balance checks. The application of mass balance constraints for carbon in fuels, carbonates, biomass wastes and hydrofluorocarbons and nitrogen balances for domestic and commercial wastewater constitute tier 2 quality control measures. All carbon entering the economy in fuels is accounted for—either as emissions from fuel combustion, emissions from the use of fossil fuels as reductants, non-energy uses, use of biomass sources of energy or international bunkers. Carbon balances for biomass, carbonates and synthetic gases consumption have also been implemented. The results of these checks against the principal quality objectives are set out in Table 1.5. Detailed results of the application of these balances are reported in Annex 6.

International comparability of emission estimates is systematically tested through comparisons of the implied emission factors obtained for significant sources of the Australian inventory with the distribution of implied emission factors for all other Annex I parties. The results of these analyses are included in the QA-QC discussions of individual sources in this report.

For the energy, industrial processes and waste sectors, systematic verification tests are undertaken for country-specific parameters such as emission factors utilising data collected under NGERS. The tests are undertaken in accordance with the decision tree (Figure 1.3). Country-specific parameters are tested against NGERS datasets that meet the prescribed conditions. If the mean of the NGERS dataset is significantly different to the country-specific parameter, then the parameter may be revised to reflect the new information.

The empirical research program set out in the National Inventory Improvement Plan is designed to generate information to provide the basis for verification tests for parameters in either tier 2 methods or tier 3 where private measurement activity is not undertaken (see section 10.4 for more details).

In addition, country-specific parameters may also be subjected to verification tests on an ad hoc basis as new information is obtained.

1.6.2 Quality Assurance procedures implemented

Australia's quality assurance systems operate at a number of levels. Quality assurance controls that are implemented annually include:

- the preparation of the Inventory is oversighted within DCCEE by the National Inventory Systems Executive Committee;
- the review of the National Inventory Report, prior to submission to the UNFCCC, by the National Greenhouse Gas Inventory Committee which comprises representatives of State and Territory Governments and the CSIRO. This is the principal formal external review mechanism for the report before it is finalised;
- reconciliation with DCCEE estimates by external consultants for specified sectors – electricity within energy, mineral, chemical and metal products within industrial processes and agricultural soils within agriculture;
- LULUCF sector-specific QA processes employed for the National Carbon Accounting System. These processes are document in Chapter 7 Appendix J;
- the inventory is potentially subject to audit by the Australian National Audit Office (ANAO). The ANAO is an independent office established under The Auditor-General Act 1997. It conducts performance audits of government agencies operating under the Standard on Assurance Engagements ASAE 3500 Performance Engagements issued by the Australian Auditing and Assurance Standards Board (AUASB). ANAO reports are tabled in the Australian parliament and subject to review by the Joint Committee of Public Accounts and Audit (JCPAA). The (ANAO) undertook a performance audit of the national inventory in 2009.

- opening the inventory emission estimates and methods for public review through the release of transparent and easily accessible information via the AGEIS webpage. Industry and public feedback is encouraged through an email facility to the Inventory contact point nationalgreenhouseaccounts@climatechange.gov.au.
- feedback from the public on key emission factors used has been ascertained through a public consultation process – review of the NGER Measurement Determination: Discussion paper, August 2010.
- UNFCCC expert review team processes which aim to review and improve the quality of all Annex I inventories in an open and facilitative manner. Australia's inventory has been reviewed by in-country teams in 2002, 2005, 2008 and 2010, with centralised reviews utilised in other years. Annex 6 documents how the recommendations from the 2011 review report have been implemented or will be addressed in the future.

Specific reviews of sectoral methodologies that have been performed by expert consultants that are not involved in the inventory preparation process are described in Table 1.6.

Table 1.6: Expert reviews of methodologies and activity data

Year of Review	Categories reviewed
2002-2003	4A Enteric Fermentation and 4B Manure Management. (CSIRO, ASIT Consulting, QDNRME, Hassell and Associates Pty. Ltd)
2004	Review of Savanna burning (CSIRO)
2005-06	Emission factors for liquid fuels (GHD Pty Ltd)
2006	Methodologies in the iron and steel and petroleum refining sectors (GHD Pty Ltd) Industrial wastewater and waste incineration methodologies (O'Brien Consulting) Flooded decommissioned coal mines (L. Lunarzewski, Consultant)
2007	Review of Industrial Processes sector (M. Tsaranu, international expert from UNFCCC reviewer roster) Review of Waste sector (Hyder Consulting 2007a,b)
2008	Review of key FullCAM model parameters and assumptions in the LULUCF sector (M. Apps, W Kurts, P. Smith and Q. Zhang, international experts from UNFCCC review roster and/or authors of IPCC guidelines)
2009	Review of waste generation and disposal improvements; and Review of DOC _f values (S. Guendehou, international expert from UNFCCC reviewer roster)

Additional quality assurance activities that have been undertaken include the:

- review of uncertainty estimates by the CSIRO Atmospheric Research Division in 2005 according to strict review protocols.
- review of AGEIS which is managed by DCCEE in accordance with the COBIT framework. The system has been subject to review by Protiviti Consulting in May 2007 and was examined as part of the ANAO performance audit in 2009.

1.6.3 Verification Activities

Australia does not attempt to provide systematic verification through atmospheric concentration estimates given likely scale and related complexities, uncertainties and costs. The CSIRO, however, operates a Baseline Air Pollution Station at Cape Grim in Tasmania. Data on the concentrations of synthetic gases – HFCs and PFCs – have been collected and have been analysed with the aim of providing an independent assessment of emissions of these gases in Australia (see chapter 4). The Australian inventory is tested extensively for comparability with the inventories of other Annex I parties. The implied emission factors and other key parameters for specified variables are reviewed for comparability against the implied emission factors for all other Annex I parties. Specific t-tests are performed to test whether the implied emission factors derived from the Australian inventory are significantly different to the mean of all other

Annex I parties. The results of these tests are recorded in the *National Inventory Systems: Evaluation of Outcomes* document.

As the Australian inventory is in transition to a mix of tier 2/tier 3 methods for many sectors, future verification activities will focus on the development of assessments of tier 3 emission outcomes against the results of associated tier 2 models.

1.6.4 Treatment of confidentiality

The reporting of emissions from some *industrial processes* sources is subject to confidentiality restrictions. Activity data for soda ash production and use, magnesia production, nitric acid production, ammonia production, acetylene use and N₂O use in aerosols and anaesthesia are commercial-in-confidence and, due to the direct relationship between activity and emissions, emissions estimates by gas species have been reported at aggregated levels.

1.7 GENERAL UNCERTAINTY EVALUATION

Uncertainty is inherent within any kind of estimation, be it an estimate of the national greenhouse gas emissions, or the national gross domestic product. Managing these uncertainties, and reducing them over time, is recognised by the IPCC *Good Practice* reports (IPCC 2000, 2003) as an important element of inventory preparation and development. While it is in some cases possible to continuously monitor emissions, it is not usually practical or economic to do so. This leads to estimations based on samples or studies being used which carry a degree of additional uncertainty attached to them. Uncertainty also arises from the limitations of the measuring instruments, and the complexity of modelling highly variable sources of emissions over space and time, particularly for some biological sources.

Australia has conducted uncertainty analysis across the sectors of *energy*, *industrial processes*, *agriculture*, *land use*, *land use change* and *waste* in line with the IPCC *Good Practice* reports (2000, 2003). These estimates have been subjected to an independent review by the CSIRO.

Emission estimate uncertainties typically are low for carbon dioxide from energy consumption as well as from some industrial process emissions ($\pm 4\text{--}5\%$). Uncertainty surrounding estimates of emissions are higher for agriculture, land use, land use change and forestry and synthetic gases. A medium band of uncertainty applies to estimates from fugitive emissions, most industrial processes and non-CO₂ gases in the energy sector.

The sectoral estimates presented in Annex 7 show that the uncertainty ranges reported for the various components of the Australian inventory are largely consistent with the typical uncertainty ranges expected for each sector, as identified in the IPCC *Good Practice* reports.

At an aggregate level, using IPCC good practice methods (tier 1), the overall uncertainty surrounding the Australian inventory estimate for 2009 is estimated at $\pm 5.7\%$. The reported uncertainty for the trend in emissions is estimated to be $\pm 8.2\%$. When the LULUCF sector is excluded from the analysis the uncertainty is estimated at $\pm 2.2\%$ for the 2009 inventory estimate and $\pm 2.0\%$ for the trend in emissions. The IPCC approach provides accurate estimates of uncertainty under certain restrictive assumptions that do not always hold for most countries' inventories. Consequently, the Department of Climate Change and Energy Efficiency is planning to undertake more extensive tier 2 analyses of uncertainty using Monte Carlo analyses in its future work programme.

1.8 GENERAL ASSESSMENT OF COMPLETENESS

The inventory is considered to be largely complete with only a few minor sources not estimated, due to either a lack of available information or methodology in the IPCC Good Practice Guidance. These sources are considered to be insignificant, however, when compared with the inventory as a whole.

The assessment of completeness is used to help identify areas where methodologies can be developed, and where additional sources of data may be sought. For example, new data has become available which has allowed an improvement in the completeness of certain sources for this inventory. In particular, the collection of data through NGERS has enabled an improvement in the completeness of 2.A.3 other carbonate consumption due to improved coverage of companies as a result of the self-identification provisions of the NGERS, compared with the more limited coverage of the previous voluntary survey conducted for the inventory.

Data collected under the *NGER Act 2007* has also enabled the estimation of fugitive CO₂ emissions from underground coal mining activities and CO₂ from food and drink production which were reported for the first time in the *National Inventory Report 2008*.

In order to provide assurance as to the completeness of the activity data used in the estimation of emissions for the inventory DCCEE also undertakes Tier 2 QA/QC checks of fossil fuels, biomass and carbonate materials. These checks ensure that the estimated supplies of carbon into the economy are reconciled with the estimated uses of carbon, ensuring that any bottom-up data collection processes are reconciled with data available at aggregate levels.

1.9 GEOGRAPHICAL COVERAGE

The Australian inventory covers the six states (New South Wales, Victoria, Queensland, South Australia, Western Australia and Tasmania), the mainland territories (Northern Territory, Australian Capital Territory and Jervis Bay Territory) and the associated coastal islands.

The geographical coverage of the Australian inventory also includes emissions from the following external territories:

- Australian Antarctic Territory
- Norfolk Island
- Christmas Island
- Cocos Islands
- Heard and McDonald Islands.

The following external territories are also covered but are included in the state statistical territories by the Australian Bureau of Statistics:

- Coral Sea Islands (Queensland)
- Ashmore and Cartier Islands (Northern Territory).

The coverage of emissions/removal categories for the external territories is as follows:

- fuel combustion, waste and HFC emissions associated with refrigeration are estimated;
- fugitive and industrial processes emissions are assumed to be not occurring;
- agriculture and land use, land use change and forestry emissions and removals are not estimated but are likely to be negligible.

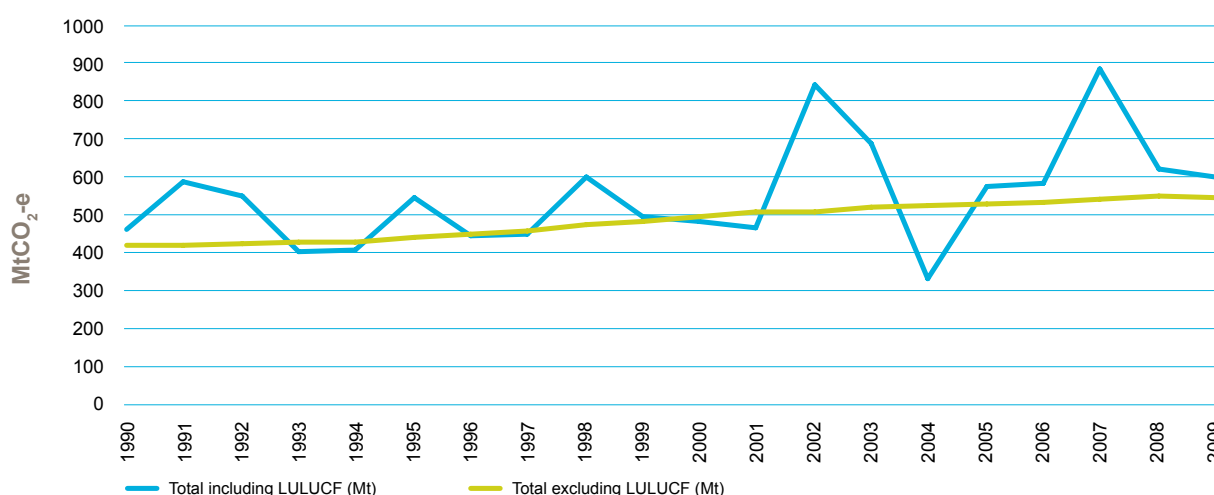
2. TRENDS IN GREENHOUSE GAS EMISSIONS

2.1 EMISSION TRENDS FOR AGGREGATED GREENHOUSE GAS EMISSIONS

Australia's total greenhouse gas emissions excluding the LULUCF sector were 545.8 million tonnes (Mt) CO₂-equivalent (CO₂-e) in 2009. This represents a decrease of 5.0 Mt or 0.9 % on net emissions recorded in 2008 and an increase of 30.5 % (127.5 Mt) above 1990 levels (Figure 2.1).

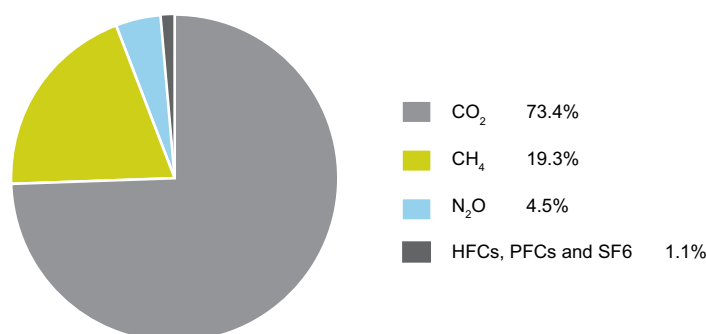
When the *land use, land use change and forestry* sector emissions and removals are included in the total, Australia's net greenhouse gas emissions in 2009 were 599.8 Mt CO₂-e³. This represents an increase of 138.3 Mt or 30.0% on net emissions recorded in 1990. From year to year the LULUCF sector may change from a net source to a net sink. In 2009, the LULUCF sector was a net source of 54.0 Mt. The trends in the LULUCF sector are primarily driven by inter-annual climate variability and natural disturbances which mask other, underlying patterns in the sector directly associated with human activities.

Figure 2.1: Australia's total net greenhouse gas emissions from 1990 to 2009



Carbon dioxide is the most important of the greenhouse gases in Australia's inventory (excluding LULUCF) with a share of 73.4 % (400.3 Mt) of the total CO₂-e emissions, followed by methane, which comprises 20.6% (112.7 Mt CO₂-e) (Figure 2.2). The remaining gases make up 6.0 % (32.7 Mt CO₂-e) of Australia's greenhouse gas emissions.

Figure 2.2: Contribution to total net CO₂-e emissions (excluding LULUCF) by gas, 2009



³ This total does not reflect Australia's emissions and removal under the Kyoto Protocol accounting framework as the inclusion of LULUCF emissions and removals are restricted to a subset of activities (i.e. Afforestation, Reforestation and Deforestation which have occurred since 1990).

The combined *energy* subsectors (including stationary energy, transport and fugitive emissions) were the largest source of greenhouse gas emissions in 2009 comprising 76.5 % of emissions (Table 2.1) followed by the *agriculture* sector (15.5 %).

Table 2.1: Australia's net greenhouse gas emissions by sector under the UNFCCC, 2009

Sector and Subsector	CO ₂	CH ₄	N ₂ O	HFCs	PFCs	SF ₆	Total	Share of total
	Mt CO ₂ -e							%
1 All energy (combustion + fugitive)	380.7	33.7	2.9	NA	NA	NA	417.4	76.5%
<i>Stationary energy</i>	291.8	1.2	1.0	NA	NA	NA	294.1	53.9%
<i>Transport</i>	81.4	0.5	1.8	NA	NA	NA	83.6	15.3%
<i>Fugitive emissions from fuel</i>	7.6	32.0	0.0	NA	NA	NA	39.7	7.3%
2 Industrial Processes	19.6	0.1	3.4	6.2	0.3	0.1	29.6	5.4%
3 Solvent and other product use ^(a)	NA	NA	IE	NA	NA	NA	IE	IE
4 Agriculture	NA	65.3	19.5	NA	NA	NA	84.7	15.5%
6 Waste	0.0	13.6	0.4	NA	NA	NA	14.1	2.6%
Total net emissions (excluding LULUCF)	400.3	112.7	26.7	6.2	0.3	0.1	545.8	100.0%
5 Land use, land use change and forestry	49.5	3.1	1.3	NA	NA	NA	54.0	9.0%
Total net emissions (including LULUCF)	449.9	115.8	27.5	6.2	0.3	0.1	599.8	100.0%

(a) confidential N₂O emissions from *solvent and other product use* are reported under *industrial processes*

2.2 EMISSION TRENDS BY GAS

Of the individual greenhouse gases, emissions of carbon dioxide increased by 43.9% between 1990 and 2009. Nitrous oxide emissions increased by 37.9%; HFC emissions increased by 454.9%; PFC emissions fell by 92.2%. Methane emissions fell by 2.9%. SF₆ emissions increased by 7.4%.

Table 2.2: Change in total net CO₂-e emissions by gas (excluding the LULUCF sector), 1990–2009

Greenhouse gases	1990 Mt CO ₂ -e	2009 Mt CO ₂ -e	1990% of Total	2009% of Total	Changes Mt	% Change in emissions
CO ₂	278.2	400.3	66.5%	73.4%	122.2	43.9%
CH ₄	116.0	112.7	27.7%	20.6%	-3.3	-2.9%
N ₂ O	18.9	26.1	4.5%	4.8%	7.2	37.9%
HFCs	1.1	6.2	0.3%	1.1%	5.1	454.9%
PFCs	4.0	0.3	0.9%	0.1%	-3.6	-92.2%
SF ₆	0.1	0.1	0.0%	0.0%	0.0	7.4%
Total CO₂-e	418.3	545.8	100%	100%	127.5	30.5%

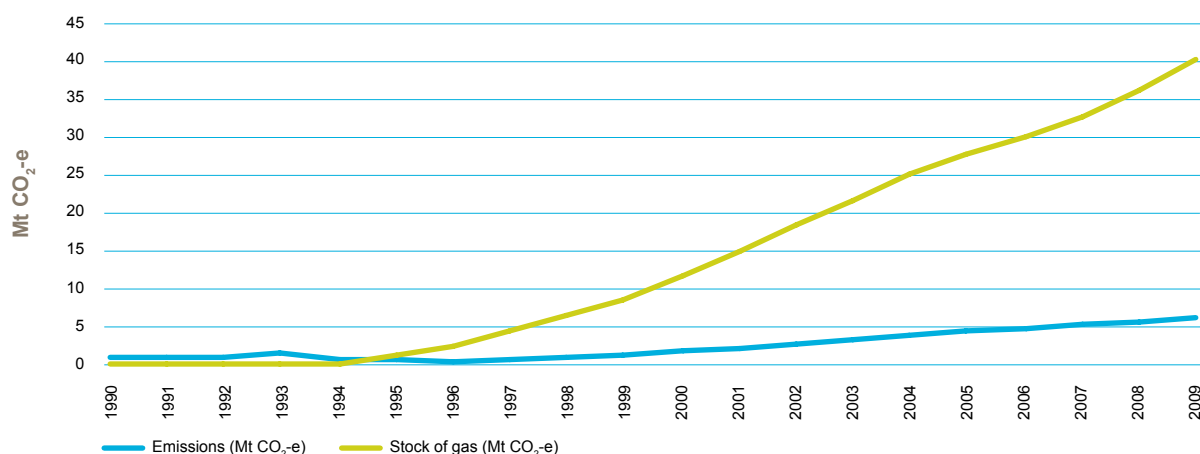
The energy sector is the major contributor to carbon dioxide emissions at 95.1 % (380.8 Mt). The growth in CO₂ represents the increased emissions from the energy sector particularly in electricity generation and road transport.

Agriculture is the main contributor of methane (57.9 %, 65.3 Mt) and nitrous oxide (74.5 %, 19.5 Mt) emissions. The growth in the N₂O emissions is driven by increases in Industrial Process emissions and also by increased emissions from the manure of intensively managed livestock and increased use of nitrogenous fertilisers in the agriculture sector.

The other greenhouse gases, HFCs and PFCs, which contribute 1.2 per cent of the total emissions, have undergone large relative changes between 1990 and 2009. Emissions of PFCs have principally decreased due to improvements in aluminium smelting processes. HFC emissions have increased because of the use of HFCs as a substitute for chlorofluorocarbons (CFCs) phased out under the Montreal Protocol.

The nature of the release of HFC emissions (gradual leakage) means they will be a growing influence on Australia's National Inventory total for some time. The stock of HFCs has been strongly growing since their introduction in 1994 (Figure 2.3). A small proportion of this stock is emitted each year through leakage however the remaining stock continues to increase and will be emitted in the future. Due to the gradual leakage rates it will take many years for HFC emissions to stabilise.

Figure 2.3: Halocarbons: estimated stock and emissions: all equipment types



Other gases and sectors such as methane emissions from solid waste disposal on land and decommissioned coal mines also have extended lifetimes to their emissions profile although the growth rate in emissions from these gases is less than the rate for HFCs.

2.3 EMISSION TRENDS BY SOURCE

Sectors with increasing emissions over the 1990 to 2009 period included *stationary energy* (up 51.1%), *transport* (up 34.6%), *fugitive emissions from fossil fuels* (up 22.9%), *industrial processes* (up 21.0%) and *land use, land use change and forestry* (up 25.1%). *Waste* (down 21.9 %) and *agriculture* (down 2.4%) emissions have decreased.

Figures 2.4 and 2.5 show the absolute annual change for each sector from 1990-2009. The principal drivers of these emission trends are as follows:

- **Energy:** The largest sectoral increase in greenhouse gas emissions over the 1990 to 2009 period, of 51.1% (99.4 Mt CO₂-e), occurred in the *stationary energy* sector, driven in part by increasing population, household incomes and export increases from the resources sector. The main driver for the increase in *transport* emissions is continuing growth in household incomes and numbers of vehicles. *Fugitive emissions* have increased due to increased production from open cut and underground coal mines.
- **Industrial processes:** The increase in emissions since 1990 is primarily driven by the growth in emissions associated with HFCs and the chemical industries.
- **Agriculture:** Between 1990 and 2001 emissions increased due to increased fertiliser use and savanna burning. Since 2002, southern and eastern Australia has experienced widespread drought conditions which has contributed to reductions in animal populations, crop production, fertiliser use, and associated emissions.

- *Waste*: The net emissions from waste have decreased as increases associated with growing populations and industrial production have been offset by increased methane recovery.
- *LULUCF*: The LULUCF sector changes between a net source and a net sink a number of times through the time-series. The trends in the LULUCF sector are primarily driven by inter-annual climate variability and natural disturbance. The trend in net emissions in the early 1990s is driven by the reduction in forest conversion emissions, however the shift from a net sink to net source in 1995, 1998 and 2002 and the large spike in 2007 are due to natural disturbances such as fires and extensive drought conditions which caused a loss of carbon from all pools.
- Trends in emissions from each sector are discussed further in chapters 3-8.

Figure 2.4: Absolute annual change in CO₂-e emissions by sector, 1990-2009

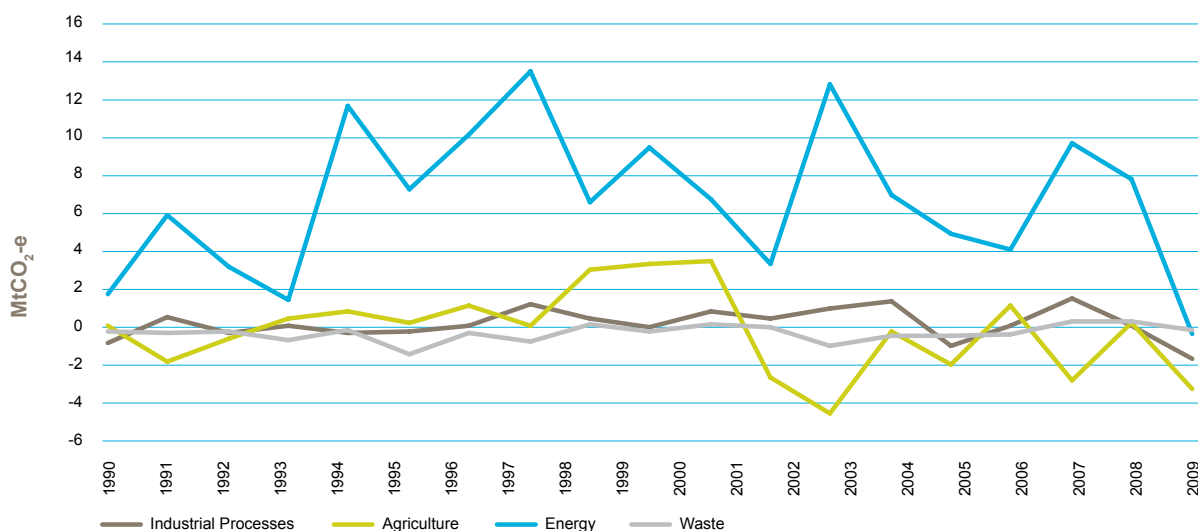
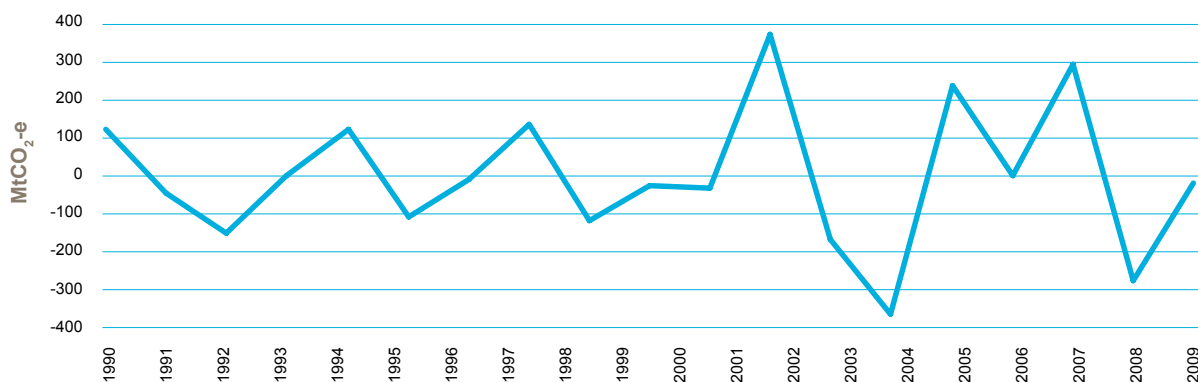


Figure 2.5: Absolute annual change in CO₂-e emissions in the LULUCF sector, 1990-2009



Emissions from some sectors are heavily reliant on the existing capital stock which influences emissions. An example of this is evident when analysing non-CO₂ emissions from road transport. In recent years updated Australia Design Rules were introduced requiring new vehicles to emit lower levels of pollutants. However, non-CO₂ emissions from this sector have reduced only gradually over the time series because it takes many years to turn over the existing vehicle stock. Figures 2.6 and 2.7 show the emissions of CH₄ and N₂O from road transportation in Australia. The gradual decrease in emissions in recent years is a result of the gradual replacement of the vehicle fleet to vehicles manufactured under more recent Australia Design Rules.

Figure 2.6: CH₄ emissions from road transportation in Australia

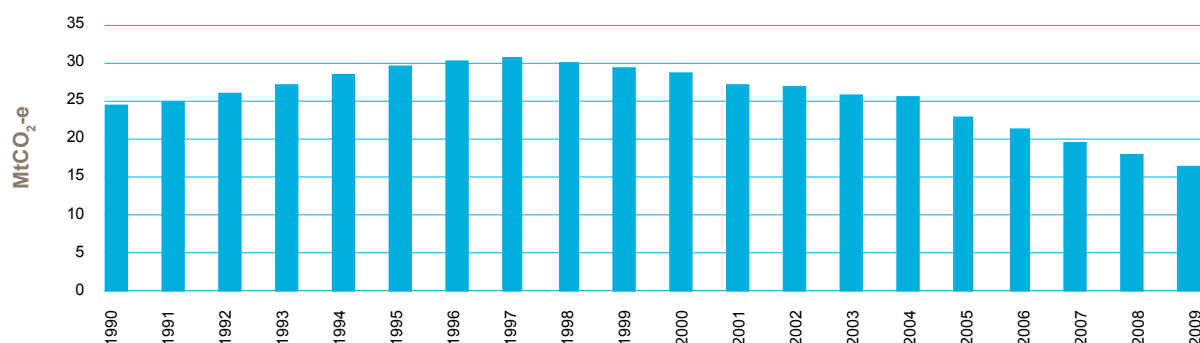
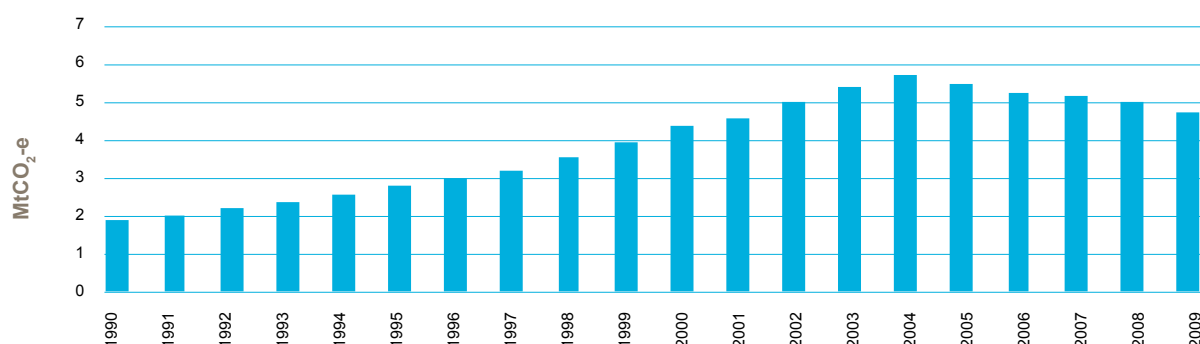


Figure 2.7: N₂O emissions from road transportation in Australia



2.4 EMISSION TRENDS FOR INDIRECT GREENHOUSE GASES AND SO₂

The indirect greenhouse gases NO_x, CO, NMVOC and SO₂ are also reported but as they have not been allocated global warming potentials they are not included within Australia's total aggregated emissions. Since 1990 there have been increases in the emissions of CO, NO_x, SO₂ and NMVOCs (Table 2.3).

The main source of NO_x is the *energy* sector, with the largest single source road transportation. The *agriculture* and *LULUCF* sectors are the largest producers of CO and NMVOCs. These gases, along with NO_x, are emitted during biomass burning. Emissions of SO₂ occur in the *industrial processes* and *energy* sectors. The principal sources being other metal production (2.C.5) (69.5 %) and electricity and heat production (1.A.1.a) (24.2%)

Table 2.3: Emissions of indirect greenhouse gases and SO₂ in 1990 and 2009

Greenhouse gas	1990	2009	Change from 1990 (Mt)	Change from 1990 (%)
NO _x	1.8	2.5	0.7	38%
CO	20.8	23.6	2.8	13%
NMVOC	2.6	2.6	-0.1	2%
SO ₂	1.6	2.6	1.1	68%

Including emissions from the LULUCF sector.

2.5 EMISSION TRENDS FOR KYOTO PROTOCOL

Australia's initial assigned amount⁴ is equivalent to 591.5 Mt CO₂-e for each year of the commitment period (2008-2012). The initial assigned amount is based on the revised 2005 inventory submitted in 2008 and reported in Australia's *Initial Report under the Kyoto Protocol, revised submission* (DCC 2008). The initial assigned amount is fixed and is the value that Australia's Kyoto Protocol emissions account will be compared with for the first commitment period.

- Under the KP accounting rules, Parties must report emissions from the Article 3.1 (Annex A) sectors, namely the *energy, industrial processes, solvents, agriculture* and *waste* sectors. However, only a subset of the LULUCF sector emissions and removals are included in the account. These include deforestation, afforestation and reforestation activities that have occurred since 1990 (Article 3.3 activities which are mandatory) and forest, cropland or grazing land management and revegetation (Article 3.4 activities which can be elected by the Party). Australia has not elected to account for any Article 3.4 activities for the first commitment period.

In 2009 *afforestation/reforestation* activities on unharvested land units contributed a net sink of 22.6 Mt CO₂-e while harvested land units contributed a net source of 7.6 Mt CO₂-e (Table 2.4). *Deforestation* activities were a net source of 41.3 Mt CO₂-e in 2009.

Table 2.4: Net emissions and removals from Article 3.3 LULUCF activities by gas in 2009

KP-LULUCF Activities	CO ₂	CH ₄	N ₂ O	Total
	Mt CO ₂ -e			
Afforestation/Reforestation				
Units of land not harvested during the commitment period	-22.6	0.0001	0.00004	-22.6
Land units harvested during the commitment period ^(a)	7.6	0.0006	0.0002	7.6
Deforestation	39.6	1.2	0.5	41.3

(a) These are the estimated emissions and removals not the accounting quantities

Australia has elected to account for the Article 3.3 activities on an annual basis. In accordance with the harvested forest sub-rule (paragraph 4 of the annex to decision 16/CMP.1) in calculating the *afforestation/reforestation* accounting quantity for 2009, the emissions on harvested land units have been set to zero where the emissions are greater than the removals on those lands units (Table 2.5). The total net emissions associated with Kyoto Protocol Article 3.1 sectors and Article 3.3 LULUCF activities in 2009 were 564.5 Mt CO₂-e.

⁴ The initial assigned amount (2,957,579.143 Gg CO₂-e) is calculated as five times 108% of base year emissions. The base year emissions (547.7 Mt CO₂-e) are the 1990 emissions from the annex A sectors and land use change as reported in the revised 2005 inventory.

Table 2.5: Emissions and removals associated with Articles 3.1 and 3.3 of the Kyoto Protocol, 2009

Sector and Subsector	Emissions Mt CO ₂ -e
Article 3.1 – Annex A sectors	
1 Energy	417.4
2 Industrial Processes	29.6
3 Solvent and other product use ^(a)	IE
4 Agriculture	84.7
6 Waste	14.1
Inventory Total	545.8
Article 3.3 – KP LULUCF activities^(b)	
Afforestation/Reforestation ^(c)	-22.6
Deforestation	41.3
Kyoto Protocol Total	564.5

(a) Included elsewhere – emissions are included in *industrial processes* for confidentiality reasons. (b) Australia has elected to account for Article 3.3 activities on an annual basis. (c) Accounting quantity in accordance with paragraph 4 of the annex to decision 16/CMP.1.

2.5.1 Latest Quarterly Trends

The Department of Climate Change and Energy Efficiency publishes preliminary estimates of emissions for each quarter. Up to date estimates of Australian emissions trends are produced by supplementing the estimation methodologies documented in the National Inventory Report with preliminary activity and economic data from government and industry sources.

Preliminary emissions estimates for the year to June 2010 showed Australia's emissions, excluding Land Use, Land Use Change and Forestry (LULUCF) activities under article 3.3 of the Kyoto Protocol, were 545Mt CO₂-e (Table 2.6); down 0.1% on the previous year.

Table 2.6: Preliminary National Inventory, for the four quarters to June quarter 2010^a

Category	Annual emissions through to the June quarter Mt CO ₂ -e ^b
<i>National Inventory – Annex A sectors</i>	
Energy – Electricity	201
Energy – Stationary energy excluding electricity	90
Energy – Transport	84
Energy – Fugitive emissions	43
Industrial processes	32
Waste	14
Agriculture	82
<i>National Inventory (excluding LULUCF)</i>	545

(a) These numbers are preliminary as more data becomes available from the Department's reference sources these preliminary activity data will be replaced and the estimates of emissions revised before submission to the UN. (b) Annual emissions through to the June quarter includes emissions from July 2009 to June 2010.

Trend emissions fell by 0.5% in the June quarter on the March quarter as emissions from electricity generation declined by 4.6%, largely from fuel switching to lower emissions technologies in the electricity generation sector. The decline is considered to be temporary. While no data is available in relation to estimates of emissions from the LULUCF sector in 2010, projections contained in Australia's 5th National Communication show deforestation emissions are likely to rebound from levels observed in 2009. Actual deforestation rates will reflect a combination of forces, including responses of land owners to higher commodity prices, recent wet weather and the impact of State regulations on clearing.

Figure 2.8: National Inventory, actual and trend emission estimates – June quarter 2000 to June quarter 2010^c



(c) In Figure 2.8, the actual emissions estimate and the trend emissions estimate for each quarter is shown. Actual emissions will fluctuate during the year as a result of seasonal weather patterns and variations in economic activity.

3. ENERGY

3.1 OVERVIEW

Total emissions from the *energy* sector for 2009 were estimated to be 417.4 Mt CO₂-e (Table 3.1). *Energy industries* were the main contributor, accounting for 54.6% of emissions from the *energy* sector. Other significant contributors to total *energy* emissions were *transport* (20.0%), and *manufacturing industries and construction* (10.7%).

Energy sector emissions increased by 44.4% between 1990 and 2009. Emissions from the sector decreased by 0.3 Mt (0.1%) from 417.6 Mt in 2008. The main contributor to the decrease in emissions was *manufacturing industries and construction*, which decreased by 2.7 Mt CO₂-e.

Table 3.1: Energy sector CO₂-e emissions, 2009

Greenhouse gas source and sink categories	CO ₂ -e emissions (Gg)			
	CO ₂	CH ₄	N ₂ O	Total
1 ENERGY	380,758	33,745	2,852	417,355
A. Fuel combustion activities	373,173	1,707	2,820	377,700
1 Energy industries	226,830	269	695	227,794
a Electricity and heat production	205,851	230	662	206,743
b Petroleum refining	5,155	2	13	5,170
c Manufacture of solid fuels	15,824	38	19	15,881
2 Manufacturing industries and construction	44,208	51	254	44,514
3 Transport	81,374	475	1,796	83,645
a Civil aviation	5,943	1	58	6,002
b Road transportation	70,116	351	1,699	72,166
c Railways	2,367	2	21	2,390
d Navigation (domestic)	2,905	120	19	3,043
e Other transportation	43	0	0	43
4 Other sectors	19,357	910	67	20,334
5 Other	1,404	1	8	1,413
a Lubricants	474			474
b Mobile (military)	930	1	8	939
B. Fugitive emissions from fuels	7,585	32,038	32	39,655
1 Solid fuels	1,027	27,676		28,704
2 Oil and natural gas	6,557	4,362	32	10,951

3.1.1 Stationary Energy

Stationary energy principally comprises fossil fuel combustion in *electricity and heat production* and *manufacturing and construction industries*. Total estimated emissions from *stationary energy* combustion were 294.1 Mt CO₂-e in 2009, equal to 53.9% of net national emissions (excluding LULUCF).

The *energy industries* subsector includes fuel combustion in electricity generation, petroleum refining, gas production and solid fuel manufacture. *Electricity and heat production* (1.A.1.a) contributed 206.7 Mt CO₂-e or 70.3% of *stationary energy* emissions in 2009. This category includes emissions only from electricity generation because heat production as defined by the IPCC does not occur in Australia. Estimated emissions from the remaining *energy industries* subsectors were 21.1 Mt in 2009.

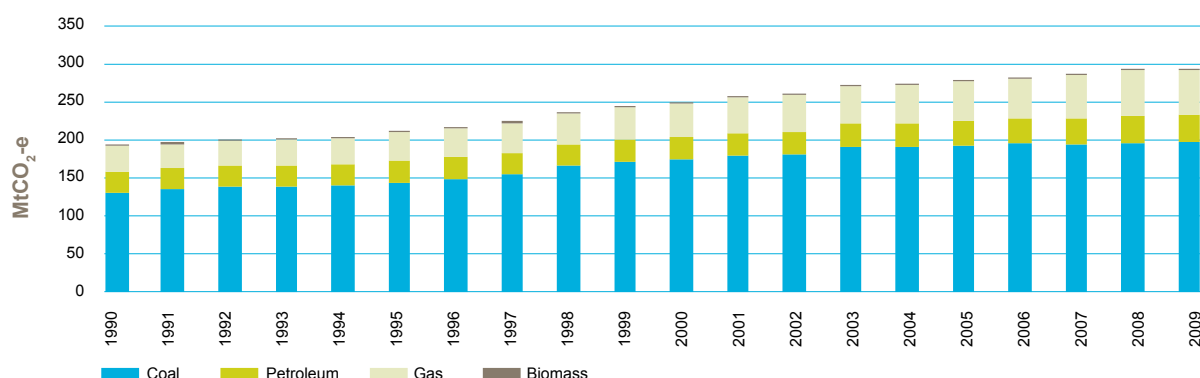
The *manufacturing industries and construction* subsector (1.A.2) emissions were 44.5 Mt CO₂-e in 2009. This subsector includes direct emissions from fuel combustion in manufacturing industries, ferrous and non-ferrous metals production, plastics production, construction and non-energy mining. These calculations do not fully reflect the greenhouse impact of these industries, as the emissions generated from the production of electricity used in these industries are included under *electricity and heat production* (1.A.1.a).

Estimated emissions from *other sectors* (1.A.4) were 20.3 Mt CO₂-e in 2009. This subsector comprises direct fuel combustion in the residential, commercial and institutional sectors, including energy used in mobile equipment in *agriculture, forestry and fishing* industries. However, as with *manufacturing*, much of the greenhouse impact of these sectors arises from their large consumption of electricity, which is not reflected in this figure alone (reported under 1.A.1.a). *Other* (1.A.5) includes emissions from *lubricants* (0.5 Mt) and *military transport* (0.9 Mt). Emissions from *lubricants* arise from the combustion of engine oil in vehicles.

Trends

Emissions from *stationary energy* increased by 51.1% (99.4 Mt) between 1990 and 2009, including an increase in emissions from coal combustion of 50.3% (66.4 Mt) in the same period (Figure 3.1). Although coal accounted for the highest absolute increase in emissions over this period, emissions related to gaseous fossil fuels have shown the largest relative growth, increasing by 79.7% (26.3 Mt) between 1990 and 2009. Emissions from liquid fossil fuels increased by 27.0% (7.5 Mt) in the same period. Biomass emissions decreased by 39.1% (0.7 Mt) between 1990 and 2009. Between 2008 and 2009, emissions from *stationary energy* increased by 0.2% (0.6 Mt).

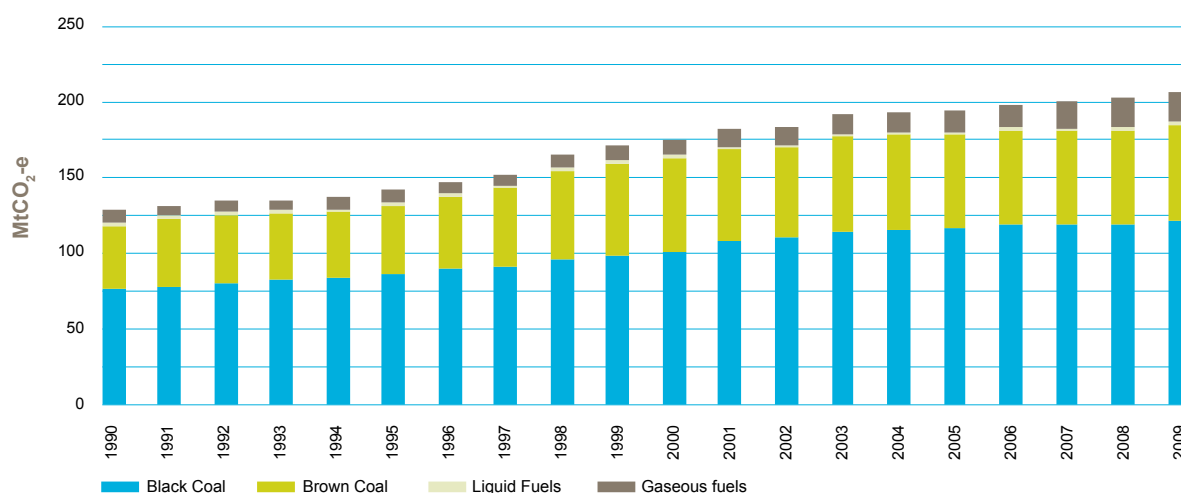
Figure 3.1: Total CO₂-e emissions from stationary energy combustion by fuel, 1990–2009



Electricity generation *emissions* increased by 3.6 Mt (1.8%) from 2008 to 2009, and by 77.3 Mt (59.7%) from 1990 to 2009 (Figure 3.2). The 2008 to 2009 electricity generation emissions increase (1.8%) was greater than the increase in electricity generation (1.1%). This was partially due to an increase in the brown coal proportion of energy use increasing by 0.2% to 27.8%, a continued decline in hydroelectricity generation driven by weather conditions and annual variability in facility specific emission factors.

Emissions from *stationary energy* subsectors, other than *electricity generation*, decreased by 3.0 Mt (3.3%) between 2008 and 2009, and increased overall by 22.1 Mt (34.0%) from 1990 to 2009. Emissions from the *manufacturing industries and construction* subsector decreased 5.7% (2.7 Mt) between 2008 and 2009 and increased by 24.9% (8.9 Mt) from 1990 to 2009.

Figure 3.2: CO₂-e emissions from electricity generation by fossil fuels, 1990–2009



3.1.2 Transport

In 2009 transport contributed 83.6 Mt CO₂-e or 15.3% of Australia's net emissions (excluding LULUCF).

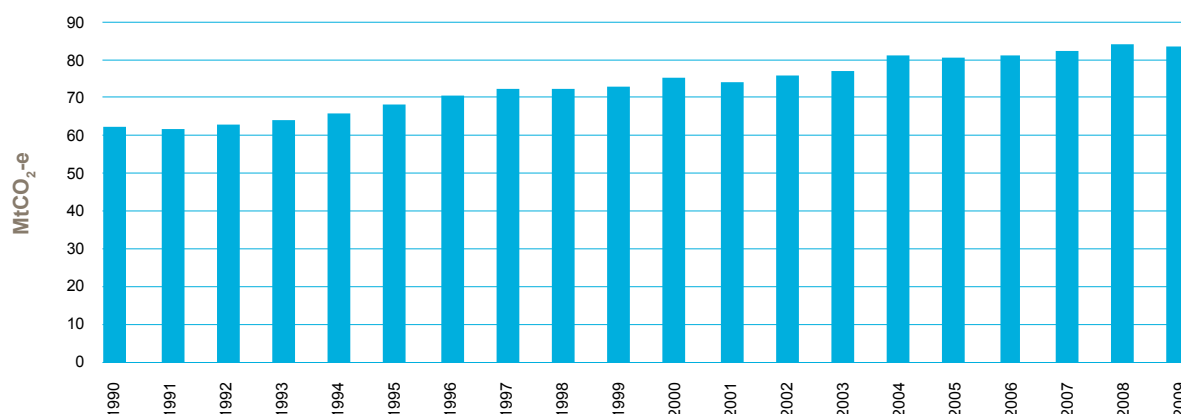
The major source of transport emissions in Australia is road transportation, which accounts for 86.3% (72.2 Mt) of transport emissions. This outcome is principally driven by the importance of motor vehicles as modes of transportation of passengers and freight in Australia. Passenger cars account for 41.5 Mt and trucks (light and heavy) and buses 30.5 Mt. Other sources are far smaller: domestic aviation contributed 7.2% (6.0 Mt), domestic navigation 3.6% (3.0 Mt), and railways 2.9% (2.4 Mt).

Fuel used in *international transport* (*international aviation* and *marine 'bunkers'*) is by international agreement reported separately from the national total net emissions. In 2009, international bunker fuels generated 12.0 Mt of emissions.

Trends

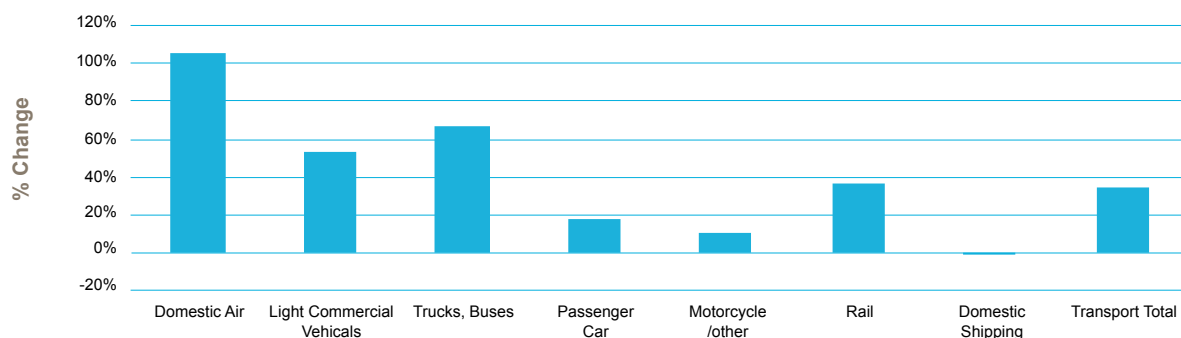
Transport emissions are one of the strongest sources of emissions growth in Australia. Emissions from this sector were 34.6% higher in 2009 than in 1990, and on average have increased by around 1.6% annually (Figure 3.3). Emissions from road transportation increased by 32.8% (17.8 Mt) between 1990 and 2009 (Figure 3.4). Emissions from passenger cars increased by 17.6% (6.2 Mt). Emissions from Light Commercial Vehicles (LCVs) and heavy duty trucks and buses have also grown strongly (53.2% and 66.7% respectively).

Figure 3.3: Total transport emissions, 1990–2009



Estimates of civil aviation (domestic air transport) emissions were 105.3% (3.1 Mt) higher than the 1990 level. Emissions have grown strongly in this sector, particularly in the early 1990s, although emissions in the 1990 base year were unusually low because of an extensive dispute with airline pilots in that year and this has contributed to the magnitude of the change. Emissions from rail have risen by 36.7% while domestic shipping emissions have fallen by 0.2%, reflecting improved productivity and changes in activity.

Figure 3.4: Comparison of growth in transport emissions by subcategory, 1990–2009



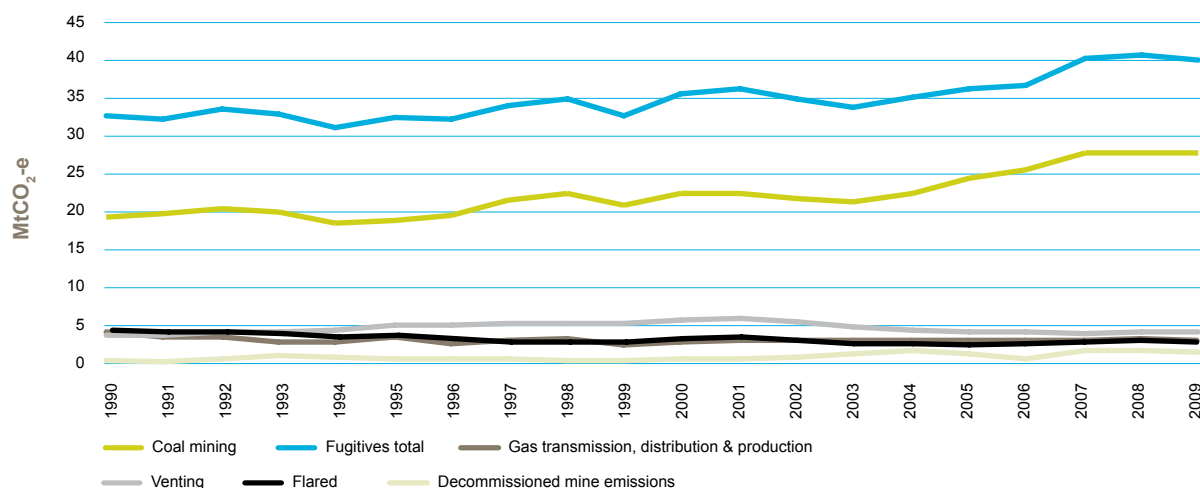
3.1.3 Fugitive Emissions

Total estimated *fugitive emissions* for 2009 were 39.7 Mt CO₂-e, representing 7.3% of net national emissions (excluding LULUCF). Net *solid fuel* emissions contributed 72.4% (28.7 Mt) of *fugitive* emissions. *Oil and natural gas production, processing and distribution* account for the remaining 27.6% (11.0 Mt) of *fugitive* emissions.

Trends

Overall *fugitive* emissions increased 22.9% (7.4 Mt) between 1990 and 2009, and decreased by 0.7% (0.3 Mt) from 2008 to 2009 (Figure 3.5). From 1990 to 2009 fugitive emissions from *solid fuels* increased by 48.8% (9.4 Mt) and *oil and natural gas* emissions decreased by 15.6% (2.0 Mt).

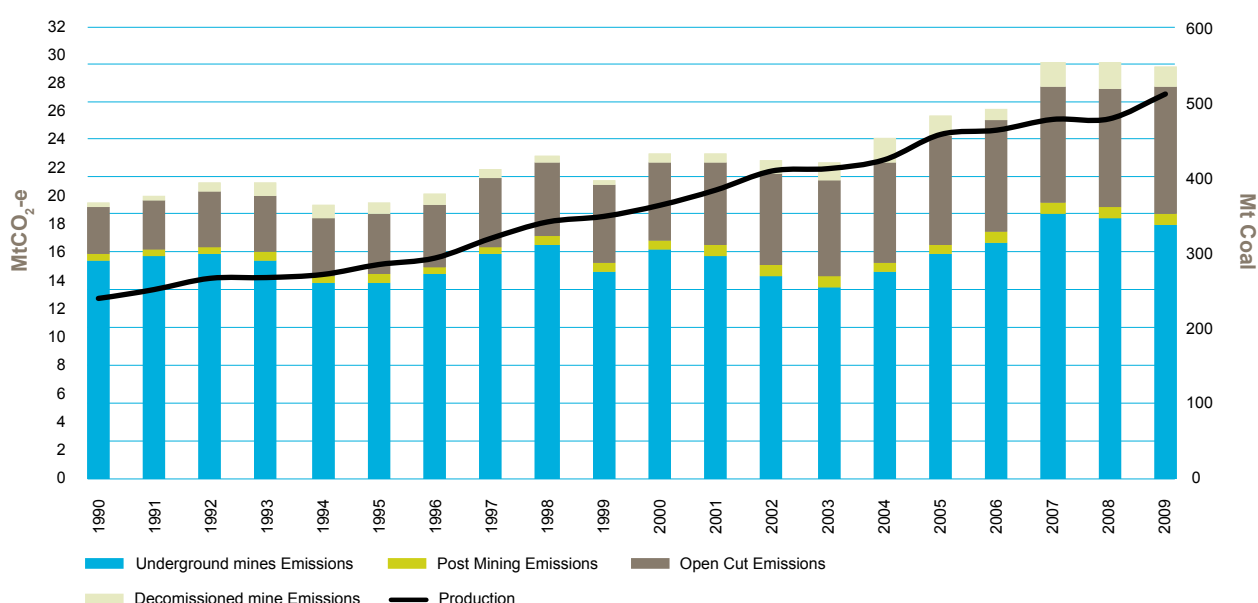
Figure 3.5: CO₂-e fugitive emissions by category, 1990–2009



Solid fuel emissions decreased by 0.6% (0.2 Mt) between 2008 and 2009. Underground mine emissions fell by 2.7% (0.5 Mt), driven by a relative increase in the production mix from less gassy coalfields of Queensland and western New South Wales. Emissions from surface mines increased by 8.0% (0.7Mt) between 2008 and 2009. Emissions from decommissioned mines have decreased 19.3% (0.3 Mt) between 2008 and 2009.

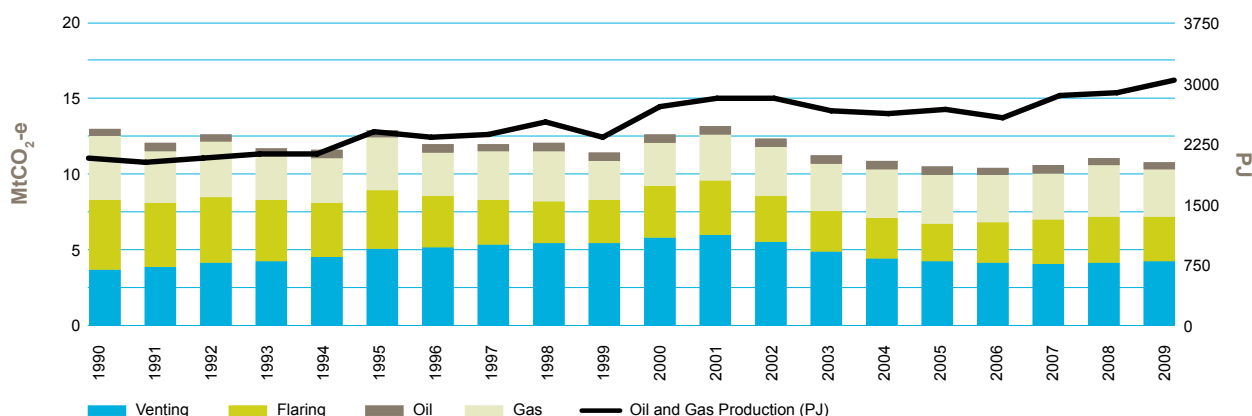
Emissions tend to fluctuate from year to year depending on the volume of coal mined and the share of production from underground mines of varying gas contents. Mine production of coal has increased from 241 Mt in 1990 to 511 Mt in 2009, an increase of 112%. Methane emissions have not grown as fast as activity principally because, since 1998, there has been an increasing trend in activity from surface mines compared to that of underground mines (Figure 3.6) and, within underground mines, a decreasing share of production from the gassiest southern New South Wales coalfield (Figure 3.31). In addition, technologies to recover and utilise coal mine waste gas for electricity generation have been increasingly adopted in underground mining.

Figure 3.6: Fugitive CO₂-e emissions from coal mining activities, 1990–2009



Oil and natural gas fugitive emissions decreased by 15.6% (2.0 Mt) between 1990 and 2009 (Figure 3.7). This compares with a 47% increase in production activity. The decrease in emissions relative to the increase in activity is the result in particular of improvements in gas distribution and a reduction in the emissions from flaring. Between 1990 and 2009, emissions from oil-related activities decreased by 18.5% (0.1 Mt) and emissions from gas-related activities decreased by 20.6% (0.9 Mt). Emissions from venting increased by 2.8% (0.1 Mt) from 2008 to 2009 and compared with 1990, emissions were higher by 15.0% (0.6 Mt). Flaring-related emissions decreased by 2.4% (0.1 Mt) from 2008 to 2009, and emissions in 2009 were lower than 1990 levels by 35.5% (1.6 Mt).

Figure 3.7: Fugitive CO₂-e emissions from oil and gas production, 1990–2009



3.2 OVERVIEW OF SOURCE CATEGORY DESCRIPTION AND METHODOLOGY – ENERGY

The *energy* sector includes emissions from the combustion of fossil fuels (1.A.1 *energy industries*; 1.A.2 *manufacturing industries and construction*; 1.A.3 *transport*; 1.A.4 *other sectors*; and 1.A.5 ‘*other*’) as well as *fugitive emissions from the extraction of fossil fuels* (1.B).

The use of solid, liquid and gaseous fuels for energy use has been identified as key sources in Australia’s inventory.

The methodology for estimating emissions from fossil fuel combustion in the stationary energy sectors is consistent with the IPCC tier 2 approach. Tier 2 methods may be regarded as those dividing fuel consumption on the basis of sample or engineering knowledge between technology types which are sufficiently homogenous to permit the use of representative emission factors. Emissions for the Transport sector have been estimated with a mix of Tier 1, 2 and Tier 3 approaches.

The Australian Bureau of Agricultural and Resource Economics and Sciences (ABARES) compile Australia’s Energy Statistics which estimates Australian energy consumption by fuel and economic sector for the purpose of meeting Australia’s reporting commitments to the International Energy Agency. These estimates are compiled from an annual fuel and electricity survey conducted by the Bureau, supplemented by a variety of other sources of information. The survey provides a comprehensive and detailed ‘bottom-up’ quantification of energy use in Australia.

In future editions of Australia’s Energy Statistics the primary data source will be the NGERS data set which will replace the fuel and electricity survey. The coverage of the NGERS data set will be larger than the previous survey and is expected to improve the robustness of the national energy statistics, particularly in relation to the allocations between sectors. In addition, ABARES are undertaking work to ensure that sub threshold data is also included in the Australian Energy Statistics, through additional data collection using non-NGERS sources and statistical methods.

To ensure internal consistency and completeness, the data are reconciled with ‘top-down’ statistics on the supply and use of all major fuels in Australia collected from the suppliers of those fuels, i.e. the coal, oil, gas and electricity industries.

3.2.1 CO₂ Emissions and Emission factors

In general, the estimate of emissions of carbon dioxide used for each fuel, k , in each economic sector, h , is estimated by:

$$E_{hk} = (F_{hk} \cdot EF_{hk} \cdot P_k / 100) - S_{hk} \cdot 44/12 \dots \dots \dots (3.1)$$

Where: E_{hk} is the amount of CO₂ emitted from fuel k in economic sector h (in Gg);

F_{hk} = the amount of fuel k combusted in sector h (in PJ);

EF_{hk} = the CO₂ emission factor (in Gg CO₂/PJ) for fuel k ;

P_k = the oxidation factor (in percent) of fuel k ; and

S_{hk} = the amount of carbon sourced from fuel k which is stored in sector h (in Gg).

Emission factors for CO₂ depend only on the chemical composition of the fossil fuel concerned under IPCC methods. For fuels having well defined and/or stable chemical composition, CO₂ emission factors can be specified with considerable accuracy. This is particularly the case for natural gas and for petroleum products, with the exception of fuel oil, which may vary considerably in composition, and to a lesser degree for coals, which can vary in their composition of both combustible components (carbon, volatiles) and non-combustible components (ash, moisture).

Coal

Approximately 90% of all coal consumed in Australia is used by the electricity generation industry. Under NGERs all electricity generators who consume coal as their primary fuel must sample and analyse their coal and report their facility specific CO₂ emission factor. The reported emission factors are illustrated in Figure 3.8. After the electricity industry, the largest user of coal in Australia is the steel industry. The steel industry has provided a representative CO₂ emission factor of 91.8 Gg/PJ for black coal used in iron/steel/coke production (L.Leung, BHP 2001, pers. comm). This figure has been further verified by industry data obtained from NGERs as being representative. For other uses of coal, CO₂ emission factors are reported in Table 3.2.

Coke

The CO₂ emission factor for coke is derived from a carbon balance conducted on the coke oven subsector. Carbon input into coke ovens is estimated and balanced against carbon contained in the fuel and product outputs from coke ovens. The carbon content of coke is determined as the carbon content required to achieve a carbon balance for the overall coke oven process. The resulting coke emission factor varies slightly from year to year depending on the balance of inputs and outputs, in a range between 105.9 and 109.9 Gg/PJ which is comparable to the IPCC default factor (Table 3.A.19). The underlying data used to estimate the coke EF is confidential due to the sector being characterised by a limited number of producers.

Coal By-Products

Coal by-product fuels are defined as coke oven gas, coal tar and liquefied aromatic hydrocarbons. They are produced largely as a by-product of coke oven processes, however liquified aromatic hydrocarbons can also be produced from petroleum refining. The coke oven gas emission factor has been advised by the steel industry to be 37 Gg/PJ (Deslandes and Kingston 1997). The steel industry has also advised a representative emission factor for coal tar of 81.8 Gg/PJ. Liquefied aromatic hydrocarbons consist of compounds such as benzenes, toluene, xylene. Because of their similarities with naphtha and solvents, the same emission factor of 69.7 Gg/PJ was assigned to these products. Liquefied aromatic hydrocarbons and coal tars are generally used as feedstocks in the production of fossil fuel products rather than being combusted for energy. Therefore an assumption was made for the use of these fuels within the Petroleum and coal product manufacturing sector (ANZSIC Group 252), that 75% of the carbon contained in these fuels was stored in products and 25% combusted, which is consistent with IPCC guidelines.

Refined Petroleum Products

Australian oil tends to be of the light crude variety and the petroleum products generated by Australian refineries reflect the characteristics of these supplies. The country-specific emission factors for marketable petroleum products for this inventory are taken from GHD 2006a, which reports the results of a review of Australian petroleum products. Emission factors are listed in Table 3.2. A representative value to be used for fuel oil, derived from IPCC 1997, was confirmed by discussion with Nabalco (J. Bawdin, pers. comm.), one of the largest industrial users of fuel oil in Australia, and Shell (J. Le Cornu, pers. comm.).

Other Petroleum Products

In the Australian Energy Statistics sectors, Basic Chemicals (ANZSIC Group 253), Oil and Gas Mining (ANZSIC Subdivision 12) and Basic Non-Ferrous Metals (ANZSIC Group 272) (after excluding petroleum coke from the latter sector), petroleum products not elsewhere classified (nec) consists largely of naphtha. The emission factor for naphtha of 69.7 Gg CO₂/PJ, (IPCC 1997), was therefore used in these sectors. For all other Australian Energy Statistics sectors in which petroleum products nec appears as a fuel type, an emission factor of 69.7 Gg CO₂/PJ is used based on IPCC 1997 default for refinery feedstocks and other oil.

Petroleum refining consumes refinery gas/liquids and refinery coke in the process of converting to raw crude oil to refined products. Emission factors of 54.7 Gg CO₂/PJ (refinery gas and liquids) and 92.7 Gg CO₂/PJ (refinery coke) are used from the 2006 Guidelines (IPCC 2006), where appropriate factors have been provided or where there are no fuel specific factors in the previous IPCC guidelines (1997). NGERS data also made available facility-specific emission factors for several of the petroleum refineries. A decision to utilise these factors for the relevant refineries while maintaining the default factors for the remainder, was made in consultation with the decision tree in section 1.6.1. In doing so, it was recognised that refinery emission factors for these fuel types are strongly linked with the specific technology types and process configurations inherent in individual refineries.

Recycled tyres are combusted for energy within Cement, Lime, Plaster and Concrete (ANZSIC Group 263). An emission factor of 81.6 Gg CO₂/PJ was sourced from the US Energy Information Administration (GHD 2006b). The emission factor is conservative as it makes no allowance for any biomass component of tyres, as no data is available on natural versus synthetic rubber tyre composition.

Solvents, Lubricants, Greases and Bitumen

Australian information on CO₂ emission factors for these products is not available. The factor for solvents (69.7 Gg/PJ) is based on IPCC 1997. For the remaining fuels, emission factors from the 2006 Guidelines (IPCC 2006) are used where updated factors have been provided or where there are no fuel specific factors in the previous IPCC guidelines (1997) – 80.7 Gg/PJ for bitumen, and 69.6 Gg/PJ for lubricants and greases.

Natural Gas

A national emission factor has been estimated for natural gas using data on the composition of natural gas in each pipeline system, as published by the Australian Gas Association (various years), weighted by the volumes of gas consumed from each pipeline system (see Table 3.2).

The CO₂ emission factor for natural gas varies slightly between States, depending on the composition of the gas supplied to energy users in the State, which in turn depends on the characteristics of natural gas in the fields from which supply is sourced. In these circumstances, use of a single national weighted average emission factor for all natural gas will not introduce errors at the level of aggregate national *energy* sector emissions. All emission estimates for natural gas are therefore based on national consumption data and national emission factors, except for gas used for electricity generation. Under NGERS all electricity generators, that use gaseous fuels as their primary fuel, are required to sample and analyse their natural gas or coal seam methane and report their facility specific emission factor. The reported emission factors are illustrated in Figure 3.8. For small electricity generators who do not meet the reporting thresholds of NGERS the national CO₂ emission factor for natural gas is used.

An additional adjustment is made for natural gas activity data reported in the Australian Energy Statistics as used by the chemical industry because this includes both natural gas and the separate ethane supply that is used as feedstock. The CO₂ emission factor used for the inventory was derived based on data within the ASHRAE *Handbook Fundamentals* (2001) and is 56.5 Gg CO₂/PJ. Ethane is the main source of feedstock and fuel supply for the petrochemical industry in Victoria, which is the location for a large proportion for the total Australian petrochemical industry.

Town Gas

Town gas is a minor source of emissions and is given the same emission factor as LPG. It is assumed that in the manufacture of town gas, both carbon content and energy content is reduced in the same proportion, meaning that the carbon emission factor is unchanged.

Biomass Fuels

Emissions of CO₂ from biomass fuels are not included in the national inventory but are required to be reported as a Memo item. The CO₂ emission factors for bagasse and wood/woodwaste combusted in commercial and residential sectors are listed in Table 3.2. A detailed explanation of residential wood heater emission factors is provided in section 3.6. Factors for bagasse (95.0 Gg/PJ) and ethanol (67.3 Gg/PJ) are based on IPCC 2006 as fuel specific factors were not available in IPCC 1997.

Table 3.2: Emission factors for CO₂ 2009

Fuel Type	Fuel	CO ₂ emission factor (Gg CO ₂ /PJ)
Coal derived fuels	Coal used in public electricity generation ^(a)	82.7 – 96.3
	Coal used in steel industry ^(m)	91.8
	Black coal used by other industry ^(a)	90.0
	Brown coal used by industry ^(a)	94.6
	Coke ⁽ⁿ⁾	109.9
	Coal by-products (coke oven gas) ^(b)	37.0
	Coal by-products (coal tar) ^(b)	81.8
	Coal by-products (liquefied aromatic hydrocarbons) ^(l)	69.7
	Brown coal briquettes ^(a)	95.0
Petroleum fuels	LPG ^(c)	60.2
	Naphtha ^(l)	69.7
	Automotive gasoline ^(c)	67.4
	Aviation gasoline ^(c)	67.0
	Lighting Kerosene ^(c)	68.9
	Aviation turbine fuel ^(c)	69.6
	Power Kerosene ^(c)	68.9
	Heating oil ^(c)	69.5
	ADO ^(c)	69.9
	IDF ^(c)	69.9
	Petroleum products nec ^(l)	69.7
	Refinery gas and liquids ^(e)	54.7
	Refinery coke ^(e)	92.7
	Fuel oil ^(l)	73.6
	Tyres ^(k)	81.6
	Solvents ^(l)	69.7
	Lubricants and greases ^(e)	69.6
	Bitumen ^(e)	80.7

Fuel Type	Fuel	CO ₂ emission factor (Gg CO ₂ /PJ)
Gases	Natural gas (including coal seam gas) ^(f)	51.4
	Natural gas (Basic chemicals sector) ^(f)	51.4
	Ethane ^(g)	56.5
	Town gas ^(c)	60.2
Biomass fuels	Wood and wood waste ^(h)	94.0
	Wood (For Residential subsector) ⁽ⁱ⁾	77.5
	Ethanol ^(e)	67.3
	Bagasse ^(e)	95.0

Sources: (a) NGERS 2009. (b) Deslandes & Kingston 1997. (c) GHD 2006a. (d) ABARE (e) IPCC 2006. (f) Australian Gas Association 2001. (g) ASHRAE 2001. (h) Todd 1993. (i) Todd 2005. (k) GHD 2006b. (l) IPCC 1997. (m) L.Leung BHP 2001. (n) Derived from carbon balance within coke oven/iron & steel subsectors.

Note: All emission factors expressed in terms of energy measured as gross calorific equivalents (GCV).

Oxidation Factors for CO₂

The oxidation factor is defined as the proportion of carbon contained in a fuel which is oxidised to CO₂. Oxidation factors are either IPCC default values (IPCC 1997) or explained below and are reported in Table 3.3. One minus this factor is the proportion of carbon (on a mass basis) that is stored in solid products such as ash and soot.

Table 3.3: Oxidation factors for CO₂ (non-electricity)

Fuel	Sector	Oxidation Factor
Black Coal	All Categories	98.0% ^(a)
Brown Coal	All Categories	98.0% ^(a)
Coal By-products	Other than Basic Chemicals and Petroleum and Coal Products nec	99.5%
	Basic Chemicals and Petroleum and Coal Products nec	99.0%
Coke	All Categories	98.0% ^(a)
Petroleum	All Categories	99.0% ^(a)
Gas	All Categories	99.5% ^(a)
Biomass	Residential	100%
	All Other Categories	98.0%

Source: (a) IPCC (1997, Volume 3)

The oxidation factor for combustion of biomass for Residential is combined with the associated CO₂ emission factor, and is therefore reported as 100%. The oxidation applied to the combustion of biomass has been set at 98.0% for all sub-categories other than the Residential and Accommodation, Cultural and Personal (Division H, P & Q) categories. Less efficient combustion in these sectors has been assumed and hence a lower oxidation factor applied.

Not all carbon consumed is combusted. Some fossil fuels are used as feedstocks for the production of goods that act as a temporary store of carbon. Storage of carbon in Australia occurs principally in outputs produced in the chemical industry but also occurs in other products such as bitumen and steel.

Facility specific CO₂ emission factors reported under the NGERS incorporate the rate of oxidation based on analysis of ash contents. In such cases the application of the default oxidation factors would double count the effect of incomplete combustion, so an oxidation factor of 100% is applied.

3.2.2 Non-CO₂ Emissions

In addition to emissions of CO₂, the combustion of fuel in stationary sources results in the emission of CH₄, N₂O, NO_x, CO, and NMVOCs. Of these, methane and nitrous oxide account for around 1% of emissions on a carbon dioxide equivalent basis in this sector. The magnitude of these emissions is dependent on a large number of factors including fuel type, equipment design and emission control technology and is therefore inherently more complex and more uncertain than estimates of CO₂ emissions.

For non-CO₂ gases, emissions are estimated by:

$$E_{hkl} = F_{hk} \cdot E_{thkl} \quad \text{..... (3.2)}$$

where: E_{hkl} = amount greenhouse gas l emitted from combustion of fuel type k, in economic sector h (in Gg);
 F_{hk} = amount of fuel type k combusted in sector h (in PJ);
 E_{thkl} = technology weighted emission factor (in Gg/PJ) for greenhouse gas l, from fuel type k in sector h.

The characteristics of the capital stock are an important determinant of the non-CO₂ emissions generated by the combustion of fossil fuels. Consequently, emission factors for non-CO₂ are capital- and technology-specific and require capital specific information to be collected including equipment type, technology and in some cases, the age of capital.

The non-CO₂ factors are derived from IPCC 1997 and USEPA 1995 default values for uncontrolled emissions from various source categories, corrected for control technologies in use in Australia. In Australia, emissions from stationary fuel combustion sources are controlled to varying degrees. The emission factors for uncontrolled fuel combustion sources, which must be used in the absence of reliable information on the equipment in use, means that the Australian inventories may overestimate some non-CO₂ emissions.

The emission factors for non-CO₂ greenhouse gases for various equipment types are summarised in Table 3.A.4. In the absence of evidence to differentiate N₂O variations in measured N₂O concentrations between boilers (i.e. differences cannot be attributed either to differences in boiler type (e.g. tangentially-fired), boiler size, boiler load or combustion modifications (e.g. low NO_x burners)), it is assumed that the N₂O emission factors are dependent on fuel type only.

Data on the quantities of energy used by each type of equipment, disaggregated by ANZSIC Group and by major fuel type (coal, oil, gas and biomass) for the manufacturing industries, are provided by ABARES. These data are used to compile a set of weighted emission factors for each type of fuel combusted in each economic sector, reflecting the mix of equipment types in use in those sectors.

For the other economic sectors not covered by the above analysis fuel use by equipment type and emission factors for equipment types were estimated as follows:

- For ANZSIC class Division A (Agriculture, Forestry, Fishing), it was assumed that all diesel is used in mobile equipment. It is assumed that the small quantities of other fossil fuels consumed in Division A are used in the agricultural industry, in miscellaneous small combustion equipment.
- For Division E (Construction), mobile equipment emission factors are used.
- For Divisions F, H, I, J, K and L, it was assumed that all fuel combustion occurs in miscellaneous small combustion equipment.
- For sectors 65 and 67, it was assumed that consumption of gaseous fuels occurs in gas turbines (used to power compressors in gas transmission and distribution systems) and all consumption of liquid fuels occurs in mobile equipment.
- For the Residential sector, it was assumed that all combustion of coal, liquid fuels and gas occurs in miscellaneous residential combustion equipment.

In ANZSIC sector 361, public electricity generation, data is available on the relevant equipment data for each power station.

3.2.3 SO₂ Emissions

Data on emission factors was obtained from the following sources:

- for petroleum products Australian Institute of Petroleum and the National Pollutant Inventory (DSEWPC)
- for natural gas and LPG Australian Gas Association
- for coal (default values) the former Australian Government Department of Primary Industries and Energy.

For the electricity industry specific SO₂ emission data have previously been obtained from power station operators. If historical data is not available defaults are used as listed in Table 3.4. For other sectors, the emission factors are derived from data from the Australian Institute for Petroleum, the Australian Gas Association and the former Australian Government Department of Primary Industries and Energy.

Table 3.4: SO₂ emission factors

Fuel	SO ₂ emission factors (Gg SO ₂ /PJ)
Black coal	0.37
Brown coal	0.15
LPG	0.002
Aviation gasoline	0.008
Kerosene	0.057
Heating oil	0.057
ADO	0.057
IDF	0.057
Fuel oil	1.282
Natural gas	0.002

Source: Australian Institute for Petroleum, National Pollutant Inventory (petroleum refining), Department of Primary Industries and Energy (pers. comm. 1998) (for default coal values) and Australian Gas Association.

For both CO₂ and non-CO₂ gases, total national emissions are calculated by summing the estimated emissions from each fuel in each sector across all fuels and across all sectors.

3.2.4 Activity Data

The Australian Bureau of Agricultural and Resource Economics and Sciences (ABARES), and predecessor organisations, have compiled statistics of Australian energy use by economic sector and fuel since the 1970s. These statistics are compiled from an annual fuel and electricity survey conducted by the Bureau supplemented by a variety of other sources of information. The statistics provide a comprehensive and detailed ‘bottom-up’ quantification of energy use in Australia. They are reconciled with ‘top-down’ statistics of all major fuels in Australia, collected from the suppliers of those fuels, i.e. the coal, oil, gas and electricity industries.

In forthcoming editions of Australia’s Energy Statistics ABARES will use data collected under NGERS as the primary source of energy consumption data. NGERS reporting is compulsory for facilities over specified energy and emissions thresholds. This is expected to provide more coverage than was previously available from the voluntary fuel and electricity survey. The use of NGERS data is expected to improve the quality of Australia’s Energy Statistics. NGERS data provides energy consumption disaggregated by fuel type, fuel use and ANZSIC sector required by ABARES. In addition, ABARES are undertaking work to ensure that sub threshold data is also included in the Australian Energy Statistics, through additional data collection using non-NGERS sources and statistical methods.

The data are presented in common energy units (PJ) on an individual State basis. ABARES also collects statistics of energy use by equipment (technology) type. These have been used to compile the technology weighted sectoral emission factors for non-CO₂ greenhouse gases. Several re-allocations to the ABARES statistics are required in order to:

- break down energy consumption into sub-sectors where this is required to match emission factor data, but is not done by ABARES,
- identify and allow for stored carbon,
- separate coke production from other parts of the iron and steel industry, also as specified in the IPCC guidelines,
- eliminate double counting of gas leakage from the gas distribution system;
- allocation of fuel use to the industrial process sector for the estimation of emissions from the use of fuels as reductants.

ABARES undertakes a reconciliation at the level of the supply and use of energy in the economy at the level of energy units. The ABARES analysis ensures that all energy entering the economy is accounted for by end-uses.

Activity data for the time-series 1990 to 2009, reported by category level and fuel type, are available on the AGEIS website: <http://ageis.climatechange.gov.au/QueryAppendixTable.aspx>

3.2.5 Feedstock and Non-Energy Fuel Use

Activity data and emissions associated with the non-energy use of fuels are not reported within the *Fuel combustion* subsector. In accordance with the IPCC Guidelines, they are reported under the *industrial processes sector* and *fugitive emissions from fuels* sub-sector as follows.

- Reported in *industrial processes*
 - Coke and natural gas where used as a reductant in the integrated coke/iron and steel production – reported in 2.C.1 Iron and Steel Production;
 - Black coal where used as a reductant in synthetic rutile production – reported in 2.B.5 Chemical Industry – Other;
 - Black coal, coke, petroleum coke and fuel oil where used as a reductant in base metal production – reported in 2.C.2 and 2.C.5 Ferroalloys and Other;
 - Petroleum coke where used as a reductant in titanium dioxide production – reported in 2.B.5 Chemical Industry – Other;
 - Petroleum coke, coal tar and coke used for anodes in aluminium production – reported in 2.C.3 Aluminium Production;
 - Natural gas used in Ammonia production – reported in 2.G Other;
 - Coke where used as a reductant in soda ash production– reported with other emissions from soda ash production in 2.A.3 Limestone and Dolomite Use.
- Reported in *fugitive emissions from fuels*
 - Oil refinery flaring- reported in 1.B.2.a. Oil Refining/Storage
 - Natural gas leakage – reported in 1.B.2.b Natural Gas Distribution.

There were 12.0 Mt CO₂ emissions from non-energy fuel use reported in the *industrial processes* sector of the inventory. Coke used as a reductant within iron and steel production was the most significant non-energy use of fuel, resulting in the reporting of 6.2 Mt CO₂ within the *industrial processes* sector. The production of aluminium and other metals accounted for another 2.4 Mt CO₂ emissions from non-energy fuel use of reported in the *industrial processes* sector.

The use of ethane and petroleum products used as feedstocks within chemicals production resulted in the equivalent of 2.0 Mt CO₂ being stored in products such as plastics, and therefore were not included in the estimate of emissions from the *energy* sector. A further 4.6 Mt CO₂ was associated with non-energy fuel use stored in other products. This includes lubricants, bitumen, solvents and coal oils and tars from coking ovens.

3.2.6 QA/QC

The Carbon Balance

A carbon balance for all years was undertaken in terms of the supply and use of carbon from fuels in the economy. All carbon entering the economy is accounted for—either as emissions from fuel combustion, emissions from the use of fossil fuels as reductants, non-energy uses, use of biomass sources of energy and international bunkers. While the predominant outcome of carbon entering the economy is emissions, a small portion of the total is stored in carbon-containing products or non-oxidised as ash.

Tables detailing the results of the carbon balance can be found in Annex 6.

Comparison with international data

Implied emission factors for all major fuels are tested for differences against the mean of the population of all other Annex I data. For each major fuel, the t-tests conducted show that the implied CO₂ emission factors for Australian fuels are not significantly different to the mean of the implied emission factors for the Annex I population with one exception. For non-CO₂ fuels implied emission factors are not significantly different to the mean of the Annex I population with the exception of liquid fuels for transport purposes which is discussed in more detail in section 3.5.4. The Australian Bureau of Agricultural and Resource Economics and Sciences (ABARES) is the common source of energy data both for the preparation of the national inventory through its publication, Australian Energy Statistics, and for Australia's report to the International Energy Agency (IEA). Some differences occur from year to year between the activity data in the inventory CRF tables and those published by the IEA. A project has been undertaken to reconcile the data provided by ABARES to the IEA with the published ABARE data used in the inventory. Following discussions with ABARES and an examination of the 2007 IEA submission data, an improved understanding of the underlying causes driving differences in the IEA data has been gained. The principal reasons for differences include a) differences in data year alignment – for some fuels published data relates to Australian financial years (July to June) while IEA data relates to calendar years and b) differences in the IEA reporting structure compared with that of the Australian Energy Statistics which has led to differences in average energy conversion factors used for major fuels.

ABARES have initiated a project to realign the format of the AES with that used by the IEA. One of the anticipated benefits of this project is a reduction in the observed differences between the AES and the IEA.

3.3 SOURCE CATEGORY 1.A.1 – ENERGY INDUSTRIES

3.3.1 Source Category Description

This category includes emissions from fuel combustion within electricity generation, petroleum refining and other energy manufacturing industries such as coke ovens, briquette production, coal mining, oil and gas extraction, and natural gas production, transmission and distribution. ABARES report energy consumption for economic sectors defined using the Australia New Zealand Standard Industrial Classification (ANZSIC) developed by Australia's national statistical agency, the Australian Bureau of Statistics. The mapping of data to IPCC classifications from the ANZSIC codes is complete and reported in Table 3.5.

Table 3.5: Relationship between IPCC source categories and ANZSIC sectors: Energy Industries

IPCC Source Category	ANZSIC Subdivision/Group/Class			
	Division	Sub-division	Group/Class	Description
1.A.1 Energy Industries				
a Electricity and heat production	D Electricity, Gas and Water Supply	36	361	Electricity supply
b Petroleum refining	C Manufacturing	25	251	Petroleum refining
c Solid fuel transformation and other energy industries	B Mining	11		Coal mining (incl. briquette production)
	B Mining	12		Oil and gas extraction (incl. gas processing and LNG production)
	C Manufacturing	27	2711 (part)	Coke ovens associated with Basic iron and steel manufacturing
	D Electricity, Gas and Water Supply	36	362	Gas supply
	I Transport and Storage	65	6501 (part)	Pipeline transport of gas

Note: (a) This comprises only electricity generated at single purpose power stations and at co-generation plants that are predominantly orientated to producing and exporting electricity. Accounting for all fossil fuel combustion associated with other co-generation occurs in the industry sector within which co-generation occurs. There is no public generation of distributed heat in Australia.

3.3.2 Methodology

In summary, emissions for the *energy industries* category are estimated using Tier 2 approaches and country specific factors (Table 3.6).

Table 3.6: Summary of methods and emission factors: Energy Industries

Categories	CO ₂		CH ₄		N ₂ O	
	Method applied	Emission factor	Method applied	Emission factor	Method applied	Emission factor
1A1a Public electricity	T2	PS, CS	T2	CS	T2	CS
1A1b Petroleum refining	T2	CS, PS	T2	CS	T2	CS
1A1c Manufacture of Solid Fuels	T2	CS	T2	CS	T2	CS

Notes: T1 = Tier 1. T2 = Tier 2. T3 = Tier 3. CS= Country-specific. D= IPCC default, PS = Plant Specific.

Electricity Generation (ANZSIC Group 361) (1.A.1.a)

Electricity generation includes power for supply to the grid (whether the power stations are owned by public or private corporations). Where it is not possible to separately identify the quantity of fuel consumed for electricity generation – for example, in electricity generation at industrial sites, the fuel consumed and the associated emissions are allocated to the industry involved and hence not included in the emissions for this sub-category. Public heat production does not occur in Australia.

Choice of emission factors

A tier 2 approach is used for the key category of electricity generation in which emission factors for fuels such as coal vary from source to source and over time. The fundamental reporting unit in this sector is the individual power station. Data is collected from power stations through the NGERs. Under NGERs, facilities over certain thresholds are required to submit data on fuel consumption, fuel energy content, fuel emission factors (incorporating oxidation factors), emission estimates and the amount of electricity generated and sent out to the Greenhouse and Energy Data Officer. Power stations must sample and analyse their primary solid and gaseous fuels in accordance with the requirements and standards listed in the *National Greenhouse and Energy Reporting (Measurement) Determination*. The adoption of these methods and standards ensures accuracy and comparability in the facility specific information reported. This data provides facility specific energy content and emission factors for the solid and gaseous fuels consumed in each power station.

When the NGERs was established the methods to be used by power stations were aligned with those that applied under the Generator Efficiency Standards program – as detailed in the *Generator Efficiency Standards Technical Guidelines*. The Generator Efficiency Standards program had been in place in Australia since 2000 and data collected under this program has been utilised in the national inventory throughout the time series. The adoption of consistent methods in the NGERs and the Generator Efficiency Standards program ensured time series consistency in the emission estimates in the national inventory.

Country-specific emission factors are utilised for minor (mainly liquid) fuels.

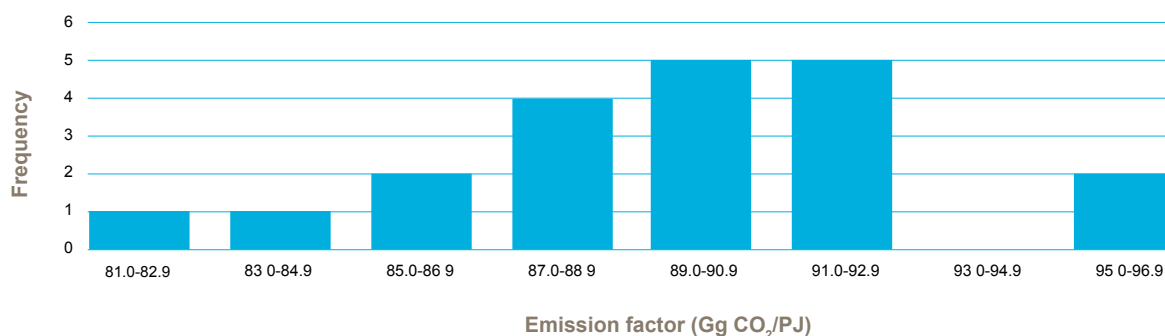
Activity data

Under NGERs data is received from all large and medium sized power stations in Australia. There are around 100 such fossil fuel based power stations in Australia at present (the number depends on how generating units at the same site are treated). NGERs data has resulted in a significant increase in the number of power stations where facility level data is available (increasing from 50-60 to around 100). The energy use of the small power stations, that do not meet the NGERs reporting thresholds, are estimated as the difference between the total of reported values under NGERs and Australian Bureau of Agricultural and Resource Economics and Sciences (ABARES) energy statistics for ANZSIC class 361. This approach has been adopted throughout the time series. Therefore the improved coverage of power stations under NGERs in 2009 does not alter the method for estimating total fuel consumption in this sector. In 2009 the coverage of individual coal power station data was comprehensive and displaced the necessity of using the coal data in ABARES' energy statistics for ANZSIC class 361 in calculations. However, the calculation method of the difference between the total of reported values and ABARES energy statistics for ANZSIC class 361 was used for other liquid and gaseous fuel types.

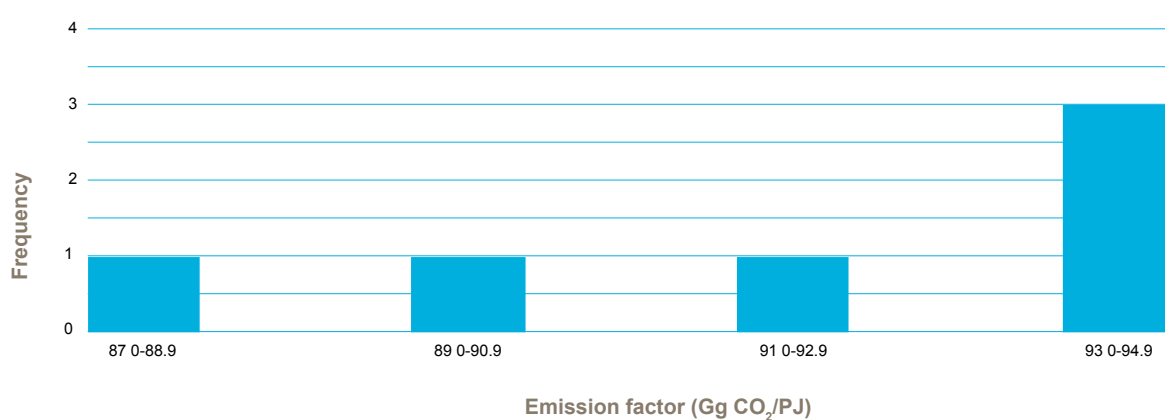
Under the NGERs oxidation factors and the emissions factors are linked in that power station operators report CO₂ emission factors including the effects of oxidation based on analysis of ash contents and in accordance with *NGERS Measurement Determination*. In such cases applying the default oxidation factor of 99% for electricity generation coal would double-count the effect of incomplete combustion, so an oxidation factor of 100% is used. Throughout the time-series CO₂ factors for many facilities do not include the effects of incomplete oxidation. For these facilities oxidation factors are applied as reported by the power station. If the power station has not reported a facility specific oxidation factor then the national default factor is applied.

Figure 3.8: Emission factors for CO₂ in electricity generation, 2009

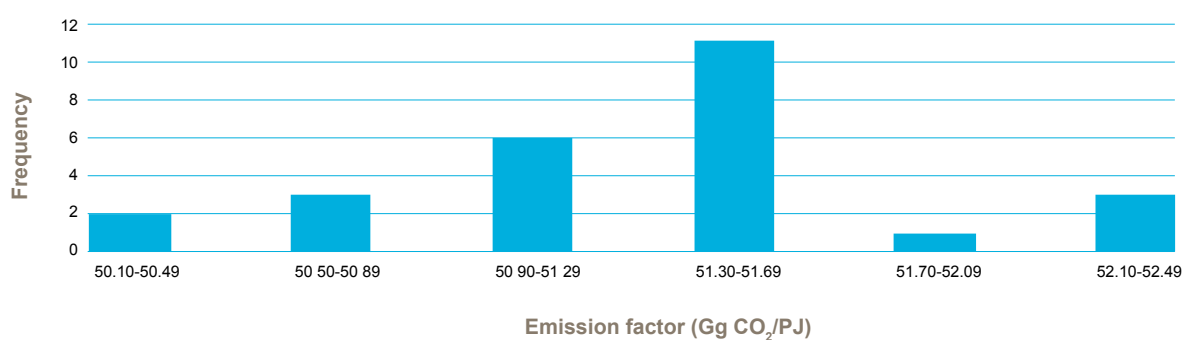
a) Black coal power station



b) Brown coal power station



c) Natural gas and waste power station



Source: NGRS 2009

Notes: Values incorporate the effect of partial oxidation of fuels.

Petroleum Refining (ANZSIC Group 251) (1.A.1.b)

The main fuels used by petroleum refineries are refinery gas/liquids and natural gas along with some minor use of other liquids fuels. The Australian Energy Statistics (ABARE 2010) reports refinery feedstock, i.e. essentially crude oil, as the major input, together with other, undefined, petroleum products. The various market petroleum products are shown as energy outputs. The total energy content of the products produced by the sector is less than the energy content of the petroleum input, with the difference being energy consumed by the refining processes (distillation, cracking etc.). The fuel from which this energy is derived is obtained from the crude oil input and is referred to as refinery fuel.

Choice of emission factor

NGERS data made available facility-specific emission factors for the fuels; refinery gas and liquids, refinery coke and natural gas from several of the petroleum refineries. A decision to utilise these factors for the relevant refineries while maintaining the default factors for the remainder, was made in consultation with the decision tree in section 1.4.1. In doing so, it was recognised that refinery emission factors for these fuel types are strongly linked with the specific technology types and process configurations inherent in individual refineries.

Activity data

The refinery fuel balance contained in ABARE 2010 is analysed using a model that examines the expected refinery plant efficiency in the conversion of crude oil to final products, taking into account factors such as the change to low sulphur diesel. For years where ABARE 2010 has difficulty providing a reasonable balance (2000-2008) the modelled data has been used for refinery fuel consumption.

Detailed fuel consumption data was made available via NGERS for all Australian oil refineries for 2009. The efficiency of refinery fuel consumption is influenced by technology type and process configurations, which change over time. Therefore it was considered appropriate that the NGERS data be used for 2009 (being the best data available), however this data would not necessarily be reflective of refinery conditions for previous years. Therefore, the 1990 to 2008 times series using ABARES and modelled data was maintained rather than attempt a recalculation using proxy data (ie refinery efficiency) obtained from the 2009 NGERS data. Time series analysis of fuel use and emissions demonstrate that time series consistency has been maintained in the approach taken in this NIR.

The NGERS has provided data for the first time on the emissions associated with the burning of refinery coke to restore the activity of the catalyst during the refining process. Given that this component of petroleum refining emissions has previously been included within total refinery fuel combustion, it was decided to continue with this practice for this submission in order to maintain time series consistency. This remains consistent with practice followed by most other countries and the guidelines are unclear as to where emissions from this source should be reported. However, for transparency purposes, these emissions from refinery coke have also been noted in the Fugitives – petroleum refining section of the NIR. Refinery flaring is accounted for in the Fugitive Fuel Emissions sector.

Manufacture of Solid Fuels and Other Energy Industries (1.A.1.c)

The manufacturing of solid fuels and other energy industries sector, 1.A.1c, comprises six ANZSIC sectors:

- Coke Oven Operation (ANZSIC Class 2711, part),
- Briquette Manufacture (ANZSIC Class 1102, part),
- Coal Mining (ANZSIC Subdivision 11, remainder),
- Oil and Gas Extraction (ANZSIC Subdivision 12),
- Other Transport, assumed to be gas pipeline transport (ANZSIC Subdivision 65),
- Gas Supply (ANZSIC Group 362).

Estimated emissions are derived from equations 3.1 and 3.2 and the emission factors reported in Tables 3.2 and 3.3 and Table 3.A.1.

The *Coke Oven Operation (ANZSIC Class 2711, part)* sub-sector is effectively a subsidiary activity of the Iron & Steel industry but is classified by the IPCC as an energy transformation industry and hence is reported separately. This sub-sector is both a consumer of black coal and coal by-products and a producer of coke and coal by-products. Consequently, fuel combustion is calculated by deducting derived fuels produced by the sector from energy inputs.

The consumption of natural gas in the *Other Transport (Natural Gas Transmission) (ANZSIC Subdivision 65)* as listed by ABARES under ANZSIC Subdivisions 65–67 is assumed to be used in gas turbines to drive pipeline compressors, and is thus considered an energy transformation activity. Emissions from the small amount of other energy used in these Subdivisions are accounted under Commercial/Institutional.

The Gas Production and Distribution (ANZSIC Group 362) sector is also one of the energy transformation industries, manufacturing town gas from both natural gas and LPG. Fuel consumption, as indicated by the *ABARE Energy Supply and Demand Statistics* consists of:

- natural gas and LPG used to make town gas,
- other gas, including both natural gas and town gas, used by the industry for its own purposes,
- gas leakage.

The quantity of town gas produced is shown as an energy output of the sector in the *ABARE Energy Supply and Demand Statistics*. It was assumed that all LPG is converted to town gas, and none is combusted in the conversion process. LPG consumption was therefore offset in full against an equal quantity (in terms of energy content) of town gas produced. The remaining town gas production was subtracted from total natural gas consumption. Gas leakage, reported under 1.B *fugitive fuel emissions*, was also subtracted from the remaining net natural gas consumption. The figure remaining was assumed to be the quantity of natural gas combusted in the sector. In each year, this figure is roughly a quarter of total fuel consumption in the sector as shown in ABARES statistics.

3.3.3 Uncertainties and Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas. Figure 3.8 provides an illustration of the level of variability in reported carbon dioxide emission factors for coal burned in electricity generation.

Time series variability of GHG implied emission factors are likely to be influenced by changes in fuel mix within categories, and changes of actual fuel emission factors. Notable examples of where such variations occur in 1.A.1 *energy industries* are set out below:

- 1.A.1.c *manufacture of solid fuels and other energy industries* – CO₂ from solid fuels: The implied emission factor (IEF) declines by 10% between 1990 and 2001. This can be explained by the relative rise of coal by-products—coke oven gas as a fuel (with a relatively low emission factor of 37 Gg/PJ) at the expense of black coal.
- 1.A.1.a *public electricity* – CO₂ from biomass: Biomass combustion for electricity consists of a growing proportion of biogas from landfill. Biogas has a relatively low CO₂ emissions factor compared to other biomass fuel, hence Australia's CO₂ biomass IEF is relatively low.

No major deviations are evident in emission and activity data trends.

3.3.4 Source Specific QA/QC

This source category is covered by the general QA/QC measures of the greenhouse gas inventory discussed in section 1.6. Results for the reference approach for the *energy* sector, reported in Annex 4, and the carbon reconciliation reported in Annex 6, provide quality control checks for this sector.

Fuel and generation data for 1.A.1.a *public electricity* are compiled by the Department of Climate Change and Energy Efficiency from NGERS reported data and from ABARES energy data. Inputs are reconciled and emission data is fully reconciled against the outputs from the AGEIS to ensure the accurate reporting in this sector.

Fuel and generation data are also checked and reconciled against the alternative data sources of the Energy Supply Association of Australia (ESAA) and the Australian Energy Market Operator (AEMO). These comparisons confirm the consistency of the estimates to a high level of accuracy and show that all energy/carbon has been accounted for.

3.3.5 Recalculations Since the 2008 Inventory

Recalculations to 1.A.1 *energy industries* are detailed at the sub-category level in Table 3.7.

Key reasons for recalculations in Energy industries include;

- Revisions in ABARES Australian Energy Statistics updating the estimates of fuel consumption in certain industries due to the availability of improved data.
- An improved understanding of detailed fuel type use from NGERS data has led to a refinement in the understanding of aggregated fuel types reported by ABARES in the AES.

Table 3.7: 1.A.1 Energy Industries: recalculation of total CO₂-e emissions, 1990-2008

	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1.A.1.a Electricity and heat production (a)				
1990	129,456	129,456	–	0.00
2000	175,323	175,323	–	0.00
2001	182,511	182,511	–	0.00
2002	183,741	183,741	–	0.00
2003	191,964	191,964	–	0.00
2004	193,511	193,511	–	0.00
2005	194,147	194,147	–	0.00
2006	197,949	197,949	–	0.00
2007	199,959	199,959	–	0.00
2008	204,295	203,061	-1,166	-0.57
1.A.1.b Petroleum refining (b)				
1990	5,753.1	5,835.6	82.5	1.43
2000	6,406.2	6,493.1	86.9	1.36
2001	6,513.3	6,598.9	85.6	1.31
2002	6,433.8	6,517.5	83.7	1.30
2003	6,333.7	6,413.8	80.2	1.27
2004	5,909.6	5,980.9	71.3	1.21
2005	5,888.3	5,958.8	70.5	1.20
2006	5,297.3	5,358.1	60.8	1.15
2007	5,822.5	5,885.6	63.1	1.08
2008	5,627.0	5,684.2	57.2	1.02
1.A.1.c Manufacturing of solid fuels and other energy industries (c)				
1990	7,888.3	7,931.1	42.8	0.54
2000	10,880.9	10,881.3	0.3	0.00
2001	10,859.1	10,859.3	0.2	0.00
2002	11,124.5	11,124.7	0.2	0.00
2003	13,224.9	13,226.0	1.1	0.01
2004	13,635.8	13,314.2	-321.6	-2.36
2005	14,530.9	14,176.5	-354.4	-2.44
2006	15,093.1	14,621.4	-471.6	-3.12
2007	16,019.9	15,622.3	-397.6	-2.48
2008	16,436.3	16,167.1	-269.2	-1.64

Recalculation explanation

(a) ABARE Australian Energy Statistics

Updates to estimates of natural gas and liquid fuel consumption in 2008.

(b) Improved understanding of fuel type use from NGERS data

Change in EF for petroleum products nec

Previously a aggregated fuel type (BTX) was used along with a representative emission factor. However NGERS data has provided the actual fuel type use which has allowed reallocation of fuel use reported by ABARE as being liquefied aromatic hydrocarbons and coal tar. Therefore, the respective EFs have been applied to the individual fuel types.

The EF for ppnec use as reported by ABARE in the AES was changed from 68.6 Gg CO₂/PJ to that of the IPCC 1997 default for refinery feedstocks and other oil (69.7 Gg CO₂/PJ) as it was considered more appropriate for the fuel types consumed.

(c) ABARE Australian Energy Statistics

Improved understanding of fuel type use from NGERS data

Reallocation of diesel use in mining to the Transport sector for years 2004 to 2008. Natural gas use increased in 2006 and decreased in 2007 and 2008.

Coal use in Coke production has been changed from applying a black coal EF of 90 Gg CO₂/PJ to a appropriate EF for coking coal of 91.8 Gg CO₂/PJ

3.3.6 Planned Improvements

Uncertainty data reported by corporations under the National Greenhouse and Energy Reporting system will be analysed to see how they can be incorporated into the national inventory estimates of uncertainty.

Some of the country specific emission factors for liquid fuels were estimated more than ten years ago. A study is planned to assess these emission factors and to assess whether any updates to the emission factors are required.

Currently it is not always possible to identify the quantity of fuel consumed for the purpose of electricity generation – for example, in electricity generation at industrial sites, the fuel consumed and the associated emissions are allocated to the industry involved. A project is planned to analyse NGERS data to inform an estimate of the amount of energy, and resulting emissions, that result from the generation of electricity at these facilities.

3.4 SOURCE CATEGORY 1.A.2 – MANUFACTURING INDUSTRIES AND CONSTRUCTION

3.4.1 Source Category Description

This source category includes emissions from fuel combustion in manufacturing, construction and non-energy mining. This includes both stationary and mobile equipment such as earth moving and mining equipment.

ABARES report energy consumption for economic sectors defined using the Australia New Zealand Standard Industrial Classification (ANZSIC). The mapping of ANZSIC codes against IPCC classifications is complete and given in Table 3.8.

Table 3.8: Relationship between IPCC source categories and ANZSIC sectors: Manufacturing and Construction

IPCC Source Category	ANZSIC Subdivision/Group/Class			
	Division	Sub-division	Group/Class	Description
2. Manufacturing Industries and Construction				
A Iron and Steel		27	271	Iron and steel manufacturing (excl. Coke ovens)
B Non-Ferrous Metals		27	272	Basic non-ferrous metal manufacturing
C Chemicals		25	252	Petroleum and coal product manufacturing
		25	253	Basic chemical manufacturing
		25	254	Other chemical product manufacturing
		25	255	Rubber product manufacturing
		25	256	Plastic product manufacturing
D Pulp, Paper and Print		23		Wood and paper product manufacturing.
		24		Printing, publishing and recorded media
E Food Processing, Beverages and Tobacco	C Manufacturing	21		Food, beverages, tobacco manufacturing
F Other (part)	B Mining	13, 14, 15		Metal ore mining, Other mining, Services to mining
F Other (part) (all other manuf.)		22		Textiles, clothing, footwear and leather manufacturing

IPCC Source Category	ANZSIC Subdivision/Group/Class			
	Division	Sub-division	Group/Class	Description
F Other (part) (non-metallic mineral products)		26	261	Glass and glass product manufacturing
F Other (part)		26	262	Ceramic product manufacturing
		26	263	Cement, lime, plaster and concrete product manufacturing
		26	264	Non-metallic mineral product manufacturing n.e.c.
F Other (part) (all other manuf.)		27	273, 274, 275, 276	All other metal product manufacturing
F Other (part)		28		Machinery and equipment manufacturing
		29		Other manufacturing
F Construction	E Construction	41, 42		Construction

3.4.2 Methodology

The emissions for *manufacturing industries and construction* are estimated using tier 2 approaches. Emissions estimated from activity data are based on the national survey of energy consumption by industry sector and fuel type compiled by the ABARES. CO₂ emission factors are country-specific and direct industry advice on the use of CO₂ emissions factors has been adopted for the use of coal by-products within 1.A.2.c *chemicals*, black coal within 1.A.2.a *iron and steel*, and natural gas in general. Non-CO₂ emission factors have been calculated using a sectoral equipment-weighted average approach and are reported in Table 3.A.2. More detail is provided for the metal and chemicals industries.

Table 3.9: Summary of methods and emission factors: Manufacturing and Construction

Category	CO ₂		CH ₄		N ₂ O	
	Method applied	Emission factor	Method applied	Emission factor	Method applied	Emission factor
1A2a Iron and Steel	T2	CS	T2	CS	T2	CS
1A2b Non-Ferrous Metals	T2	CS	T2	CS	T2	CS
1A2c Chemicals	T2	CS	T2	CS	T2	CS
1A2d Pulp, Paper and Print	T2	CS	T2	CS	T2	CS
1A2e Food Processing, Beverages and Tobacco	T2	CS	T2	CS	T2	CS
1A2f Other	T2	CS	T2	CS	T2	CS

Notes: T1 = Tier 1, T2 = Tier 2, T3 = Tier 3, CS= Country-specific, D= IPCC default.

Iron and Steel (ANZSIC Group 271) (1.A.2.a)

The methodology in the *iron and steel* sub-sector is somewhat more complex than many other sections of the inventory. This complexity arises from a number of factors:

- The operation of Coke Ovens is considered to be an energy transformation industry, and hence must be reported separately to the rest of the iron and steel emissions.
- The production of coke yields a variety of by-products, including coke oven gas, coal tar, liquefied aromatic hydrocarbons and naphthalene, each having quite different calorific values and emission factors. Coke oven gas is used as fuel in coke ovens and adjacent steelworks, while the other products are in general not combusted, but are used as feedstock in the chemical industry.
- Overall, the Coke Ovens sector is a producer of coke, most of which is consumed in the Iron and Steel sector and some of which is exported to other sectors (and other countries).

- The operation of blast furnaces to produce pig iron also produces yet another coal by-product, blast furnace gas, which is a low calorific value fuel consisting mainly of CO (and atmospheric nitrogen), used elsewhere in the steelworks. For the purpose of calculating CO₂ emissions, the production and subsequent combustion of blast furnace gas is ignored, and it is assumed that all coal and coke used in the iron and steel industry undergoes complete oxidation to CO₂, apart from the small adjustments resulting from application of oxidation factors and a further allowance for carbon sequestered in steel.
- The use of coke, as well as natural gas in hot briquetted iron production is regarded primarily as a chemical process rather than fuel combustion under IPCC reporting guidelines. Consumption and emissions are therefore reported under the *industrial processes* sector 2.C.3 rather than the *energy* sector.
- Post 2001, steel producers have fed pulverised black coal direct into blast furnaces. This is currently counted as a fuel input within the energy sector. It is anticipated that the collection of NGERS data may assist in the possible reallocation of pulverised black coal where used as a reductant to the Industrial Processes sector at some time in the future.

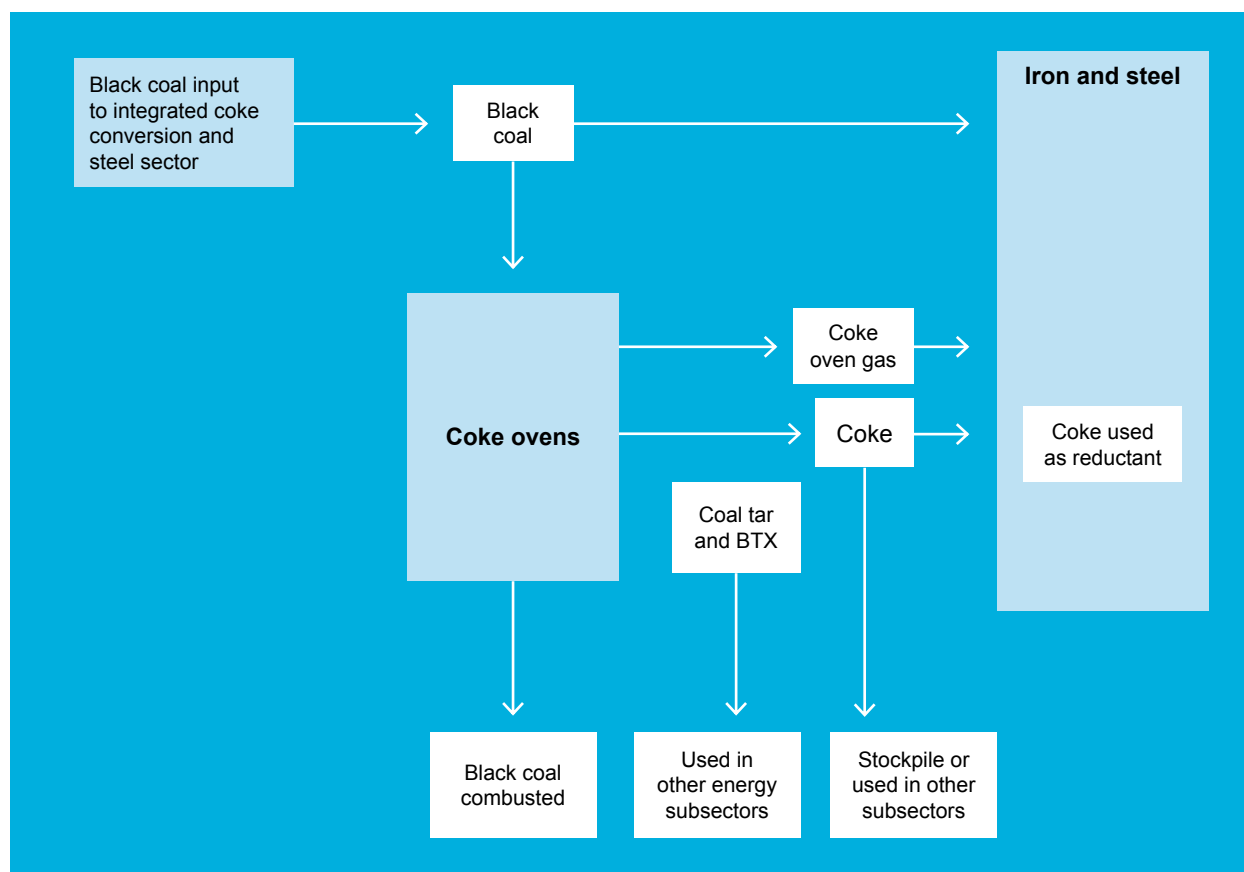
Although Coke Ovens (ANZSIC code 2711) are in operation in the iron and steel industry, they are considered an energy transformation industry under the IPCC methodology. Therefore Coke Ovens must be separated from the other parts of the iron and steel industry, so that it can be reported under IPCC category 1.A.1.c.

The statistics show that production of both coke and coal by-products exceed consumption within the sectors, i.e. the iron and steel industry as a whole is a net producer of coke and coal by-products. Only the estimate of consumption is used to estimate emissions from the Iron and Steel sector. Some of the remaining production may appear elsewhere in the national inventory if it is consumed as fuel by other industries in Australia, in which case the emissions are allocated to the consuming industry.

Production consumed elsewhere includes some coke (though in most years the majority of surplus coke produced by the industry is exported from Australia), and surplus coal by-products, most of which are consumed by the Coal and Petroleum Products sector.

A schematic chart showing energy flows within the integrated coke oven/Iron and Steel subsectors is shown in Figure 3.9. All coke used for reduction is reported in the *industrial processes* sector. Some electricity is produced as a by-product of the iron and steel production process. Given the primary role of the fossil fuel inputs in the iron and steel production process, those emissions are not re-allocated to electricity production.

Figure 3.9: Coke oven and Iron and Steel energy flow chart



Non-Ferrous Metals (ANZSIC Group 272) (1.A.2.b)

The consumption of petroleum products nec (meaning other, unspecified petroleum products ‘not elsewhere counted’) in this sector includes petroleum coke and coal tar used to make carbon anodes for aluminium production. CO₂ emitted from oxidation of carbon anodes in aluminium smelters is accounted in UNFCCC category 2.C.3. The quantity of petroleum coke and coal tar consumed in this sector, as advised by industry each year, is therefore subtracted from energy consumption of petroleum products nec and coal by-products, in order to eliminate double counting. It is assumed that the remaining energy consumption of Petroleum Products nec consists of naphtha. Some use of black coal in the production of synthetic rutile as well as black coal, coke, petroleum coke and fuel oil for base metal smelting occurs for reductant purposes. Therefore these fuel quantities are also deducted from the *energy* sector fuel consumption and reported under the *industrial processes* sector.

Chemicals (ANZSIC Subdivision 25) (1.A.2.c)

This sub-sector spans five ANZSIC classes:

- Petroleum and coal product manufacturing (ANZSIC Group 252),
- Basic chemical manufacturing (ANZSIC Group 253),
- Other chemical product manufacturing (ANZSIC Group 254),
- Rubber product manufacturing (ANZSIC Group 255),
- Plastic product manufacturing (ANZSIC Group 256).

The Chemicals sector (ANZSIC Subdivision 25) is a major energy user. Most of the energy is used by the Petroleum Refining and Basic Chemical Manufacturing sub-categories. Energy use in these two sub-categories is separately reported at the national level.

Non-energy use of natural gas in the production of ammonia is regarded as an industrial process and is therefore reported under the *industrial processes* sector rather than the *energy* sector, in order to prevent double counting. Likewise, the non-energy use of petroleum coke for titanium dioxide production and coke oven coke used in soda ash production are also reported within the *industrial processes* sector.

The calculation of emissions in the Chemicals sector must identify and allow for carbon stored in products. Sequestration takes place in the Petroleum and Coal Product Manufacturing (ANZSIC Group 252) and Chemical Manufacturing (ANZSIC Group 253) sub-categories, where fossil fuels are used as feedstock. Data is also obtained directly from chemical companies in order to estimate the quantity of carbon sequestered in products from feedstocks, with emissions estimates adjusted accordingly.

Coal by-products constitute the largest fuel input into the *Petroleum and Coal Product Manufacturing* (ANZSIC Group 252) sector. It is assumed that these consist of coal tar and liquefied aromatic hydrocarbons and that, in the absence of specific information about this industry sector in Australia, 75% of this fuel is sequestered in long lived coal products, following the default assumption of the IPCC Methodology.

The *Basic Chemical Manufacturing* (ANZSIC Group 253) sub-category includes the major bulk chemical manufacturing enterprises producing fertilisers, other nitrogenous chemicals, polymer resins (plastics) and carbon black. The fossil fuel feedstocks used include natural gas (methane), ethane, propane, butane, propylene and naphtha. Ethane, propane and butane may be either 'naturally occurring', i.e. sourced directly from oil and gas fields, or derived from crude oil as by-products of refining. In Australia, all ethane is derived from naturally occurring sources, while both naturally occurring and ex-refinery propane and butane are used. Propylene and naphtha are refinery products. ABARES statistics include ethane within the reported total natural gas consumption, after appropriately adjusting for the different energy content of ethane. ABARES also group propane and butane together as LPG and group propylene and naphtha as petroleum products nec.

The important outputs of this sector can be classified into two components:

- synthetic resins (polymers),
- nitrogenous fertilisers and other nitrogenous products.

A third component, carbon black manufacture, uses significant quantities of fossil fuel feedstock as a source of carbon, however relatively little is combusted. A fourth, methanol, has been manufactured in Australia since 1994.

Synthetic Resins

The balance between combustion and storage in products varies greatly between chemical plants, depending on the production processes involved and the configuration of the particular plant. Therefore the quantity of feedstock supplied to chemical plants is not a useful indication of the quantity of stored carbon. The only reliable guidance comes from the quantities of chemical products produced. The major products in which fossil carbon is sequestered include polyethylene, polypropylene, synthetic rubber and styrene. Other bulk plastics are made in Australia from imported monomers, e.g. PVC made from imported vinyl chloride monomer. These imported monomers contain large quantities of fossil carbon, but since this has not been derived from primary fossil fuels (crude oil, petroleum products and natural gas) produced in or imported to Australia, this carbon is not estimated.

The IPCC Methodology assumes that default fractions of specified fossil fuel products, e.g. ethane, naphtha, are sequestered. The National Inventory utilises the actual production figures provided by the companies making the products concerned. The analysis is nevertheless relatively complex, because most products are derived from several different feedstocks. The carbon contents of the various feedstocks and basic chemical products used in estimating the carbon sequestration are reported in Tables 3.10 and 3.11.

The quantities of feedstocks used in the Chemical sub-sector, and the associated amounts of carbon stored in products, are detailed in CRF Table 1.A(d) – Feedstocks and non-energy use of Fuels. For 2009, 19.4 PJs of ethane and 11.7 PJs of petroleum (naphtha) feedstocks resulted in the storage of 329 kt and 212 kt of carbon in long life products respectively. The emissions of ethane and naphtha combusted as fuels are reported in the National inventory under 1.A.2c Chemicals. In 2009, net emissions from the combustion of ethane were 0.2 Mt CO₂-e, while 2.0 Mt CO₂-e of naphtha emissions were reported.

Carbon Black

Carbon black is produced in Australia by partial oxidation of petroleum feedstocks and used in a variety of long lived products, including tyres.

Table 3.10: Feedstock assumptions in basic chemicals

Feedstock	Carbon Fraction	Calorific Value (GCV)
Ethane	0.80	(a)
Propylene	0.86	52.2
Naphtha (Benzine)	0.84	48.1
Gas Oil (ADO)	0.85	45.6
Carbon Black Feedstock	(a)	(a)

Sources: ABARE and Energy Strategies Analysis. (a) Data is provided on in a confidential manner annually from the relevant companies and hence is not reported here.

Table 3.11: Product assumptions in basic chemicals

Product	Carbon Fraction
Polyethylene	0.86
Polypropylene	0.86
BR / SBR	0.86
Styrene	0.92
Carbon black	1.00

3.4.3 Uncertainties and Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas.

Time series variability of GHG implied emission factors are likely to be influenced by changes in fuel mix within categories, and changes of actual fuel emission factors. Notable examples of where such variations occur in Manufacturing Industries and Construction 1.A.2 are set out below:

- 1.A.2.a *iron and steel* – CO₂: The use of coke in *iron and steel* is reported in *industrial processes* sector in accordance with the IPCC guidelines. Of the two remaining solid fuels: coal and coke oven gas, the coke oven gas has a relatively low CO₂ emission factor of 37 Gg/PJ compared to 91.8 Gg/PJ for coal. This tends to lower the overall CO₂ IEF for solid fuels. Likewise, the liquid fuel CO₂ IEF is also relatively low, being driven by the large and increasing use of LPG (CO₂ emission factor of 60.2 Gg/PJ) compared to other liquid fuels with higher emission factors.
- The IEF trend for solid fuels shows an increase in recent years. This is due largely to steel producers feeding pulverised black coal direct into blast furnaces.
- 1.A.2.c *chemicals*: Emissions and IEFs for *chemicals* are influenced by the mix of end products which sequester carbon. The production mix of the Australian chemicals industry has changed over time, resulting in a variable trend.

3.4.4 Source Specific QA/QC

This source category is covered by the general QA/QC of the greenhouse gas inventory in section 1.6.

3.4.5 Recalculations Since the 2008 Inventory

Recalculations to 1.A.2 *manufacturing and construction* are detailed at the sub-category level in Table 3.12. Key reasons for recalculations include;

- changes to the ABARES national energy statistics as a result of ABARES revisions to estimates in response to improved data and reallocations of fuel use at the sub-category level to improve time series consistency;
- an improved understanding of detailed fuel type use from NGERS data has led to a refinement in the allocation of emission factors to some fuel types reported by ABARES in the AES; and
- NGERS data has provided greater detail on the use of fuel as reductants which has facilitated the reallocation of some fuels previously reported as fuel combustion to the Industrial Processes sector.

Table 3.12: 1.A.2 Manufacturing and Construction: recalculation of total CO₂-e emissions, 1990-2008

	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1.A.2.a Iron and steel (a)				
1990	2,636.1	2,640.6	4.5	0.17
2000	2,415.5	2,423.4	7.9	0.33
2001	2,445.1	2,444.8	-0.2	-0.01
2002	2,657.4	2,659.5	2.0	0.08
2003	2,836.9	2,837.1	0.3	0.01
2004	3,254.3	3,254.3	0.0	0.00
2005	3,399.5	3,400.1	0.6	0.02
2006	3,280.4	3,280.3	-0.1	0.00
2007	3,457.3	3,456.9	-0.4	-0.01
2008	3,506.7	3,598.4	91.6	2.61
1.A.2.b Non-ferrous metals (b)				
1990	11,351.3	10,836.0	-515.3	-4.54
2000	13,460.8	12,945.5	-515.3	-3.83
2001	12,697.7	12,182.1	-515.6	-4.06
2002	12,898.6	12,381.7	-516.9	-4.01
2003	13,260.5	12,745.7	-514.9	-3.88
2004	13,566.5	13,049.0	-517.5	-3.81
2005	14,800.3	14,281.1	-519.1	-3.51
2006	14,690.2	14,169.9	-520.3	-3.54
2007	15,108.2	14,711.5	-396.7	-2.63
2008	16,273.5	15,825.2	-448.4	-2.76
1.A.2.c Chemicals (c)				
1990	5,708.1	5,801.7	93.6	1.64
2000	6,345.5	6,397.6	52.1	0.82
2001	6,635.7	6,883.3	247.6	3.73
2002	6,212.0	6,415.0	203.0	3.27
2003	5,485.3	5,833.6	348.3	6.35
2004	5,688.1	5,661.9	-26.1	-0.46
2005	5,420.3	5,421.9	1.6	0.03
2006	5,348.0	6,029.6	681.6	12.75
2007	5,097.9	5,891.8	793.9	15.57
2008	5,474.0	5,911.1	437.1	7.98

	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)		(Gg CO ₂ -e)	(%)
1.A.2.d Pulp paper and print				
1990	1,300.1	1,300.6	0.5	0.04
2000	1,783.2	1,783.3	0.1	0.01
2001	1,839.1	1,839.3	0.1	0.01
2002	1,830.8	1,831.0	0.1	0.01
2003	1,805.0	1,805.2	0.1	0.01
2004	2,049.1	2,049.2	0.1	0.01
2005	2,298.0	2,298.1	0.1	0.01
2006	2,319.0	2,319.1	0.1	0.01
2007	2,303.4	2,303.5	0.1	0.01
2008	2,218.3	2,218.4	0.1	0.01
1.A.2.e Food, beverages and tobacco (d)				
1990	2,988.0	2,988.8	0.7	0.02
2000	3,205.6	3,211.4	5.8	0.18
2001	2,609.8	2,609.9	0.1	0.01
2002	2,609.6	2,609.7	0.1	0.00
2003	2,659.5	2,659.6	0.1	0.00
2004	2,711.4	2,711.4	0.0	0.00
2005	3,975.7	3,280.6	-695.1	-17.48
2006	3,787.3	3,075.1	-712.2	-18.80
2007	3,526.1	2,824.3	-701.8	-19.90
2008	3,645.5	2,916.0	-729.6	-20.01
1.A.2.f Other (e)				
1990	12,078.6	12,080.3	1.7	0.01
2000	12,092.1	12,093.5	1.4	0.01
2001	12,314.7	12,329.8	15.1	0.12
2002	13,818.7	13,837.6	18.9	0.14
2003	15,487.0	15,527.6	40.6	0.26
2004	15,870.1	15,222.2	-647.9	-4.08
2005	16,087.2	15,529.0	-558.3	-3.47
2006	15,967.1	15,292.5	-674.5	-4.22
2007	16,879.3	16,134.0	-745.3	-4.42
2008	17,569.1	16,716.4	-852.7	-4.85

Recalculation explanation

(a) Minor revision of activity data

Minor adjustment to fuel oil activity data as a result of QC processes

(b) ABARE Australian Energy Statistics

Reallocation of emissions to the Industrial processes sector

Reallocation of diesel use to the Transport sector and changes to natural gas use for 2008

NGERS data has provided greater detail on the use of fuel as reductants. This has facilitated the reallocation of some fuels previously reported as combustion to the Industrial processes sector. For non ferrous metals, this has affected minor amounts of black coal, coking coal, coke, petroleum coke and fuel oil used in metal production.

Improved facility data

A revision to black coal consumed as a reductant in the production of synthetic rutile for 2007 and 2008 has been undertaken to account for the improved facility data

(c) Improved facility data

Improved understanding of fuel type use from NGERS data

A revision to petroleum coke consumed as a reductant in the production of titanium dioxide production for 2007 and 2008 has been undertaken to account for the improved facility data.

A revision to the quantities of natural gas consumed in the production of ammonia has been undertaken due to improved facility data.

Previously an aggregated fuel type (BTX) was used, along with a representative emission factor. However NGERS data has provided the actual fuel type use which has allowed reinterpretation of fuel use reported by ABARE as being liquefied aromatic hydrocarbons and coal tar. Therefore, the respective EFs have been applied to the individual fuel types.

(d) ABARE Australian Energy Statistics

Reallocation of diesel use to the Transport sector for 2005 to 2008.

(e) ABARE Australian Energy Statistics

Reallocation of diesel use from mining to the Transport sector for 2004 to 2008. Changes to natural gas for 2006 to 2008

Improved understanding of fuel type use from NGERS

Previously an aggregated fuel type (BTX) was used along with a representative emission factor. However NGERS data has provided the actual fuel type use which has allowed reallocation of fuel use reported by ABARE as being liquefied aromatic hydrocarbons and coal tar. Therefore, the respective EFs have been applied to the individual fuel types.

Change in EF for petroleum products

The EF for petroleum use as reported by ABARE in the AES was changed from 68.6 Gg CO₂/PJ to that of the IPCC 1997 default for refinery feedstocks and other oil (69.7 Gg CO₂/PJ) as it was considered more appropriate for the fuel types consumed.

3.4.6 Planned Improvements

Uncertainty data reported by corporations under the National Greenhouse and Energy Reporting system will be analysed to see how they can be incorporated into the national inventory estimates of uncertainty.

Some of the country specific emissions factors for liquid fuels used in the national inventory were estimated more than ten years ago. A study is planned to assess these emission factors and to assess whether any updates are required.

3.5 SOURCE CATEGORY 1.A.3 – TRANSPORT

3.5.1 Source Category Description

This source category includes emissions from the transport sector, comprising the civil aviation, road transportation, marine navigation, railways and 'other' categories.

Activity data on fuel consumption is based on ABARES 2010. A number of mobile source categories have been allocated to the stationary source inventory because the current national data collection methods do not allocate this fuel to the transport sector but rather to the specific ANZSIC class in which it is used. In particular, emissions from miscellaneous off-road vehicles used in specific ANZSIC classifications (such as tractors and other farm vehicles, forestry vehicles, quarry trucks and front-end loaders, construction equipment, and forklifts) are allocated to the corresponding ANZSIC group and accounted for in sectors 1.A.2 and 1.A.4. It is estimated that these emissions account for approximately 17,000 Gg in 2009. More information on the assumed mobile components of stationary sources is at section 3.2.2. Emissions from mobile utility engines (such as lawn-mowers, chain-saws, portable generators and mobile compressors) and military transport are reported in sectors 1.A.4 and 1.A.5 using the methodologies detailed in this sector. Emissions from other off-road mobile sources, however, such as unregistered trail bikes, recreation vehicles and competition vehicles are reported under 1.A.3.

3.5.2 Methodology

Like other energy sub-sectors, the methodology for 1.A.3 is based on the application of 'bottom up' approaches to the estimation of emissions. Tier 2 approaches are employed for the estimation of non-CO₂ emissions from civil aviation, which takes account of fuel consumed, landing and takeoff cycles and Australian fleet characteristics. The estimation of non-CO₂ emissions from passenger and light commercial vehicles utilises a Tier 3 approach that depends on data on vehicle kilometres travelled, vehicle fleet characteristics and vehicle operating modes.

Table 3.13: Summary of methods and emission factors: Transport

Source Category	CO ₂		CH ₄		N ₂ O	
	Method applied	Emission factor	Method applied	Emission factor	Method applied	Emission factor
1A3a Civil Aviation	T2	CS	T2	CS/D	T2	CS/D
1A3b Road Transportation – passenger, light commercial and heavy vehicles	T2	CS	T3	CS/D	T3	CS/D
1A3b Road Transportation – other	T2	CS	T1	CS/D	T1	CS/D
1A3c Railways	T2	CS	T1	D	T1	D
1A3d Navigation (Domestic)	T2	CS/D	T2	D	T2	D
1A3e Other Transport	T2	CS	T1	D	T1	D

Notes: T1 = Tier 1. T2 = Tier 2. T3 = Tier 3. CS= Country-specific. D= IPCC default.

General Methodology

The emission estimate of a greenhouse gas from fuel combustion in the engines of a mobile source, using a specified fuel type, is calculated by:

$$E(l)_{ijk} = Au_{ijk} \times F(l)u_{ijk} \dots\dots\dots (3.3)$$

where: $E(l)_{ijk}$ is the emission of greenhouse gas l in gigagrams (Gg) from a mobile vehicle and age class i and technology j using fuel type k ;
 Au_{ijk} is the activity level, where u refers to either energy consumption in petajoules (PJ) or to distance travelled in kilometres (km); and
 $F(l)u_{ijk}$ is the emission factor, in units of grams of gas l emitted per megajoule of energy use (g/MJ) for CO₂ and SO₂, and grams of gas l emitted per kilometre travelled (g/km) for other non-CO₂ gases.

Fuel consumption data for the *transport* sector are taken from ABARE 2010. The main assumptions applied to ABARES energy consumption data allocate liquid fuels to off-road and military fuel uses (reported in Table 3.A.13).

The allocations of fuel to military transport have been improved in this submission as a result of direct reporting of fuel consumption by the Department of Defence. This new data has facilitated an update of the proportions of each fuel allocated to sector 1.A.5. Allocations were made in 2008 and 2009 based on direct data from the Department of Defence. Allocations were updated for 1995-2007 by linearly extrapolating the proportions between the previous reported data point (1994) and the latest reported data point (2008). It is anticipated that these proportions will be updated on an annual basis in the future. The shares used to allocate fuel consumption are reported in Table 3.A.13.

Civil Aviation (1.A.3a)

The estimation of carbon dioxide emissions from civil aviation is undertaken using a Tier 2 methodology and emission factors given in Tables 3.2 and 3.3. Non-carbon dioxide emissions from domestic civil aviation from fuel use are estimated using a Tier 2 methodology. Emissions are calculated as a function of both the landing/takeoff cycles (LTOs) and of cruise emissions for both domestic and international aircraft.

The estimation of emissions from landing and takeoff cycles requires data on the number of LTO cycles at Australian airports; data on the profile of the Australian aviation capital stock or fleet; and emission factors by type of aircraft. The data required for the total yearly LTO for the domestic and international aircraft are available from the Bureau of Infrastructure, Transport and Regional Economics within the Department of Infrastructure and Transport. The Australian aviation fleet profile is developed using the Australian Aircraft Register which is available from the Civil Aviation Safety Authority (Table 3.14). Emission factors for each aircraft type are taken from IPCC 2006 and are used to estimate weighted average LTO cycle emission factors for the domestic/interstate, regional and international aviation fleets

(Table 3.15). These emission factors most accurately reflect the technology and aircraft types currently in the Australia aircraft fleet. In a couple of instances emission factors are not available for a certain aircraft type that exists in the Australian fleet. These aircraft are allocated to the aircraft type, for which an emission factor exists, that most closely reflects the aircraft's engine characteristics.

The estimation of cruise emissions is a function of fuel use, after deduction of fuel consumption required for the LTO cycles, and cruise emission factors. Data on the yearly fuel consumption for domestic and international activity are available from ABARE 2010. Cruise emission factors are taken from IPCC (1997) (Table 3.16).

The methodology is applied to each of the eight Australian States and Territories. Differences in emission estimates across the States principally reflect differences in fuel consumption and both the number of LTO cycles and the relative importance of major interstate relative to regional LTO cycles, which impacts on the aircraft type that use State airports. National emissions are estimated as the sum of the State and Territory emissions.

Emissions from international aviation are also estimated, but are reported as a Memo item only, by international agreement.

Activity data for international bunkers is estimated by ABARES as part of the Australian Energy Statistics. ABARES utilise data published by the Department of Resources Energy and Tourism (DRET) in their Australian Petroleum Statistics (APS).

Monthly national and state petroleum statistical information are published in the APS. Quarterly sales of aviation turbine fuel, diesel and fuel oil for domestic and international uses are published on a quarterly basis.

The APS explanatory note outlines that the dissection of international and domestic/coastal is according to the predominant mode of usage by the consumer.

APS domestic aviation turbine fuel consumption is cross checked with tax and customs data. The comparison showed a difference of 4.4% for 2008 (the latest year that customs data was available). The inclusion of military transport in the APS data is expected to contribute to this difference.

Independent of the national inventory the Department of Infrastructure and Transport has developed a software tool to compute and track the carbon footprint associated with aircraft fuel uplifted in Australia. The Department of Infrastructure and Transport completed an assessment of the robustness of their results by comparing their calculated values with the APS. Their results showed that over the 2008-09 and 2009-10 period computed fuel sales using the software tool and APS estimates differed by 3.7% for domestic consumption and 2.6% for international consumption. This is concluded to be a strong independent verification of the estimates.

The estimated weighted emission factors have been held constant throughout the time series 1990–2004, and are adjusted in line with changes in the aviation capital stock from 2004 onwards.

Table 3.14: The Australian aircraft fleet, 2009, and emission factors by type of aircraft

Type of aircraft		Emission Factors				
		CH ₄	N ₂ O	NO _x	CO	NMVOC
		kg/LTO	kg/LTO	kg/LTO	kg/LTO	kg/LTO
Major inter-state						
DHC-8-100	41	0.00	0.02	1.51	2.24	0.00
DHC-8-200	12	0.00	0.02	1.51	2.24	0.00
A320	60	0.06	0.10	9.01	6.19	0.51
A330–200/300	25	0.13	0.20	35.57	16.20	1.15
BAE 146	36	0.14	0.00	4.07	11.18	1.27
B717	11	0.01	0.10	10.96	6.78	0.05
B727-200	2	0.81	0.10	11.97	27.16	7.32
B737–300/400/500	23	0.08	0.10	7.19	13.03	0.75
B737–700	19	0.09	0.10	9.12	8.00	0.78
B737–800	72	0.07	0.10	12.30	7.07	0.65
B767–200	0	0.33	0.10	23.76	14.80	2.99
B767–300	26	0.10	0.20	28.19	14.47	1.07
Regional						
SAAB 340	53	0.00	0.02	1.51	2.24	0.00
SA227	51	0.00	0.02	1.51	2.24	0.00
SA226	11	0.00	0.02	1.51	2.24	0.00
Gulfstream IV	38	0.14	0.10	5.63	8.88	1.23
EMB 110	8	0.06	0.01	0.30	2.97	0.58
EMB 120	21	0.00	0.02	1.51	2.24	0.00
Cessna 525	12	0.33	0.03	0.74	34.07	3.01
Beech 200	109	0.06	0.01	0.30	2.97	0.58
F27	121	0.03	0.02	1.82	2.33	0.26
International						
747–300	0	0.27	0.40	65.00	17.84	2.46
747–400	36	0.22	0.30	42.88	26.72	2.02
777	5	0.07	0.30	52.81	12.76	0.59

Source: CASA Civil Aircraft Register (2010), International Civil Aviation Organisation, Engine Exhaust Emissions Databank (2004).

Table 3.15: Weighted average emissions factors per Landing and Take Off cycle

Fleet	CH ₄ (kg)	N ₂ O (kg)	NO _x (kg)	CO (kg)	NM VOC (kg)
Domestic/Interstate Fleet	0.08	0.09	11.49	8.45	0.69
Regional Fleet	0.05	0.02	1.61	3.96	0.43
International Fleet	0.20	0.30	44.09	25.02	1.85

Source: DCCEE estimates.

Table 3.16: Aviation cruise emission factors (grams per tonne of fuel consumed)

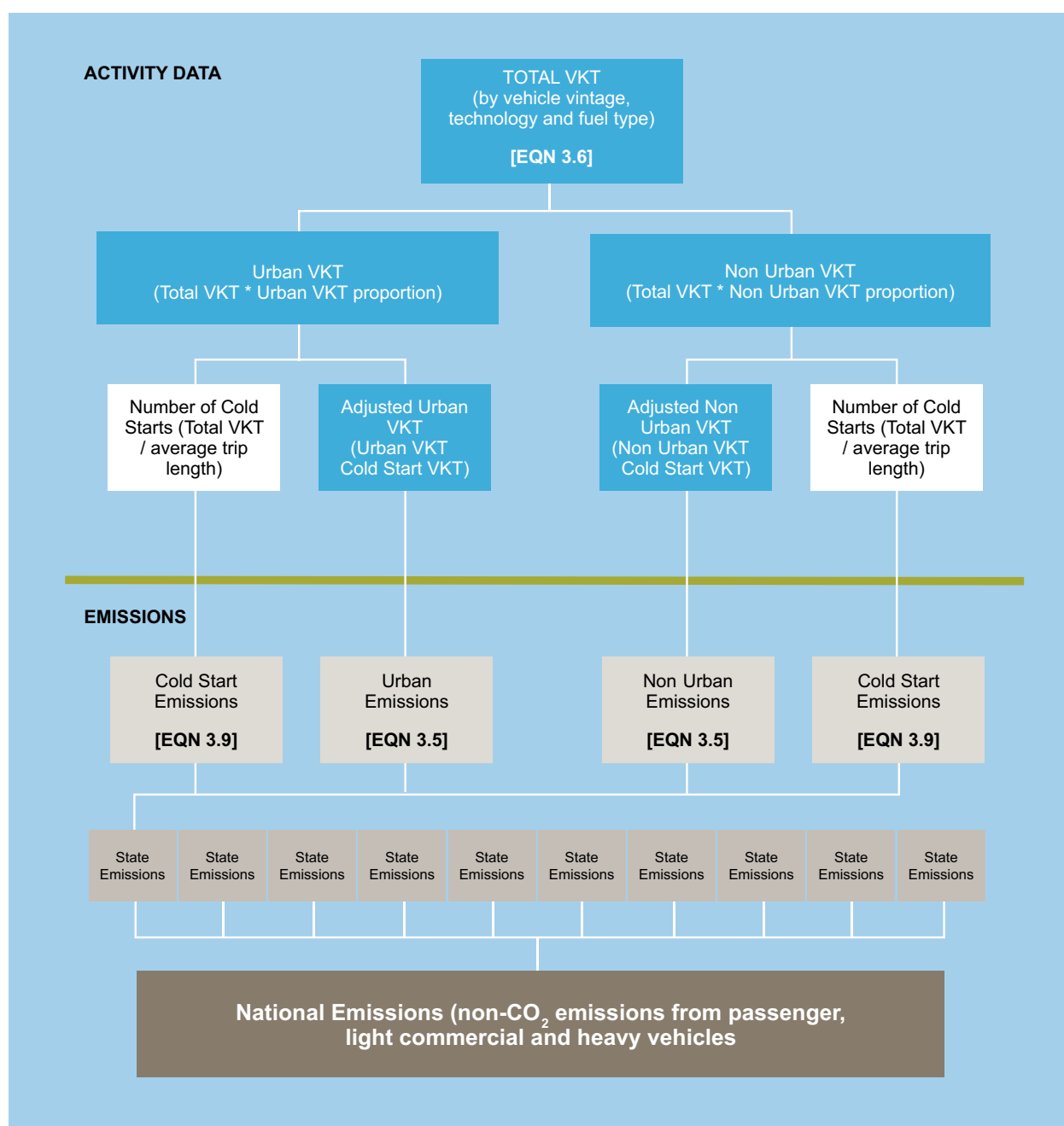
Fleet	CH ₄ (g/t)	N ₂ O (g/t)	NO _x (g/t)	CO (g/t)	NM VOC (g/t)
Domestic and Regional Fleets	0	0.1	11	7	0.7
International Fleet	0	0.1	17	5	2.7

Source IPCC (1997).

Road Transportation (1.A.3.b)

Like the aviation sector, the estimation of carbon dioxide emissions from the road transport sector is based on a Tier 2 method with emission factors given in Table 3.2. The estimation of non-carbon dioxide emissions is based on a Tier 3 method, with the emission estimates dependent on the type of vehicle, the age of the vehicle capital stock, technology, operating mode (cold versus hot) and road type (urban versus non-urban). Activity data is expressed in terms of vehicle kilometres travelled and emission factors are expressed in terms of g/km. The methodology is applied to each of the eight Australian States and Territories. Differences in emission estimates across the States principally reflect differences in fuel consumption and the impacts on non-CO₂ emission estimates of differentials in the age distribution of each State's vehicle fleet. National emissions are estimated as the sum of the State and Territory emissions (see Figure 3.10).

Figure 3.10: Methodology for the estimation of non-CO₂ emissions from passenger and light commercial vehicles



Passenger and Light Commercial Vehicles, Heavy Vehicles and Buses (1.A.3.bi-v)

Carbon dioxide emissions from all vehicle fuel sources have been estimated based on the quantity of fuel consumed by the CO₂ emission factor specific to that fuel and the proportion of that fuel which is completely oxidised.

$$E_{ijk} = A_{ijk}^{u=1} \times (F(l)_k \times P_k) \dots\dots\dots (3.4)$$

where: $F(l)_k$ is the CO₂ emission factor applicable to complete oxidation of fuel carbon content for fuel type k (where k=petrol, diesel and LPG);
 P_k is the proportion of fuel that is completely oxidised upon combustion; and
 $A_{ijk}^{u=1}$ is the activity data for vehicle type i with emission control technology j and fuel type k (and where u=1 for fuel consumption in each Australian State)

The CO₂ emission factors and oxidation factors for each fuel are summarised in Tables 3.2 and 3.3.

For all vehicles besides motorcycles consuming automotive gasoline, ethanol, diesel and LPG non-CO₂ emissions for each age class are estimated based on vehicle kilometres travelled in each State or Territory; the profile and age of the vehicle capital stock in each State; the penetration of catalytic control technology; mode of operation and road type; and vehicle and fuel specific emission factors.

It is assumed that all light duty vehicles go through a cold start phase for each trip which is associated with higher emissions due to engine and catalyst temperatures that are below optimum. The number of cold starts is derived from total VKT and an average trip length of 10km (based on data cited in VicHealth (1999), QLD Transport (2001), and EPA NSW (2000)). A cold-start duration of 3km (as cited in IPCC 2006) is used to determine the total cold start VKT. This is subtracted from total VKT to derive an adjusted total VKT value.

Emission factors vary by road type (urban versus non-urban) to reflect the different driving conditions and engine operating profiles. Distance travelled is disaggregated into urban and non-urban VKT in each State based on data in Apelbaum Consulting Group 2006.

Vehicles using automotive gasoline, ethanol, diesel and LPG are further classified by age of vehicle (using data contained in ABS 2009a). The divisions in the vehicle fleet enable differences in emissions control technology and differences in fuel efficiency across age classes to be factored into the emissions estimation. Passenger vehicles and light commercial vehicles manufactured and sold in Australia before 1976 are assumed to have no emissions control equipment. The 1976-1985 group uses a variety of non-catalytic control (such as exhaust gas recirculation) and the 1985-1997, 1998-2003, 2004-2005 and the post-2005 groups use catalytic control.

In general, non-carbon dioxide exhaust emissions from vehicles have been calculated by the following form of equations:

$$E(l)_{ijk} = A_{ijk}^{u=2} \times EF(l)_{ijk} \dots\dots\dots (3.5)$$

where: l = non-CO₂ gases; $A_{ijk}^{u=2}$ for vehicle kilometres travelled & k= automotive gasoline, diesel, and LPG;
 $EF(l)_{ijk}$ is the exhaust emission factor for gas l from vehicle type i and age class j using fuel type k for urban and rural operation in each state or territory.

where vehicle distances travelled during the hot-engine phase of operation are related to energy consumption levels using:

$$A_{ijk}^{u=2} = A_{ijk}^{u=1} / R_{ik} \times D_k \quad (3.6)$$

where: $A_{ijk}^{u=2}$ is the distance travelled for vehicle type i and age class j, using fuel type k = automotive gasoline, diesel, and LPG; and
 R_{ik} is the average rate of fuel consumption (in l/km, given in Tables 3.A.15-3.A.17) for vehicle type i and age class j, using fuel type k; and
 D_k is the energy density of fuel type k (in MJ/L)

and where

$$EF(I)_{ijk} = (ZKL_{ijk} + DR_{ijk} \times CumVKT_{ijk}) \quad (3.7)$$

where: $EF(I)_{ijk}$ is the emission factor for gas I from each vehicle type i and age class j, using fuel type k=automotive gasoline, diesel, and LPG;
 ZKL_{ijk} is the zero kilometre level emissions of a gas I from vehicle type i and age class j;
 DR_{ijk} is the deterioration rate for vehicle type i and age class j; and
 $CumVKT_{ijk}$ is the cumulative VKT for vehicle type i and age class j, and fuel type k, in each state or territory

and where

$$CumVKT(I)_{ijk} = \sum_{t=1-n} A_{ijk}^{u=2} \quad (3.8)$$

where $A_{ijk}^{u=2}$ is the average distance travelled (in km) by vehicle type i and age class j, using fuel type k=automotive gasoline, diesel, and LPG in each State or Territory summed over time.

Cold start emissions are derived using equation 3.9:

$$Ecs_{ijk} = CS_{ijk} \times EFcs_{ijk} \quad (3.9)$$

where: Ecs_{ijk} are the cold start emissions for vehicle type i and age class j, using fuel type k = automotive gasoline, diesel, and LPG;
 CS_{ijk} is the number of cold starts for vehicle type i and age class j, using fuel type k = automotive gasoline, diesel, and LPG;
 EF_{ijk} is the cold start emission factor (g/start) for vehicle type i and age class j, using fuel type k = automotive gasoline, diesel, and LPG

Data on fuel consumption for individual vehicle types is derived from ABARE 2009a and ABS (2008a and various issues). The data on fuel consumption rates are taken from ABS (2008a, and various issues). The profile and age of the passenger vehicle stock in each State required for equation 3.7 is taken from ABS (2009a, and various issues). The vehicle stock from each historical year varies largely due to vehicle sales from each particular year, which in turn is largely driven by the prevailing economic conditions. For example the vehicle stock in 1991 is lower than surrounding years which is likely due to lower vehicle sales due to an economic recession affecting Australia at the time. Data required to estimate VKT for individual vehicle and age classes are given in Tables 3.A.15 to 3.A.17.

Emissions of CH_4 from motor-vehicles are a function of the emission and combustion control technologies present as well as vehicle operating conditions. Emission factors chosen for passenger and light commercial vehicles were obtained from Australian sources where these were available and applicable to the vehicle fleet and its various modes of operation and fuel types (see Table 3.19). A major empirical study of emissions from the operation of light duty petrol vehicles was undertaken in 2009. The results of this study were analysed for this national inventory (Orbital 2011(see box 3.1)). The use of disaggregated, country-specific emission factors expressed in terms of emissions per kilometre travelled is consistent with the IPCC tier-3 methodology. Where country-specific emission factors were not available IPCC 2006 has been used as a source reflecting the most recent technology-specific information available on the performance of transport vehicles.

Box 3.1: Emissions Factor Derivation from NISE 2 (Second National In-Service Emission Study) Measurements of Vehicles from the Australian Fleet

The DCCEE engaged Orbital Australia to derive factors for use in the National Inventory based on the NISE 2 dataset. Orbital were asked to estimate emission factors (EFs, g/km) and deterioration rates (DRs, g/km/km) for several GHG pollutant species and under various operating conditions.

Emission factors were estimated for petrol passenger vehicles and LCVs. Emission factors were estimated for NO_x, CO, CH₄ and N₂O for four ADR (Australian Design Rule) age groupings (ADR 37/00, ADR 37/01, ADR 79/00, ADR 79/01) under hot urban, non-urban and cold urban driving conditions. Deterioration rates were also estimated for each gas and vehicle design combination.

Table 3.17 sets out the year of manufacture of the vehicles analysed in the Orbital Australia report.

Table 3.17: Year of manufacture and ADR of vehicles analysed

Australian Standard	Year of Manufacture
ADR 37/00	1994-1998 (3-way catalytic converters)
ADR 37/01	1999-2003
ADR 79/00	2004-2005
ADR 79/01	2006-2009

NISE 2 Dataset

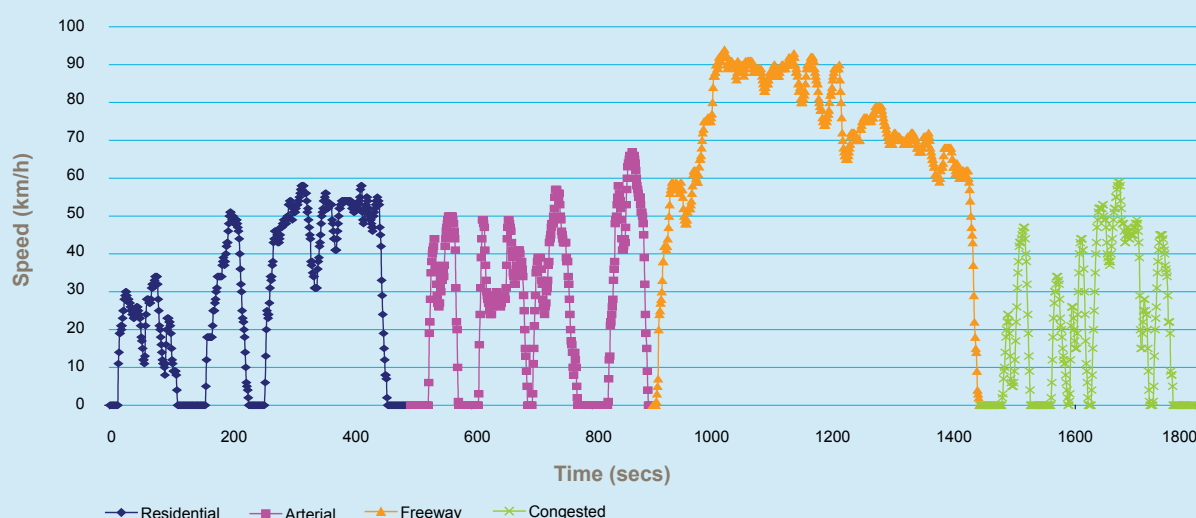
The NISE 2 dataset represents the latest survey of Australian in-service emissions for the petrol vehicle fleet. The NISE 2 consists of results from 347 vehicles.

The 347 vehicles represent four ADR (Australian Design Rule) age groupings and six vehicle sub-types. Within each age/type grouping the vehicle make/model representation is approximately proportional to their representation of the Australian vehicle fleet as at the time of the NISE 2 study.

At the time of testing 345 of the 347 vehicles tested had a VKT of between 0 and 300,000km.

The Petrol Composite Urban Emissions Drive Cycle for light duty vehicles (Petrol CUEDC) was developed as part of the NISE 2 project as a means of better representing driving under Australian conditions. The drive cycle consists of four phases representing residential, arterial, freeway and congested traffic conditions and is shown graphically in Figure 3.11.

Figure 3.11: Drive Trace for Australian Petrol CUEDC (light duty petrol vehicles)



All vehicles, regardless of ADR group, undertook a hot start CUEDC. Some vehicles were scheduled to perform multiple hot start CUEDCs as part of assessing repeatability. As this drive cycle was a constant across all vehicle categories and ADR groupings this formed the basis of the analysis. The Australian Petrol CUEDC cycle consists of four driving phases; three representing urban and one representing freeway (highway or Non-Urban) driving. For the purpose of the calculation of hot EFs and DRs the urban cycle is assumed to be phase1 (residential) portion of the CUEDC. This allows direct comparison to cold data which is only available for phase1. The Non-Urban cycle is assumed to be the phase3 (freeway) portion of the CUEDC.

Only a subset of the vehicles (approximately half) undertook cold start CUEDCs.

Total hydrocarbon, CO, NO_x, CO₂ and CH₄ emissions were measured for all vehicles in the NISE 2 project. These pollutants were measured by convention analytical benches, from dilute bag samples for the respective drive cycle phase. N₂O emissions were measured directly for only a sub-fleet of 100 vehicles. N₂O emissions were measured by integration of FTIR readings taken over the drive cycle using an MKS MultiGas™ 2030 HS. CO₂ and CH₄ emissions were used as the cross-reference between the two measurement systems.

The vehicle count of available datasets were as shown in Table 3.18. This shows that the datasets were complete (equal to 347) for Hot Urban and Non-Urban (hot) regulated emissions, but not for N₂O and Cold Urban. The latter would require some data substitution in order to complete the datasets. In undertaking the analysis for this inventory study, the NISE 2 dataset is reviewed and where necessary (for example for Cold Urban conditions and N₂O pollutant results) expanded by substituting data not available from vehicles actually tested. This substitution process is cross-checked where possible before emission factors are calculated for the required vehicle groupings.

Table 3.18: NISE 2 Dataset Size

	THC, CO, NO _x , CO ₂ , CH ₄	N ₂ O
Hot Urban	347	96
Non-Urban	347	95
Cold Urban	175	39

Calculation of Emissions Factors

The term Emission Factor (EF) can take on different meanings depending upon the data source from which it is calculated. In the Orbital Australia report, emission factors are defined in the form below:

$$EF_b = EF_0 + DR * kms + [other factors]$$

For the study by Orbital Australia, EF_b is the parameter which is calculated directly from the NISE 2 data. The NISE 2 test data represents 347 individual vehicle results (in g/km) at different VKT. The average of these directly gives the emission rate (or emission factor) of the particular grouping at the grouping's average VKT.

Emission factor results (EF_b) are summarised in appendix tables 3.A.5.

DR indicate how vehicle emissions change with increasing VKT.

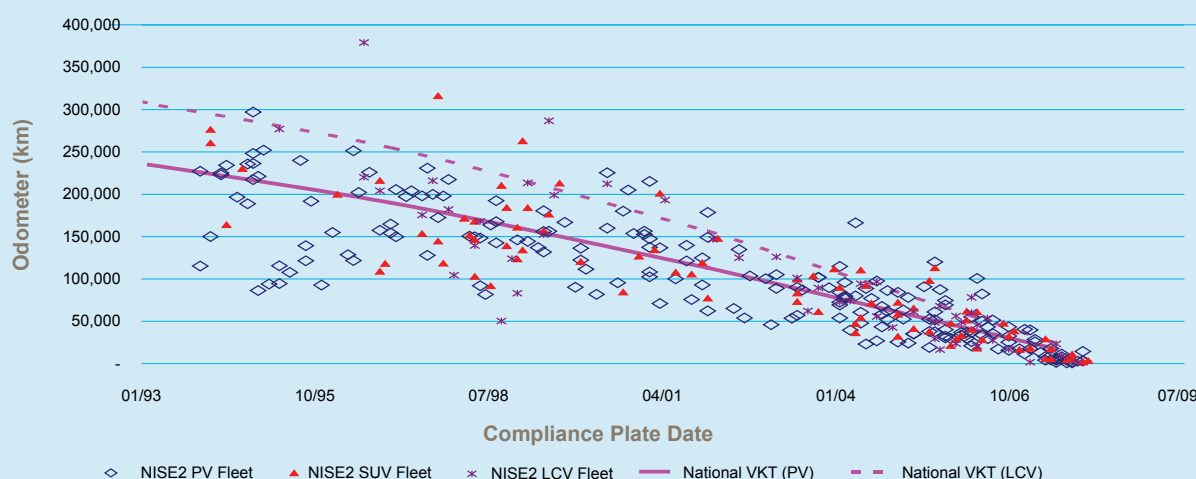
EF_0 can be derived as the emissions factor at 0 VKT and is estimated from the calculated EF_b and DR.

Calculation of Deterioration Rates

Deterioration Rates are calculated to indicate how vehicle emissions may change with increasing VKT. Generally, deterioration rates are positive indicating that emissions increase with mileage as vehicle emission control systems degrade.

The NISE 2 study can be considered a “snapshot in time” where the vehicles sampled were targeted to have average VKT for their age. The list of vehicle odometer readings are graphically represented in Figure 3.12.

Figure 3.12: NISE 2 Test Vehicle Odometer reading at Start of Test



Deterioration rates for each gas, vehicle design category and vehicle type combination are calculated by fitting a linear regression to the scatter of measured emissions (see Figures 3.13-3.20). Since the dataset includes a range of results, the regression line is not automatically generated with confidence. The fitted line is adjusted taking into account the vehicle technology, initial compliance level and type of drive cycle.

Only a single DR is estimated for each gas/vehicle age/type sub-category. It should be noted that DR values can be sensitive to the test cycle used. For catalyst equipped vehicles, which are the vast majority in the NISE 2 study, the Cold Urban DRs will largely be a function of catalyst light-off time degradation with VKT. The Hot Urban and Non-Urban DRs will be a function of catalyst efficiency at the exhaust temperatures and flow rates experienced during the specific drive cycle. The DRs provided in this study are based on the Hot Urban data. As a check on the

reasonableness of this approach, Orbital Australia have generated Cold Urban and Non-Urban DRs for a subgroup of vehicles. This check has indicated that Cold Urban DRs are expected to be higher, and Non-Urban DRs lower, than the Hot Urban DRs.

Deterioration rates are summarised in Appendix Table 3.A.11. The NISE 2 data showed no evidence of deterioration in the level of N_2O emissions, therefore the assumption from the previous inventory submission of an N_2O DR of 0 was retained.

Verification of Generated EF and DR Values

The reasonableness of calculated EF and DR results can be assessed by cross-referencing the generated results to a known or expected result. For this data, the reference point is the zero kilometre capability of the vehicle fleet. This reference point is based on the assumption that at zero kilometres the vehicles were generally in compliance with emission standards of the day, and that in general the deterioration over the ADR specified period is indicated to be in line with automotive engineering expectations.

This rationality check has been detailed in the Orbital report. The methodology used is to take the generated EF and DR results, scale them back to a zero-kilometre value, then combine the various driving modes as prescribed by the relevant ADR. A deterioration factor equivalent to ADR79/01 can then be computed. ADR79/01 stipulates a deterioration factor of 1.2 in the absence of experimental data.

The results show that that most vehicle categories would have ADR deterioration factors of industry expected levels. There are a few instances where the ADR deterioration factor is higher than expected. Detailed results and possible reasons for variations are outlined in the Orbital Report.

Zero mileage EFs (EF_0)

EF_0 is not specifically presented in the Orbital report. However EF_0 estimates can be derived based on the EF_b and DR values. These EF_0 estimates are presented in appendix tables 3.A.6 and 3.A.8 and are utilised in the national inventory model.

Comparison with emission factors and deterioration rates used in previous submissions

Compared to the EF and DR values in the previous inventory submission the NISE 2 based results generally show:

- Lower EF and DR values for CO , NO_x and CH_4 , across all ADR groups
- Lower EF for N_2O , but a DR of zero as is currently assumed

A number of explanations for these differences are outlined in the orbital report.

A comparison of the new and old CH_4 and N_2O EFs and DRs in the hot urban mode of operation for each ADR is outlined in Figures 3.13 to 3.20. The blue dots list the data collected for the individual vehicles in this age classification, the black linear line shows the derived DR and the red line shows the EFs and DR used in the previous national inventory submission.

Figure 3.13: ADR37/00 CH₄ Hot Urban Estimations – Passenger Vehicles

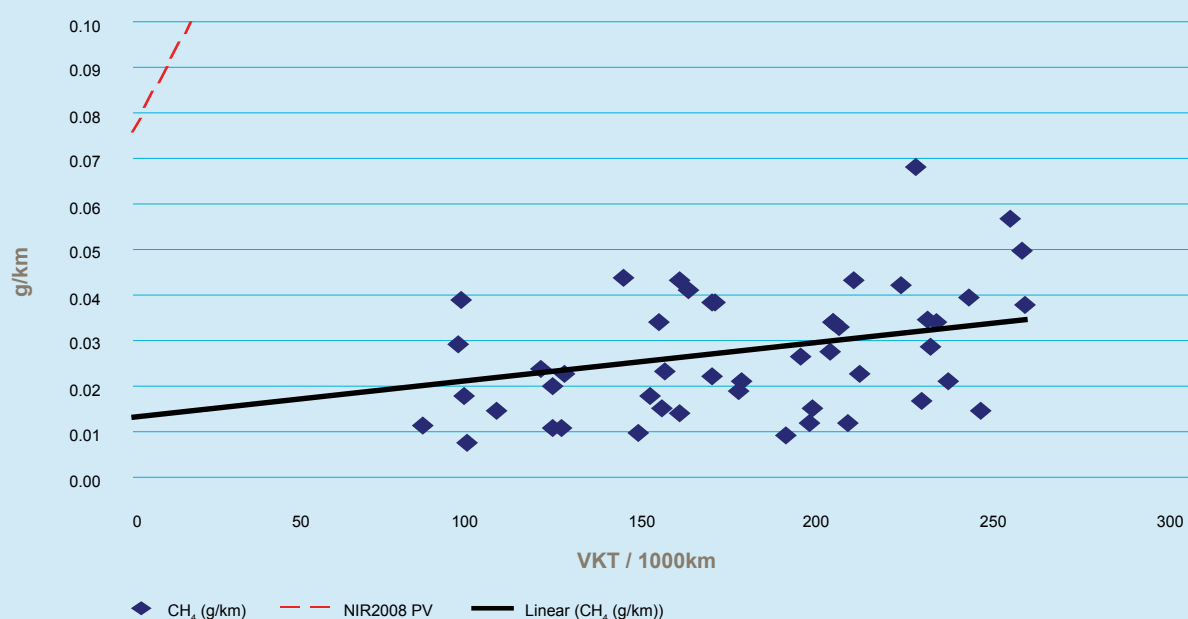


Figure 3.14: ADR37/00 N₂O Hot Urban Estimations – Passenger Vehicles

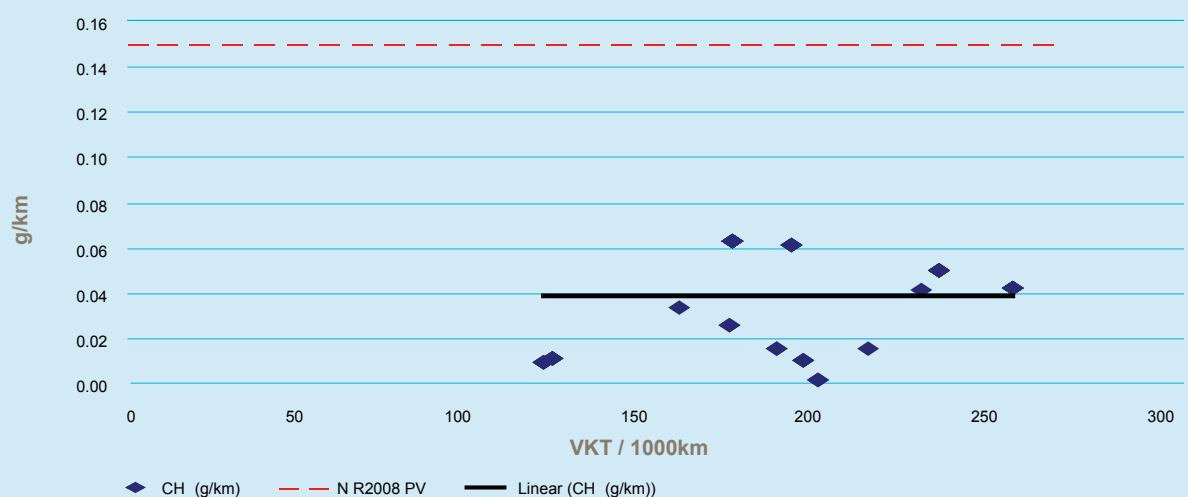


Figure 3.15: ADR37/01 CH₄ Hot Urban Estimations – Passenger Vehicles

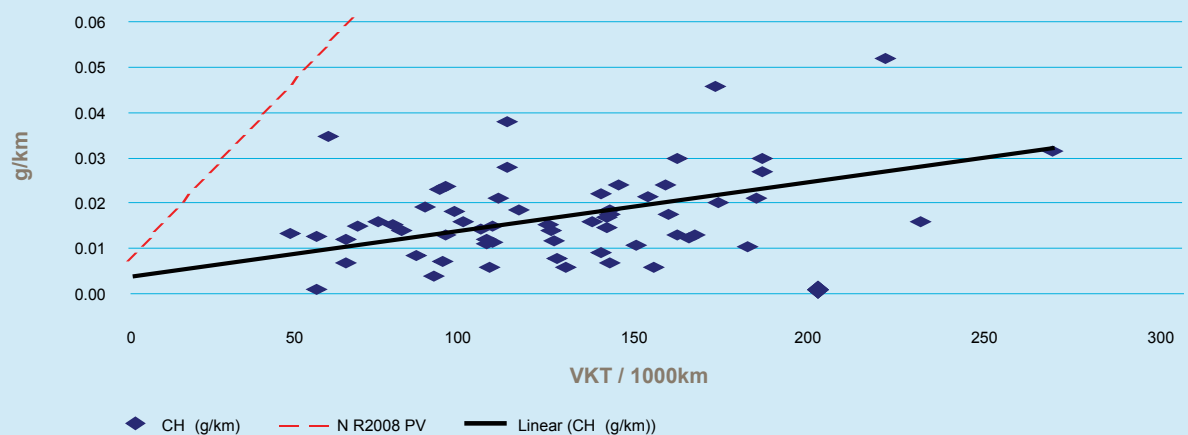


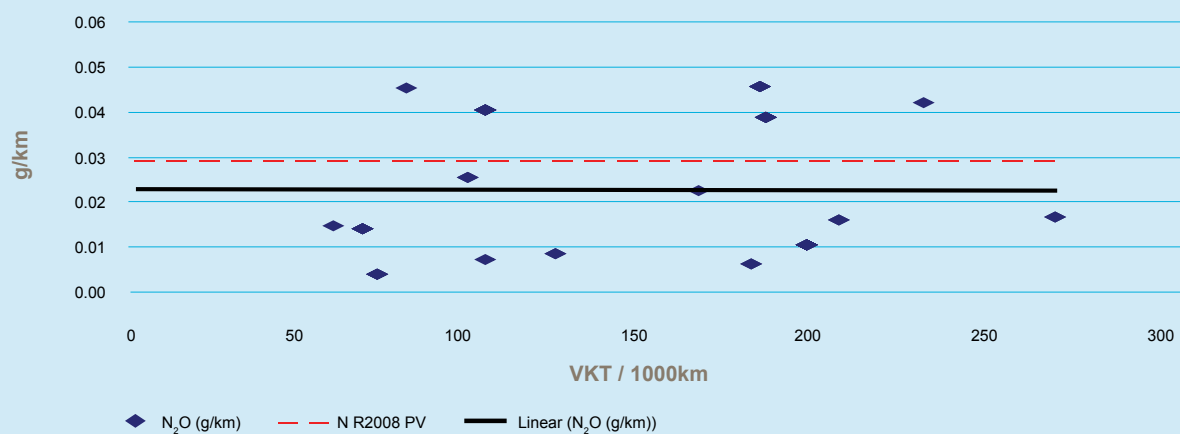
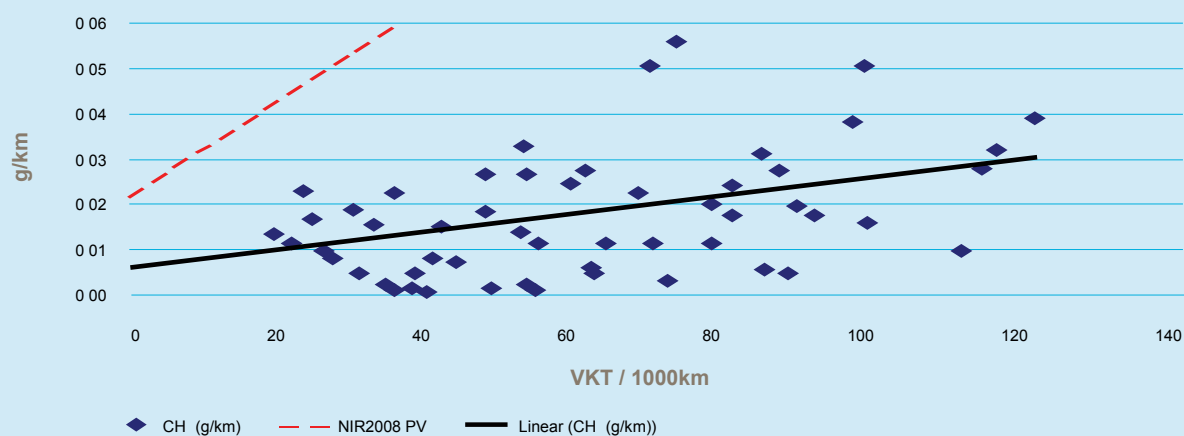
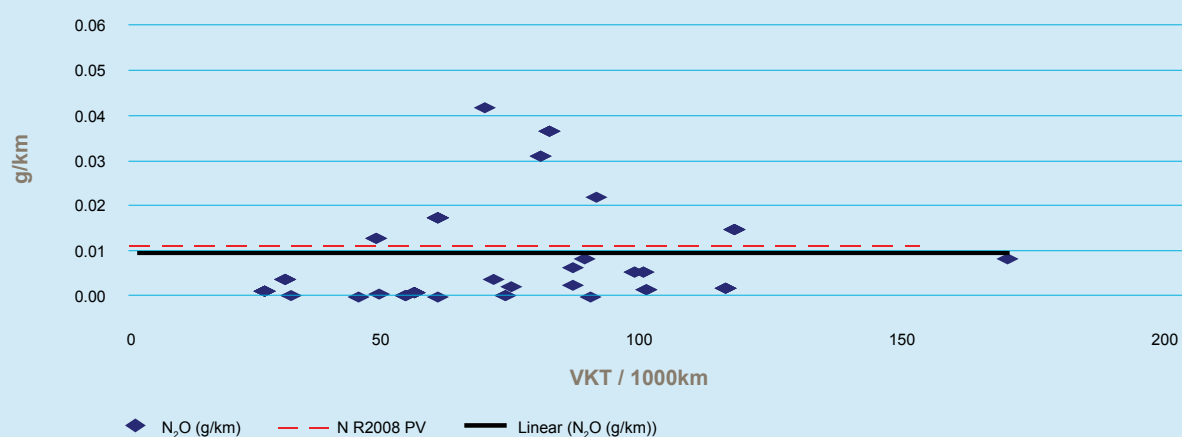
Figure 3.16: ADR37/01 N₂O Hot Urban Estimations – Passenger VehiclesFigure 3.17: ADR79/00 CH₄ Hot Urban Estimations – Passenger VehiclesFigure 3.18: ADR79/00 N₂O Hot Urban Estimations – Passenger Vehicles

Figure 3.19: ADR79/01 CH₄ Hot Urban Estimations – Passenger Vehicles

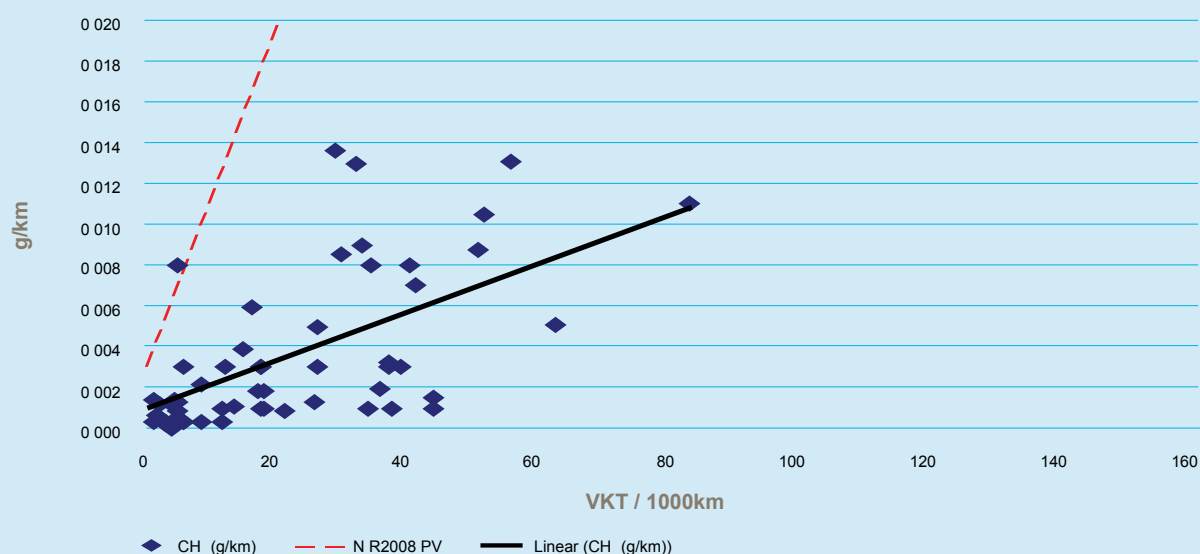
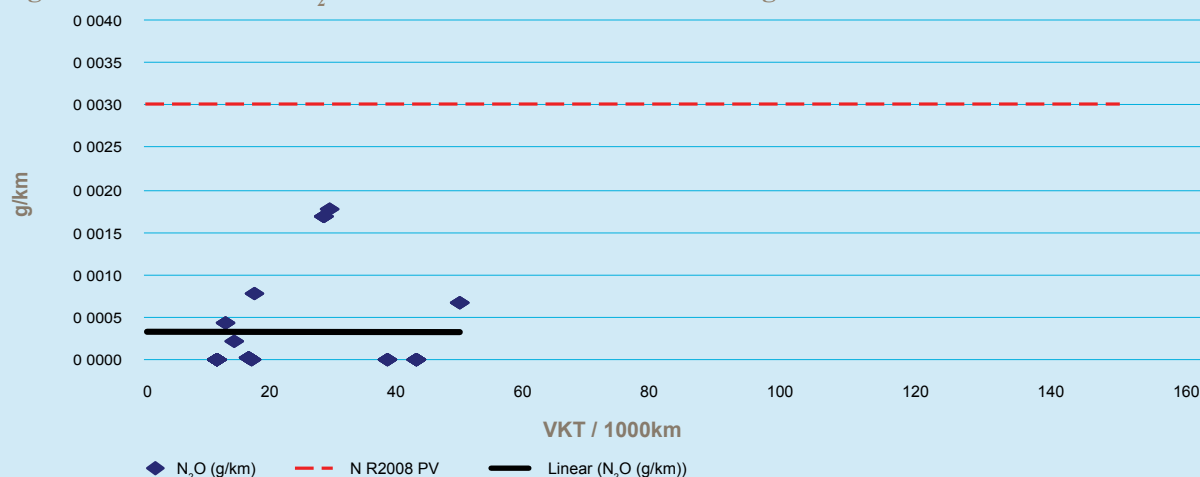


Figure 3.20: ADR79/01 N₂O Hot Urban Estimations – Passenger Vehicles



Future National Inventory Developments

Within the passenger vehicle group the Orbital Australia report identified that the factors associated with the SUV-large vehicle sub-type can vary significantly between specific vehicle make/models depending on the original ADR to which they are certified. These factors are also significantly different to the other vehicle sub-types in the passenger vehicle group. Separate EFs and DRs for SUV-Large were formulated in the report. The DCCEE will investigate further disaggregation of vehicle classifications in the next annual inventory submission.

Separate DRs were estimated for LCVs and passenger vehicles. The Department will amend the model in the next inventory submission to accommodate this further disaggregation.

EFs from the NISE 2 dataset are available for further disaggregated drive cycles (hot extra urban). The DCCEE will investigate whether the required data is available to support the further disaggregation of drive cycles in the National Inventory.

The Orbital Australia report contained separated DRs for LCVs and petrol passenger vehicles. The National Inventory model currently applies one DR to all light duty vehicles of a given vintage. In the next inventory submission DCCEE will expand the inventory model to allow further disaggregations in DRs.

Note: Significant proportions of this box were taken directly from Orbital Australia 2011

For vehicles not covered by the NISE 2 study the choice of US versus European default factors has been dictated by the exhaust emission standards in the Australian Design Rules (ADR) applicable to each particular vehicle vintage. Australian Design Rules have been harmonised with European Standards since 1996 in heavy duty vehicles. Therefore the IPCC default factors used for post 1995 heavy duty vehicles are based on European data (COPERT IV). Prior to the harmonisation with European standards, US Federal Test Protocol standards were used as the basis for ADRs. Therefore USEPA default factors cited in IPCC 2006 are used for earlier vehicle vintages where required.

Australian design rules applied to Australia's vehicle fleet, their date of introduction and the European sources for these standards are outlined in tables 3.19 and 3.20. The age-band structure of the motor vehicle emission model is based on the applicability of a given ADR to a given vehicle vintage.

Table 3.19: Australian petrol passenger car exhaust emission standards

Australian Standard	Year introduced	Source standard
ADR 79/00	2004	Euro 2
ADR 79/01	2006	Euro 3
ADR 79/02	2010	Euro 4

Source: Department of Infrastructure and Transport

Table 3.20: Australian heavy duty diesel exhaust emission standards

Australian Standard	Year introduced	Source standard
ADR 70/00	1996	Euro 1
ADR 80/00	2003	Euro 3
ADR 80/02	2008	Euro 4
ADR 80/03	2011	Euro 5

Source: Department of Infrastructure and Transport

The Orbital Australia 2011 report on NISE 2 data provided factors for light duty vehicles manufactured from 1994 onwards (see Box 3.1). For vehicles manufactured prior to 1994 factors are estimated from the Federal Office of Road Safety (FORS) 1996 – National In-Service Emissions study 1. FORS was a comprehensive testing program which looked specifically at the emissions performance of Australia's passenger car fleet and the effects of tuning and emissions control deterioration. Tests were performed on a total of 640 passenger cars. Emissions measured were limited to hydrocarbons (HC), carbon monoxide (CO) and oxides of nitrogen (NO_x). Methane was not specifically identified in the study. However, it is possible to derive CH₄ emissions from HC emissions using standard assumptions about HC speciation. Methane emission factors by road type are presented in Carnovale 1991. Approximately 10 vehicles of vintage ranging from 1976 to post 85 were tested in this study. As with FORS 1996, methane emissions were not reported directly. The Bureau of Transport and Regional Economics 2002 also provides CH₄ emission factors for a range vehicle age cohorts. These emission factors are based on those presented in Carnovale et al. (1991) and FORS (1996).

As the emission factors cited in the above reports are most applicable to vehicles in the 2-way catalyst and pre catalyst technology categories, Carnovale et al. (1991) emission factors were used for urban (congested) and non-urban (freeway) roads in the pre-1976 and 1976-1985 age cohorts and FORS (1996) emission factors were used for urban roads in the 1985-1997 (2 way catalyst) vehicle class.

For light duty petrol vehicles manufactured from 1994 onwards emission factors are taken from the NISE 2 study (see box 3.1).

There are no country-specific CH₄ emission factors available for heavy-duty vehicles. These emission factors have been taken from USEPA 1989 and IPCC 2006 as indicated in Appendix 3.A.6. Methane emission factors for post-2005 vintage vehicles (Euro 3) have been derived based on the Euro 1 COPERT

IV emission factor and an emission reduction factor over heavy Euro 1 diesel vehicles of 44% taken from Ntziachristos and Kuuridis 2007.

A summary of the emission factors used to estimate CH₄ emissions from the Australian petrol, diesel, LPG and ethanol driven passenger and light commercial vehicle fleets, as well as their respective sources, are presented in appendix Table 3.A.6.

Emissions of non-CO₂ exhaust gases may increase as the vehicle ages due to the gradual wearing of components, poor maintenance, deactivation of catalyst materials, removal of emission control equipment, oxygen sensor failure, or modification of the engine.

For vehicles manufactured prior to 1994 a study by EPA NSW (1995) analysed the combined emission test databases of EPA NSW and EPA Victoria to determine deterioration rates and zero VKT (i.e new car) emissions for the two States' combined fleet. For vehicles manufactured from 1994 onwards the deterioration rates are taken from the Orbital Australia 2011 report on NISE 2 data (see Box 3.1). The general trend is for hydrocarbon emissions to increase as the cumulative VKT increases. The deterioration rates used to derive emission factors for the passenger and light commercial vehicle fleet are shown in appendix Table 3.A.11.

The majority (345 out of 347) of vehicles tested in the NISE 2 study had a VKT between 0 and 300,000km. Most of the deterioration rates used in the transport model are sourced from this data set. Therefore the National Inventory has applied a limit to the emission factors applied after deterioration. This limit is applied at an accumulated average VKT of 300,000km per vehicle.

N₂O emission factors for Australia's petrol-fuelled passenger vehicle fleet are based on CSIRO testing (Weeks et al, 1993) of vehicles of vintage up to 1993, fitted with a range of emissions control technology. Test data on vehicles not fitted with catalysts are used for the pre 1976 and the 1976-85 age groupings and a weighted average of the catalyst equipped emissions used for the 1985-1997 and the post-1997 vehicle fleet. The emission factors in Weeks et al are comparable to those reported in IPCC (2000) and by the US EPA and COPERT IV. N₂O emission factors for light duty petrol vehicles of vintage 1994 onwards are estimated in the Orbital Australia 2011 report on NISE 2 data (see box 3.1).

Australian emissions standards as set out in Australian Design Rules (ADRs) have tended to lag those applied in Europe and the United States – emission standards for petrol passenger vehicles equivalent to Euro 3 (ADR 79/01) were introduced in Australia in 2006. Consequently, the types of emissions control technology employed in Australia also tend to lag as these are introduced in order to comply with the emissions standards.

The emission factors used to estimate N₂O emissions from the Australian petrol, ethanol, diesel and LPG driven passenger and light commercial vehicle fleets, as well as their respective sources, are presented in appendix Table 3.A.8.

Vehicles tested in the NISE 2 study were almost exclusively at 300,000VKT or lower. The emission factor after applying deterioration is assumed to reach its limit at an accumulated average VKT of 300,000km per vehicle.

There are no country-specific N₂O emission factors available for heavy-duty vehicles. These emission factors have been taken from USEPA 1989 and IPCC 2006 as indicated in appendix Table 3.A.9.

Emission factors from the 2006 IPCC guidelines are used in the road transportation sector when they are the most appropriate factors for the vehicle standards and technology that exist in the Australian road transport fleet.

Motorcycles (1.A.3.b.vi)

More generally, the estimation of emissions for motorcycles is given by equations 3.4 and 3.5. Fleet average emission factors for motorcycles are provided in appendix Table 3.A.12.

Evaporative Fuel Emissions (1.A.3.b.v)

Road vehicles using automotive gasoline emit NMVOCs both from the exhaust and through evaporation. The evaporative NMVOC emissions include:

- Running losses resulting from evaporative emissions released during engine operation. Running losses occur when the capacity of the vapour control canister and purge system is exceeded by the vapour generation rate and are greatest at low average vehicle speeds. Running losses vary with the age and type of control system of the vehicle and the trip duration.
- Hot soak losses resulting from evaporation of fuel at the end of each trip. These emissions bear little relation to the VKT for an individual vehicle. A more realistic activity on which to base these emissions is the number of trips an average vehicle would make in a given time period.
- Diurnal losses resulting from vapour being expelled from fuel tanks due to ambient temperature rises. These emissions are strongly dependent on the Reid Vapour Pressure (RVP) of the fuel, the daily ambient temperature changes and where the vehicle is parked during the day. Emissions will vary significantly between identical vehicles in different geographical regions. Diurnal emissions only occur when the temperature is rising.
- Resting losses resulting through the permeation of fuel through rubber hoses or open bottom carbon canisters. Resting losses have often been included in measurements of hot soak, diurnal and running losses (USEPA, 1991a).

Emission factors for evaporative emissions for each of the three passenger vehicle age classes have been estimated for average Australian temperatures and fuel properties and are presented in appendix Table 3.A.19.

Railways (1.A.3c)

Emissions are estimated using Tier 2 methods described by equations 3.1 and 3.2. Carbon dioxide emission factors are reported in Table 3.2 and non-carbon dioxide emission factors are reported in Table 3.21. Given data on the composition and engine types in the local fleet, an average fleet emission factor has been calculated using the individual engine emission factors in USEPA (1992). Data on fuel consumption is taken from ABARE 2010.

Table 3.21: Non-CO₂ emission factors for non-road sources

Source Category	CH ₄	N ₂ O	NO _x	CO	NMVOC
	(g/MJ)				
Rail Transport ^(a)					
ADO	0.003	0.002	1.530	0.202	0.071
IDF	0.003	0.002	1.530	0.202	0.071
Coal	0.032	0.001	0.190	0.220	0.260
Marine Transport ^(b)					
Domestic					
Petrol					
Small Craft	0.360	0.001	0.254	20.300	3.240
ADO	0.004	0.002	1.105	0.246	0.075
IDF	0.007	0.002	1.580	0.163	0.046
Fuel Oil	0.003	0.002	2.000	0.044	0.063
NG	0.243	0.001	0.243	0.095	0.029
Coal	0.032	0.001	0.190	0.220	0.260
International					
ADO	0.007	0.002	1.580	0.163	0.046
IDF	0.007	0.002	1.580	0.163	0.046
Fuel Oil	0.003	0.002	2.000	0.044	0.063

Sources: (a) USEPA (1995a); (b) Lloyd's Register (1995, and previous issue).

Navigation (1.A.3d)

Emissions are estimated using Tier 2 methods described by equations 3.1 and 3.2. Carbon dioxide emission factors are reported in Table 3.2 and non-carbon dioxide emission factors are taken from Lloyds 1995 and are reported in Table 3.21.

Emissions from international bunker fuels are also estimated, but are excluded from national emission inventory aggregates by international agreement. Activity data for international bunkers is estimated by ABARES as part of the Australian Energy Statistics. ABARES utilise data published by the Department of Resources Energy and Tourism (DRET) in their Australian Petroleum Statistics (APS).

The publishes Monthly national and state petroleum statistical information are published in the Australian Petroleum Statistics. Sales of aviation turbine fuel, diesel and fuel oil for domestic and international uses are separated on a quarterly basis.

An explanatory note in the APS outlines that the distinction between international and domestic/coastal fuel consumption data is undertaken according to the predominant mode of usage by the consumer.

3.5.3 Uncertainties and Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas. Time series consistency is ensured by use of consistent models, model parameters and datasets for the calculations of emissions estimates. Where changes to emission factors or methodologies occur, a full time series recalculation is undertaken.

3.5.4 Source Specific QA/QC

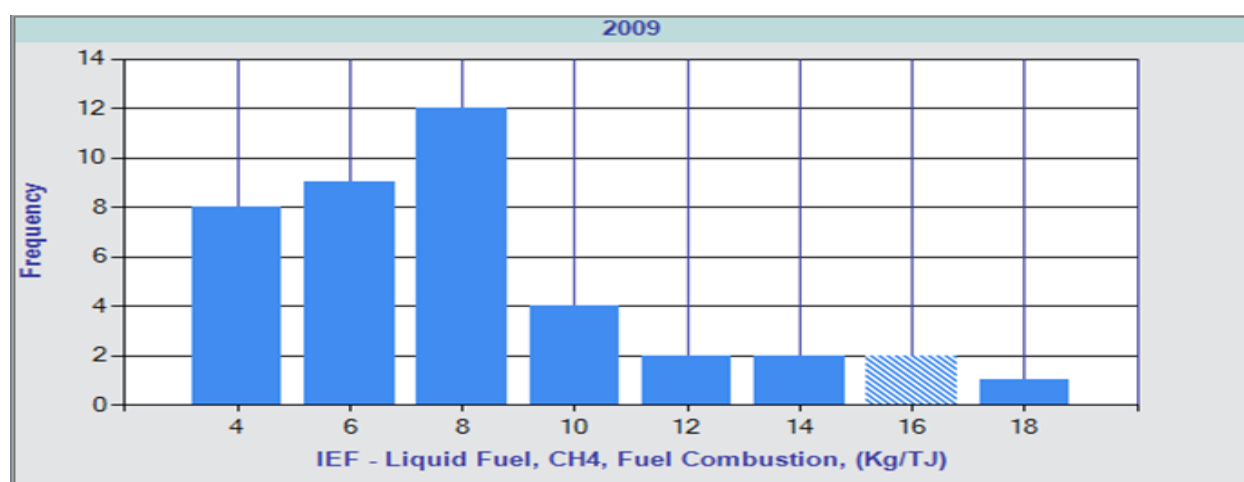
This source category is covered by the general QA/QC of the greenhouse gas inventory in section 1.6 and the fuel combustion specific QA/QC outlined in section 3.2.6.

The primary sources of activity data for this sector are the Australian Bureau of Agricultural and Resource Economics and Sciences (ABARES) and the Australian Bureau of Statistics (ABS). These two organisations have systematic quality assurance programmes in place. In addition, there are also a number of critical user organisations and alternative data sources available for this sector.

Comparisons of implied emission factors and with international data sources are conducted systematically for the Australian inventory. In the previous inventory submission it was found that implied emission factor for methane from the combustion of liquid fuels in Australia (18.1 kg CH₄/TJ) was significantly higher than those of other Annex 1 parties (7.5kg CH₄/TJ). The largest contributor to Australia's high emission factors was CH₄ emissions from road vehicles.

One of the primary purposes of the Orbital Australia report (Box 3.1) was to more accurately reflect the emissions of modern Australian light duty petrol road vehicles (the largest contributor to methane fuel combustion emissions). The outcomes of the study were generally lower CH₄ emission factors and deterioration rates for newer vehicles. The effect of these updated factors was a reduction in the implied emission factor for methane from the combustion of liquid fuels in Australia to 14.3 kg CH₄/TJ. This value is still found to be significantly higher than the Annex 1 population (see Figure 3.21).

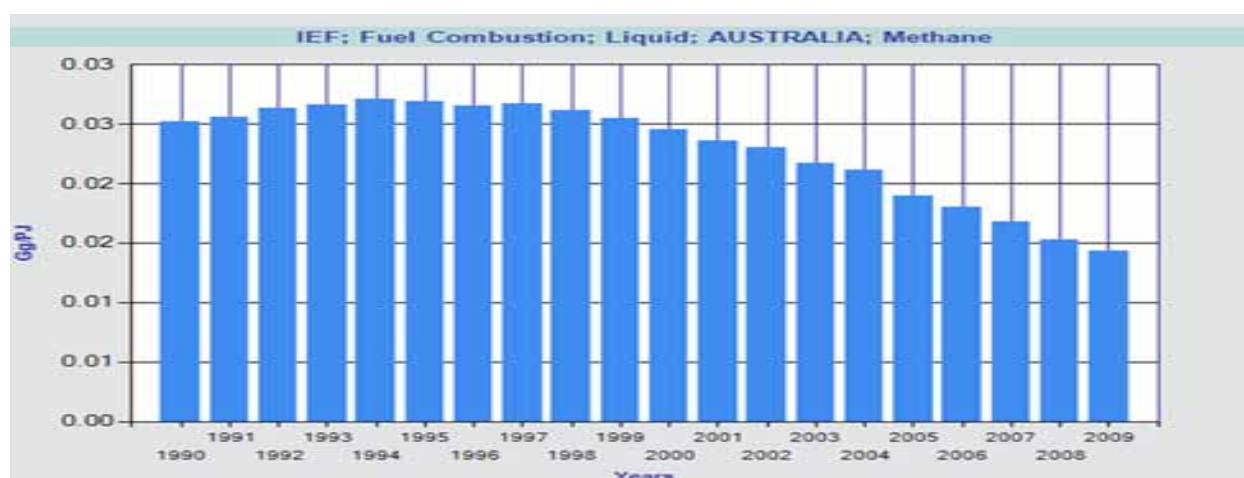
Figure 3.21: Methane implied emission factor from fuel combustion for Annex I countries and Australia



This outcome is likely due to the slow turnover in motor vehicles stocks. Over 16% of passenger vehicles and 25% of light commercial vehicles were manufacture prior to 1994 (the earliest vehicle included in the Orbital Australia analysis). It is expected that Australia's implied emission factor will continue to decrease as these older vehicles are retired from the vehicle fleet. This trend is evident in Figure 3.22 which shows Australia's implied emission factor from fuel combustion over the time series.

Further improvements are also planned for the road transport model as outlined in section 3.5.6.

Figure 3.22: Methane implied emission factor from fuel combustion for Australia



The Department of Infrastructure and Transport has developed a software tool to compute and track the carbon footprint associated with aircraft fuel uplifted in Australia. The Department of Infrastructure and Transport completed an assessment of the robustness of their results by comparing their calculated values with the APS. Their results showed that over the 2008-09 and 2009-10 period computed fuel sales using the software tool and fuel estimates reported by the APS, which is the primary source of aviation fuel consumption data for ABARES, differed by 3.7% for domestic consumption and 2.6% for international consumption. This is concluded to be a strong independent verification of the estimates.

3.5.5 Recalculations Since the 2008 Inventory

The largest recalculations resulted from the updated fuel consumption data from ABARE for 2003-2008 largely due to a reallocation in ADO fuel consumption between the transport and mining sector. Other significant recalculations resulted from the NISE 2 study providing updated emission factors, deterioration rates and deterioration limit for light duty petrol vehicles in the road transport model. Finally updated data on the amount of fuel allocated to military transport resulted in minor recalculations in the transport sector. The effect of the recalculations on 1990 and 2008 is as follows:

- 2010 Submission – 62,092 Gg CO₂-e in 1990, 80,213 Gg CO₂-e in 2008.
- 2011 Submission – 62,121 Gg CO₂-e in 1990, 83,695 Gg CO₂-e in 2008.

Table 3.22: 1.A.3 Transport: recalculation of total CO₂-e emissions, 1990-2008

	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1.A.3.a Civil Aviation (a)				
1990	2,923.5	2,923.5	0.0	0.00
2000	5,012.1	5,010.3	-1.8	-0.04
2001	5,473.0	5,470.5	-2.6	-0.05
2002	4,845.9	4,843.5	-2.4	-0.05
2003	4,725.7	4,723.1	-2.6	-0.05
2004	4,877.5	4,874.7	-2.8	-0.06
2005	5,021.6	5,018.6	-3.1	-0.06
2006	5,164.9	5,161.6	-3.3	-0.06
2007	5,343.4	5,338.8	-4.6	-0.09
2008	5,963.3	5,957.3	-6.0	-0.10
1.A.3.b Road Transportation (b)				
1990	54,330.4	54,359.5	29.2	0.05
2000	65,919.5	66,191.2	271.8	0.41
2001	64,408.7	64,685.1	276.4	0.43
2002	66,680.3	66,977.6	297.3	0.45
2003	67,371.9	68,486.8	1,114.8	1.65
2004	70,790.8	71,971.0	1,180.2	1.67
2005	69,582.9	71,520.7	1,937.7	2.78
2006	69,744.7	71,901.4	2,156.7	3.09
2007	69,419.6	72,053.1	2,633.5	3.79
2008	69,247.0	72,964.4	3,717.4	5.37
1.A.3.c Railways (c)				
1990	1,748.4	1,748.4	0.0	0.00
2000	1,566.4	1,566.4	0.0	0.00
2001	1,492.4	1,492.4	0.0	0.00
2002	1,567.0	1,567.0	0.0	0.00
2003	1,639.2	1,639.2	0.0	0.00
2004	1,815.2	1,815.2	0.0	0.00
2005	1,886.5	1,886.5	0.0	0.00
2006	1,900.8	1,900.8	0.0	0.00
2007	1,935.8	1,935.8	0.0	0.00
2008	2,026.6	2,313.2	286.5	14.14

	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1.A.3.d Navigation (d)				
1990	3,049.5	3,049.5	0.0	0.00
2000	2,363.7	2,376.9	13.2	0.56
2001	2,123.6	2,133.9	10.3	0.49
2002	2,119.8	2,132.2	12.4	0.58
2003	2,115.9	2,128.6	12.7	0.60
2004	2,281.0	2,296.3	15.3	0.67
2005	2,340.3	2,358.6	18.3	0.78
2006	2,225.7	2,248.1	22.4	1.01
2007	2,916.1	3,033.0	116.9	4.01
2008	2,931.9	2,950.0	18.1	0.62
1.A.3.e Other Transportation				
1990	40.4	40.4	0.0	0.00
2000	43.3	43.3	0.0	0.00
2001	42.2	42.2	0.0	0.00
2002	43.3	43.3	0.0	0.00
2003	43.9	43.9	0.0	0.00
2004	46.4	46.4	0.0	0.00
2005	45.0	45.0	0.0	0.00
2006	44.0	44.0	0.0	0.00
2007	44.6	44.6	0.1	0.12
2008	44.4	44.4	0.0	0.09

Recalculation explanation

(a) Improved allocation between sectors

From 1995-2008 the allocation of aviation turbine fuel between 1.A.3.a and 1.A.5.a has improved due to new data on fuel consumption reported by the Department of Defence.

(b) ABARE Australian Energy Statistics

Reallocation of diesel use between the mining and transport sector resulted in a large increase in diesel fuel consumption allocated to the transport sector between 2003 and 2008.

Updated Road Transport Factors

Emission factors, deterioration factors and the deterioration factor limit were updated from data outcomes from the NISE 2 report. These updated factors generally had a negative effect on non-CO₂ emissions in recent years due to the lower calculated emission factors 1990 emissions are higher due to the increase in the deterioration factor limit.

(c) ABARE Australian Energy Statistics

In 2008 ABARE increase their estimate of ADO consumed in this sector.

(d) Improved allocation between sectors

ABARE Australian Energy Statistics

From 1995-2008 the allocation of fuel between 1.A.3.a and 1.A.5.a has improved due to new data on fuel consumption reported by the Department of Defence

In 2007 and 2008 ABARE made some small updates to their estimates of black coal and ADO consumption in this sector.

3.5.6 Planned Improvements

A number of mobile source categories are allocated to the stationary source in the inventory because the current national data collection methods do not allocate this fuel to the transport sector but rather to the specific ANZSIC class in which it is used. A project is planned to analyse the NGERS data to investigate the magnitude of these emissions and whether the reliability, completeness and accuracy of the data is adequate to inform a reallocation of these emissions from the stationary sectors to the transport sector.

The Orbital Australia report (see Box 3.1) provided a level of detailed vehicle testing data that is at a greater level of disaggregation than is currently available in the national inventory model. The DCCEE plans to investigate and apply updates, as appropriate, to the issues listed below in its next inventory submission.

- Within the passenger vehicle group the Orbital Australia report identified that the factors associated with the SUV-large vehicle sub-type can vary significantly between specific vehicle make/models depending on the original ADR to which they are certified. These factors are also significantly different to the other vehicle sub-types in the passenger vehicle group. Separate EFs and DRs for SUV-Large were formulated in the report. The DCCEE will investigate further disaggregation of vehicle classifications in the next annual inventory submission.
- Separate DRs were estimated for LCVs and passenger vehicles. The Department will amend the model in the next inventory submission to accommodate this further disaggregation.
- EFs from the NISE 2 dataset are available for further disaggregated drive cycles (hot extra urban). The DCCEE will investigate whether the required data is available to support the further disaggregation of drive cycles in the National Inventory.
- The Orbital Australia report contained separated DRs for LCVs and petrol passenger vehicles. The National Inventory model currently applies one DR to all light duty vehicles of a given vintage. In the next inventory submission DCCEE will expand the inventory model to allow further disaggregations in DRs.

Update emission factors in the transport model in light of the updated Australian heavy duty diesel exhaust emission standards.

3.6 SOURCE CATEGORY 1.A.4 – OTHER SECTORS

3.6.1 Source Category Description

Source category *1.A.4 other sectors* is an aggregation of the following sources:

- Commercial/Institutional—a diverse category which includes direct emissions from water utilities, accommodation, communications, finance, insurance, property and business services, government and defence, education, health and wholesale and retail trade.
- Residential—emissions from fuel combustion in households, including lawnmowers.
- Agriculture, forestry and fisheries—emissions from fixed and mobile equipment.

ABARES report energy consumption for economic sectors is defined using the Australia New Zealand Standard Industrial Classification (ANZSIC). The mapping of ANZSIC codes against IPCC classifications is complete and given in Table 3.23.

Only the petroleum from ANZSIC sub-divisions 65–67 is included in this category. The natural gas consumption is accounted for within the Other Transport (Natural Gas Transmission) sub-category. Similarly, only the natural gas consumption from sub-category 62 Railway Transport is included in this category. Any other fuel consumption within sub-category 62 is assumed to be accounted for within sector 1.A.3.

3.6.2 Methodology

The methodology for this sector consists of Tier 2 approaches and country specific CO₂ emission factors. Non-CO₂ emission factors have been calculated using a sectoral equipment-weighted average approach.

CO₂ emission and oxidation factors are reported in Tables 3.2 and 3.3. Activity data are taken from the Australian Energy Statistics published by ABARES. Non-CO₂ emission factors for this sector, by ANZSIC Division, are reported in appendix Table 3.A.3.

Table 3.23: Relationship between IPCC source categories and ANZSIC sectors: Other Sectors

IPCC Source Category	ANZSIC Category			
	Division	Sub-division	Group/Class	Description
4. Other Sectors				
A Commercial, Institutional	Division D	37		Water supply, sewerage and drainage services
	Division F	45, 46, 47		Wholesale trade
	Division G	51, 52, 53		Retail trade
	Division H	57		Accommodation, cafes and restaurants
	Division I Transport and Storage	66		Services to transport
	Division I Transport and Storage	67		Storage
	Division J	71		Communication services
	Division K	73, 74, 75		Finance and insurance
	Division L	77, 78		Property and business services
	Division M	81, 82		Government administration and defence
	Division N	84		Education
	Division O	86, 87		Health and community services
	Division P	91, 92, 93		Cultural and recreational services
	Division Q	95, 96		Personal and other services
B Residential	Residential			Residential
C Agriculture, forestry, and fishing	Division A	01		Agriculture
	Division A	02		Services to agriculture, hunting and trapping
	Division A	03		Forestry and logging
	Division A	04		Commercial fishing

Table 3.24: Summary of methods and emission factors: 1.A.4 Other Sectors

Source Category	CO ₂		CH ₄		N ₂ O	
	Method applied	Emission factor	Method applied	Emission factor	Method applied	Emission factor
1A4a Commercial/Institutional	T2	CS	T2	CS	T2	CS
1A4b Residential	T2	CS	T2	CS	T2	CS
1A4c Agriculture, Forestry and Fisheries	T2	CS	T2	CS	T2	CS

Notes: T1 = Tier 1, T2 = Tier 2, T3 = Tier 3, CS= Country-specific.

Residential – Biomass Combustion (1.A.4)

The *Residential* sector also includes specific treatment of the use of firewood and also in the combustion of fuels in mobile equipment such as lawnmowers.

This category is characterised by the use of wood in residential woodheaters. Emissions are modelled using an advanced tier 2 approach which takes into account factors such as wood heater technology and replacement of older models, user operation and Australian wood.

The estimation of emissions from residential firewood use requires a more complex approach to the estimation of emissions from fossil fuels reflecting information on heater design (technology type) and the operation of wood-burning appliances, which influences the mix of emissions per kilogram of firewood consumed.

The proportion of Australian households choosing firewood as their main heating fuel peaked in the early 1990s and has decreased slowly since then. New appliances, with lower emissions of some greenhouse gas species, came on the market in the early 1990s and they have gradually been replacing older, non-certified heater models. Poor user behaviour, which significantly increases emissions of pollutants, has been the target of education campaigns and, in the past few years, programs have been aimed specifically at households with excessive visible smoke. This has led to improved appliance use.

The residential wood heater methodology has been developed for Australian conditions (see Todd 2003, and updated with Todd 2005), and incorporates factors such as appliance type and certification, wood type and moisture content and user behaviour. The composition of gaseous and particulate emissions when burning eucalypt firewood in typical Australian appliances is based on Gras (2002). A schematic diagram showing the methodology process is shown in Figure 3.23, and is also summarised in the algorithm below:

$$E_{k,n} = F_n \times S \times W \times fn_k \{ \sum PEF_n \} \dots\dots\dots 3.10$$

Where: $E_{k,n}$ = emission of greenhouse gas k in year n
 F_n = amount of fuel combusted (i.e. firewood use) in year n
 S = softwood use correction factor
 W = wet wood correction factor
 fn_k = formula linking the greenhouse gas emission factor for gas k to the particulate emission factor.
 PEF_n = weighted particulate emission factor for year n, which is summed over the mix of appliances and operator behaviour for that year, with $l = 1$ to 8
 $l(1)$ certified woodheater correctly operated
 $l(2)$ certified woodheater carelessly operated
 $l(3)$ certified woodheater very badly operated
 $l(4)$ non-certified woodheater correctly operated
 $l(5)$ non-certified woodheater carelessly operated
 $l(6)$ non-certified woodheater very badly operated
 $l(7)$ masonry open fireplace
 $l(8)$ factory built (metal) open fireplace

Description of factors

Certified and non-certified heater

- Emission factors

A base methane emission factor for certified woodheaters of 261.3 Mg/PJ has been developed by Todd (2005). It has been derived from a large database on particulate emissions from heaters meeting the requirements of Australian Standard AS4013. Over 250 different heater models have been tested at the two NATA certified (National Association of Testing Authorities) laboratories in Australia, producing a database of over 2250 individual emission tests (heaters must have three repeat tests at each of high, medium and low burn rates).

A base methane emission factor of 461.3 Mg/PJ has been applied to non-certified heaters, through the application of a factor of 1.77 to the certified woodheater emission factor. Todd (2005) based this approach on comparisons between US emission tests of non-certified heaters (referred to as 'Pre-Phase I Non-Catalytic Heaters' in US literature) and certified heaters (referred to as Phase II Non-Catalytic Heaters) (USEPA 1996). The Australian emission test for woodheaters has differences to the US test (both in test fuel, and testing procedure); however, the Australian Standard was cross-checked with two models of heater that had passed both the US (Phase II) and found to be generally similar. Thus the US ratio has been applied to Australian heaters.

- Mix of certified and non-certified heaters and open fireplaces

A survey of households in 2000, carried out as part of a CSIRO study (Gras, 2002), found that 40% of heaters were less than 6 years old (i.e. installed in 1994 or later). Taking into account the number of open fireplaces also in use (derived by Todd 2005 from a 1999 ABS survey), certified woodheaters accounted for 30.6% of all wood-burning appliances in 2000. The population of certified woodheaters has been decreased linearly to 1994, where it is zero (Todd 2005).

Operator behaviour

- Emission factors

Three operator classifications have been adopted for these calculations.

- a) 'Good' operation means a certified heater will perform as it did in the laboratory test.
- b) 'Careless' operation (or poor operation) refers to operators who pay some attention to heater performance, but are not well enough informed. A survey in Tasmania (Todd 2001) suggested at least half the heater owners fall into this category. Careless operation has been assigned emission factors 1.5 times greater than for good operators, applying to both certified and non-certified heaters (expert judgement by Todd 2005).
- c) 'Very poor' operation refers to heater operators that regularly run the heater with a slow, smouldering fire. Todd (2001) indicates 10% of households with woodheaters are in this category. The increase in emissions compared to a well-operated heater has been set at a factor of 4 based on a small number of laboratory tests (Todd 2005).

Proportion of well/poorly operated woodheaters

The proportion of good, careless and very poor woodheater operators for 2000 was set by Todd (2005) at 0.4/0.5/0.1 respectively. This is based on surveys in 1999 and 1997 that showed most households thought they operated their heaters correctly, but more detailed questioning showed that few did everything correctly. National TV campaigns (in 1997 'Breathe the Benefits') and a wide range of other education campaigns at state level suggest user behaviour has improved over time. Therefore Todd (2005) has used 0.7 (i.e. 70%) for 1990 as the proportion of heaters used carelessly.

The very poor operation grouping represents those heaters that regularly emit copious quantities of visible smoke. A 1999 Hobart survey, and feedback from local government officers involved in wood-smoke reduction programs in all states, suggests that about 10% of chimneys/flues smoke excessively. Todd (2005) has allowed for a continuous improvement over the time series, setting 1990 at 0.2, i.e. (20%) of heaters smoked excessively.

Open fireplaces

- Emission factors

No emission testing of masonry open fireplaces has been carried out in Australia. The US (US EPA 1996) value for the particulate emission factor for masonry open fireplaces (17.3g/kg) has been used by Todd (2005) to derive a base methane emission factor of 1365.8 Mg/PJ. Even though the wood species used in Australia are different from the US, this is unlikely to have a significant effect on emission factors. The CSIRO tests provide particulate emission factor of 2.3g/kg for factory-built open fireplace (sometimes referred to as heat-recovery fireplaces). This is used by Todd (2005) to derive a base methane emission factor of 181.6 Mg/PJ. It is assumed that the operator of an open fireplace has little impact on the emissions (on average) and so no correction factors for careless or very poor operation have been used (Todd 2005).

- Proportion of open fireplaces

The proportion of open fireplaces in use is based on the same CSIRO survey and a 1999 ABS survey (Todd 2005).

Softwood fuel and wet wood

- Emission factors

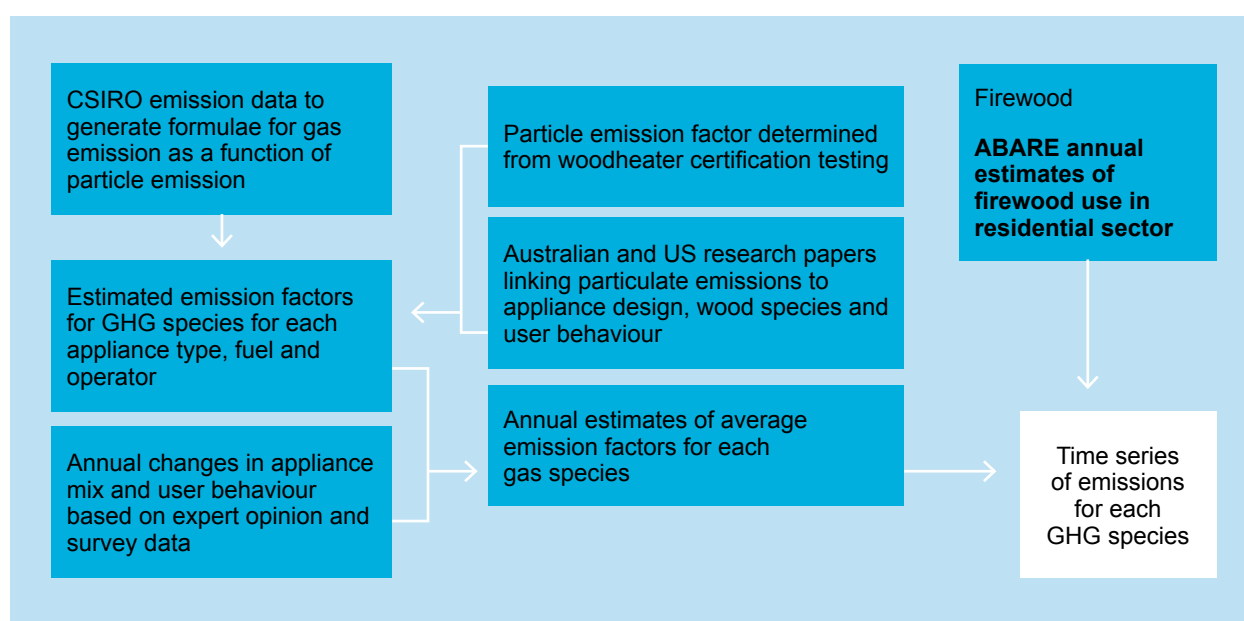
The use of wet firewood is often cited as one of the main reasons for high emissions from woodheaters. However, the CSIRO study, and other Australian studies (e.g. Todd et al. 1989a) have consistently shown that only very wet wood (i.e. unseasoned) influences emissions. High burn-rate tests carried out by the CSIRO have shown that very wet wood (moisture greater than 30%) leads to an increase in emissions by a factor of 3.5 (Todd 2005).

The use of softwood fuel in the CSIRO testing led to a large increase in emissions (by a factor of about 3.5). However, other comparative tests of hardwood and softwood emissions (Todd 1991) have shown smaller increases. Therefore, Todd (2005) has adopted a factor of 2.

- Proportion of wet wood and softwood

The 4% proportion of households using very wet wood (>30% moisture, wet weight basis) is taken from a Hobart survey (Todd 2001). The proportion of softwood used as firewood is based on several surveys (Todd et al. 1989b, Driscoll et al. 2000, Gras 2002) that consistently show around 5% of firewood consumed is softwood.

Figure 3.23: Schematic diagram of the methodology process for estimation of emissions from woodheaters



The resulting emissions factor trends are shown below in Table 3.25. With Australian standards for woodheater emissions introduced in 1992, there has been an increasing uptake of certified heaters at the expense of older, non-compliant heaters, as well as open fireplaces. Together with improving user operation, these factors work to produce an overall trend for the more complete and efficient combustion of fuelwood. This is borne out in the increasing CO₂ emission factor (i.e. more carbon is oxidised under improved combustion conditions) and decreasing methane emission factor. As a result, the implied methane emission factor varies between 1100 Mg/PJ in 1990 and 714 Mg/PJ in 2007. This range is consistent with the 2006 IPCC defaults for residential methane emission factors for woodstoves (IPCC 2006, Vol 2, Table 2.9), taking in account the inherent uncertainty of residential combustion methane emission factors of 50 to 150% (IPCC 2006, Vol 2, Table 2.12).

Table 3.25: Residential biomass emission factors

Inventory Year	Greenhouse Gas Emission Factor (Mg/PJ)						
	CO ₂	CH ₄	N ₂ O	CO	NO _x	NMVOC	SO ₂
1990	70 374	1100.1	2.3	11751.6	16.9	1393.5	1.1
2000	76 245	782.4	1.9	9421.7	21.1	991.0	1.1
2001	76 561	765.2	1.9	9296.0	21.4	969.3	1.1
2002	76 877	748.2	1.9	9170.7	21.6	947.7	1.1
2003	77 191	731.1	1.9	9045.9	21.8	926.1	1.1
2004	77 505	714.2	1.9	8921.6	22.1	904.6	1.1
2005	77 505	714.2	1.9	8921.6	22.1	904.6	1.1
2006	77 505	714.2	1.9	8921.6	22.1	904.6	1.1
2007	77 505	714.2	1.9	8921.6	22.1	904.6	1.1
2008	77 505	714.2	1.9	8921.6	22.1	904.6	1.1
2009	77 505	714.2	1.9	8921.6	22.1	904.6	1.1

Emissions from lawnmowers are estimated using Tier 2 methods described by equation (3.1). Carbon dioxide emission factors are reported in Table 3.2 and non-carbon dioxide emission factors are reported in Table 3.26. There are no fuel consumption statistics for these activities, instead allocation factors are used to derive this data from known consumption statistics. Lawn mowers are powered by small 2-stroke or 4-stroke engines and assumed to be utilised in the ratio of 60:40 (EPA NSW, 1995).

For the *1.A.4.c agriculture, forestry and fisheries* category, ABARES statistics present a single total figure for diesel fuel consumed in agriculture, fisheries and forestry. However, the types of equipment used by these industries vary quite widely (tractors, log skidders, fishing boats etc.), and therefore emission factors for non-CO₂ gases also vary widely. It is assumed that the agriculture, fisheries and forestry industries account respectively for 77%, 6% and 17% of total diesel fuel consumption by the sector as a whole. This estimate is based on the relative volumes of diesel fuel for which excise rebates were claimed, as advised by the Australian Customs Service, over the period 1988 to 1994 inclusive, and have been held constant throughout the period.

These ratios were applied to emission factors for the different types of diesel engines used in the types of equipment typical of the three sectors, to estimate weighted sectoral emission factors (Table 3.26).

Table 3.26: Non-CO₂ emission factors for non-road mobile sources

	CH ₄	N ₂ O	NO _x	CO	NMVOC
	(g/MJ)				
Other Mobile Sources					
Recreational Vehicles					
Petrol	0.03	0.0009	0.37	7	1.08
Industrial Equipment					
ADO	0.0057	0.002	1.006	0.39	0.108
LPG	0.022	0.001	0.437	5.465	0.409
Farm Equipment					
ADO	0.01	0.002	1.36	0.541	0.189
Tractors	0.0096	0.002	1.362	0.543	0.183
Non-Tractors	0.011	0.002	1.351	0.531	0.21
Utility Engines					
Petrol	0.38	0.0009	0.087	13	3.45

Sources: IPCC (1997), USEPA (1995a), F.Carnovale personal communication, 1995.

3.6.3 Uncertainties and Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas.

Time series variability of GHG implied emission factors are likely to be influenced by changes in fuel mix within categories, and changes of actual fuel emission factors.

3.6.4 Source Specific QA/QC

This source category is covered by the general QA/QC of the greenhouse gas inventory in section 1.6.

3.6.5 Recalculations Since the 2008 Inventory

Recalculations to *1.A.4 other* are detailed at the sub-category level in Table 3.27. The majority of the recalculations were made in response to changes to the ABARES national energy statistics. These changes were the result of ABARES revising estimates in response to improved data and reallocating fuel use at the sub-category level to improve time series consistency.

Table 3.27: 1.A.4 Other sectors: recalculation of total CO₂-e emissions, 1990-2008

	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1.A.4.a Commercial/institutional ^(a)				
1990	3,564.17	3,564.10	-0.1	0.00
2000	4,489.25	4,489.54	0.3	0.01
2001	4,209.70	4,210.04	0.3	0.01
2002	4,347.93	4,348.34	0.4	0.01
2003	4,290.11	4,290.50	0.4	0.01
2004	4,339.74	4,339.87	0.1	0.00
2005	4,420.74	4,420.96	0.2	0.01
2006	4,615.60	4,615.83	0.2	0.01
2007	4,658.17	4,651.81	-6.4	-0.14
2008	4,769.19	4,690.13	-79.1	-1.66
1.A.4.b Residential ^(b)				
1990	7,790.23	7,790.87	0.6	0.01
2000	8,774.50	8,775.44	0.9	0.01
2001	8,860.42	8,861.21	0.8	0.01
2002	8,773.58	8,774.33	0.8	0.01
2003	8,964.73	8,938.11	-26.6	-0.30
2004	8,783.28	8,723.78	-59.5	-0.68
2005	8,731.01	8,665.10	-65.9	-0.75
2006	8,840.37	8,780.45	-59.9	-0.68
2007	8,918.99	9,051.16	132.2	1.48
2008	9,055.52	9,199.55	144.0	1.59

	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1.A.4.c Agriculture/fisheries/forestry^(c)				
1990	3,389.66	3,389.61	0.0	0.00
2000	4,386.34	4,386.23	-0.1	0.00
2001	5,380.97	5,380.79	-0.2	0.00
2002	5,463.23	5,463.04	-0.2	0.00
2003	6,855.80	6,083.41	-772.4	-11.27
2004	6,089.34	6,089.14	-0.2	0.00
2005	6,429.24	6,429.04	-0.2	0.00
2006	5,967.90	5,967.70	-0.2	0.00
2007	5,877.21	5,877.01	-0.2	0.00
2008	5,939.57	5,939.37	-0.2	0.00

Recalculation explanation

(a) Minor revision of activity data

Minor adjustment to fuel oil activity data as a result of QC processes

(b) ABARE Australian Energy Statistics

Minor revisions to LPG and town gas activity data

(c) ABARE Australian Energy Statistics

Reallocation of diesel use to the Transport sector for 2003.

3.6.6 Planned Improvements

Some of the country specific emissions factors for liquid fuels used in the national inventory were estimated more than ten years ago. A study is planned to assess these emission factors and to assess whether any updates are required.

A study is planned to update factors in the model used to estimate residential firewood combustion to incorporate the latest available data.

3.7 SOURCE CATEGORY 1.A.5 – OTHER (NOT ELSEWHERE CLASSIFIED)

Emissions from *1.A.5 other* are estimated using a mix of Tier 1 and Tier 2 approaches using emission factors set out in Tables 3.2 and 3.3.

Table 3.28: Summary of methods and emission factors: Other (Not Elsewhere Classified)

Category	CO ₂		CH ₄		N ₂ O	
	Method applied	Emission factor	Method applied	Emission factor	Method applied	Emission factor
1A5 Other	T1/T2	CS	T1/T2	CS	T1/T2	CS

Notes: T1 = Tier 1, T2 = Tier2, CS= Country-specific

3.7.1 Source Category Description

The source category *1.A.5 other* consists of emissions arising from the oxidation of lubricants, as well as fuel used in mobile equipment within defence operations.

3.7.2 Methodology

This sector includes emissions of CO₂ arising from the oxidation of lubricants. Lubricants, together with bitumen and solvents, are non-fuel products of crude oil, which are included in the energy statistics compiled by ABARE 2009a. It is assumed that 60% of lubricants are not oxidised during engine operation, i.e. not actually combusted (Australian Institute of Petroleum, pers. comm.). Therefore the stated ABARES consumption of lubricants and greases is reduced by 60% before emissions are estimated. Emissions of gases other than CO₂ are included with the emissions arising from fuel combustion in the engine type concerned in the relevant sector. Some lubricants may be incinerated subsequent to use. Any emissions from this source are included in the Waste sector.

Emissions from military vehicles are estimated using Tier 1 methods described by equation 3.3 and 3.4. Carbon dioxide emission factors are reported in Table 3.2 and non-carbon dioxide emission factors are reported in appendix Table 3.A.12.

The allocations of fuel to military transport have been improved in this submission as a result of direct reporting of fuel consumption by the Department of Defence. This new data has facilitated an update of the proportions of each fuel allocated to sector 1.A.5. Allocations were made in 2008 and 2009 based on direct data from the Department of Defence. Allocations were updated for 1995-2007 by linearly extrapolating the proportions between the previous reported data point (1994) and the latest reported data (2008). It is anticipated that these proportions will be updated on an annual basis in the future. The shares used to allocate fuel consumption are reported in table 3.A.13.

It is assumed that solvents are either emitted as NMVOC, in which case they are reported in the *solvents* sector, or as CO₂, following incineration, in which case they are reported in the *waste* sector. It is assumed that all fossil carbon in bitumen is sequestered.

3.7.3 Uncertainties and Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas.

3.7.4 Source Specific QA/QC

This source category is covered by the general QA/QC of the greenhouse gas inventory in section 1.6.

3.7.5 Recalculations Since the 2008 Inventory

Recalculations made to *1.A.5 other* are detailed at the sub-category level in Table 3.29. The changes are due to a reallocation of emissions between the transport sector and military transport. This reallocation affected 1995-2008 and is due to new data reported directly by the Department of Defence.

Table 3.29: 1.A.5 Other: recalculation of total CO₂-e emissions (Gg), 1990-2008

	1990	2000	2001	2002	2003	2004	2005	2006	2007	2008
1.A.5.a Stationary – lubricants										
2010 submission (Gg CO ₂ -e)	560	568	588	599	616	668	707	487	454	471
2011 submission (Gg CO ₂ -e)	561	569	589	600	617	670	709	488	455	472
Change (Gg CO ₂ -e)	0.9	0.9	1.0	1.0	1.0	1.1	1.2	0.8	0.7	0.8
Percent Change	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2
1.A.5.b Mobile – Military transport (a)										
2010 submission (Gg CO ₂ -e)	453	680	675	630	612	637	660	684	1,106	1,162
2011 submission (Gg CO ₂ -e)	453	636	631	577	555	570	585	600	932	975
Change (Gg CO ₂ -e)	0	-44	-44	-54	-57	-67	-75			
Percent Change	0	-6.4	-6.6	-8.5	-9.3	-10.6	-11.3			

Recalculation explanation:

(a) From 1995-2008 the allocation of fuel between 1.A.3.a and 1.A.5.b has improved due to new data on fuel consumption reported by the Department of Defence.

3.7.6 Planned Improvements

All relevant data are kept under constant review.

3.8 SOURCE CATEGORY 1.B.1 – SOLID FUELS

3.8.1 Source Category Description

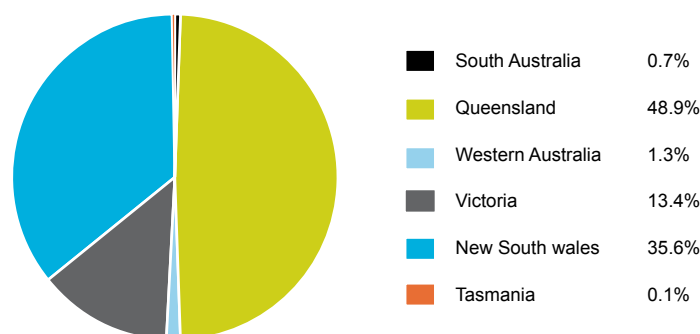
This source category covers fugitive emissions from the production, transport and handling of coal, and emissions from decommissioned mines. It does not include emissions arising from the conversion of coal into coke. Coverage of emissions for 1.B.1 Solid Fuel emission categories, by gas, are shown in Table 3.30. Carbon dioxide emissions are reported for underground coal mines. Estimates for carbon dioxide emissions from surface mines and decommissioned mines are not currently available but will be considered for reporting in the inventory as data becomes available under NGRS.

Table 3.30: B.1 Solid Fuels – Emissions source coverage

Category	CH ₄ emissions	CO ₂ emissions
1.B.1.a.i Underground coal mines		
Mining	YES	YES
Post-mining	YES	
1.B.1.a.ii Surface mines		
Mining	YES	Possible under NGRS
1.B.1.b Solid fuel transformation	IE (IP Metals)	
1.B.1.c Other Decommissioned mines	YES	Possible under NGRS

The great majority of Australia's resources and production of black coal are located on the east coast of Australia in New South Wales and Queensland. A very small quantity of bituminous black coal is also mined in Tasmania. In Victoria, large quantities of brown coal are mined in open cut operations, of which the CH₄ content of this coal is regarded as very low. Sub-bituminous coal is mined in Western Australia, while a low rank sub-bituminous coal is mined in South Australia. The share of coal production from Australian states for 2009 is shown in Figure 3.24.

Figure 3.24: Share of coal production from Australian states – 2009



In New South Wales, the principal coal fields are the Southern, Newcastle, Hunter and the Western New South Wales. In Queensland, they can be identified as the Northern Bowen Basin, the Central Bowen Basin and the Southern Basin. Since 1990 there has been strong growth in production from the Hunter and Bowen Basins and declines from the Southern and Newcastle Basins (see Figure 3.26).

There can be wide variations in both the gas content and the composition of the gas across Australian coal basins and across coal fields within the basins. The variability and characteristics of coal gas in eastern Australia have been described by Thomson (2010) as a response to a number of distinct geological and biogenic processes namely:

- the coalification processes;
- tectonic history;
- magmatic activity;
- groundwater flow; and
- biogenesis.

In Australia, the methane in coal layers has its origins in the coalification process following the initial burial of biomass. Burial of biomass reached a peak during the mid cretaceous period when it was estimated to be around 2.5 to 4 km depth resulting in coal layers reaching saturation with thermogenic methane. As gas is generated during the coalification process, coal is able to store the gas within its micropore structure. The quantity of gas able to be adsorbed within the coal increases with pressure and decreases temperature. Temperature and pressure both increase with depth. Therefore, the upper limit of gas able to be held within coal will follow an adsorption isotherm, which describes the pressure/temperature relationship at the point where the coal is fully saturated with gas. The isotherm is useful for representing a theoretical cap on the gas content of coal at any given depth. In the Permian coal basins of Australia's east coast, coal layers greater than 500 -600m in depth will tend to be close to saturation with thermogenic methane (Thomson 2010).

It is rare, however, for coals saturated with methane to be mined. This is because uplifting and rifting of the strata in geological periods following the coalification process provided opportunities for gas to escape through fracture systems and resulting in the upper coal layers becoming undersaturated with methane. For Australia, this started from the late cretaceous period with New Zealand rifting away from the Australia east coast, with the associated uplifting and subsequent erosion of the coal bearing regions.

The undersaturated coal layers were then receptive to new sources of gas. Extensive magmatism activity in the tertiary period introduced carbon dioxide into the upper, undersaturated coal layers. In more recent times, methanogen bearing groundwater flows through the surface fracture system since the late Tertiary period has introduced biogenic methane into the upper coal layers (Thomson 2010).

A generalised model to describe the variation of gas in coal along the east coast coal bearing regions as a result of these processes has been described by Thomson (2010) and is shown in Figure 3.25. Localised geological features can also have a large influence on subsurface gas characteristics at a mine level scale, for example faults and dykes can provide opportunities for gas to escape or be trapped and influence

groundwater flows for biogenesis. In summary, the coal gas type and distribution characteristics of the eastern coalfields can be viewed as a result of the history of large scale processes overlaying localised geological features. Most near surface coal deposits on the east coast are undersaturated, as a function of their geological history.

Coal mining on the west coast of Australia is confined within a small coal field within the Collie basin. The Collie basin coal deposits were formed by the transport of material rather than the bed forming in situ. The coal beds are also commonly associated with a sandstone roof providing opportunities for gas to escape over time. The understanding of the geological characteristics, current and historical mining practices and anecdotal evidence suggest the basin is characterised by low methane content. The lack of gas has been to an extent that it has not been a mining safety issue, which has resulted in a lack of historical data and analysis for greenhouse gas emission assessment. Work is continuing within the ACARP methodology research framework in order to develop low cost emission estimation methods for low gas mines such as those in the Collie basin. It is also expected that data obtained from mining operational practices and NGERs will provide a better understanding of Collie Basin gas characteristics.

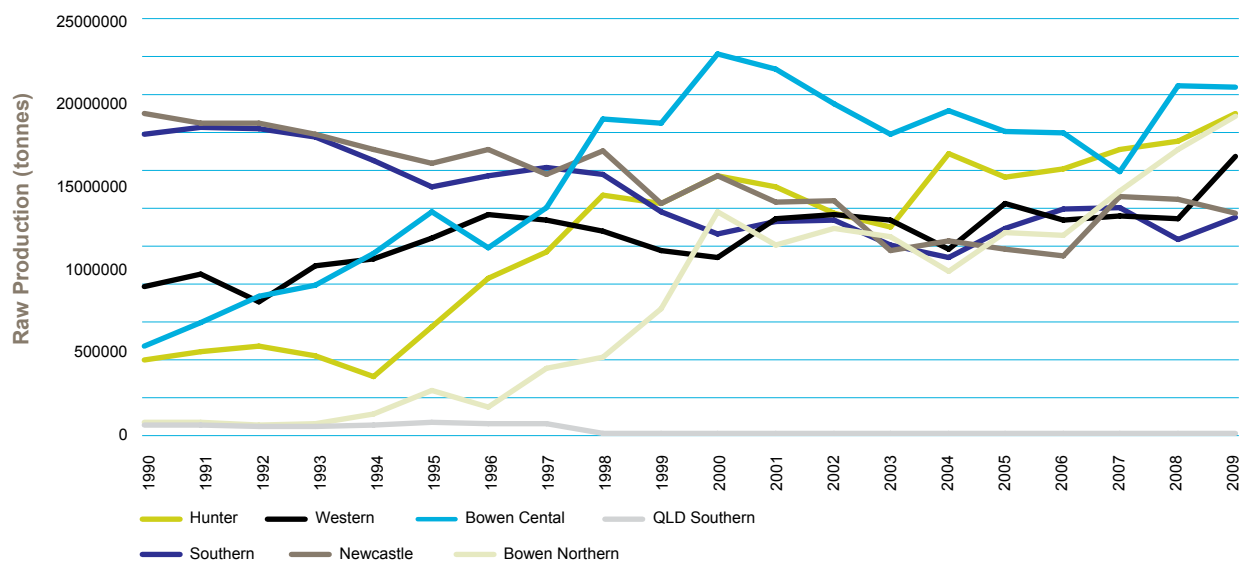
Figure 3.25: Generalised model of gas variation in the subsurface for east coast Australia

Zone 1	Surface zone to ~ 100m of very low gas – CO ₂ dominant
Zone 2	Biogenic zone, 100 to 250/300m. Methane increasing with depth
Zone 3	Mixed gas zone. Biogenic and thermogenic undersaturated CH ₄ . Magmatic CO ₂ present
Zone 4	Thermogenic methane, increasing to saturation with depth

Source: Thomson (2010)

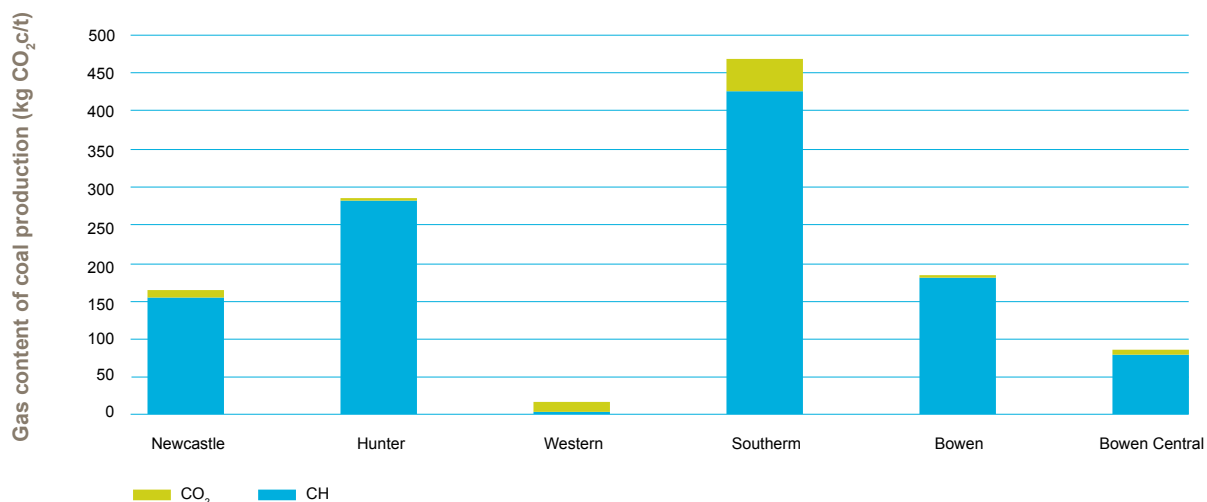
In 2009, there were 43 underground mines and 84 open cut mines operating nationally while emissions are estimated for 112 decommissioned mines. The extensive use of NGERs data for underground mine emissions which is based largely on measured mine ventilation emissions shows that in some basins, at some depths, over half of the gas contained in the coal seams may be carbon dioxide on a volume basis. Carbon dioxide accounts for around five per cent of fugitive emissions estimated on a carbon dioxide equivalent mass basis. The gas type and content of different coal fields varies significantly, as shown in Figures 3.26 and 3.27, which details the average gas content profile of underground production by coal field. The gassiest coal field is the Southern New South Wales while the least gassy field is the Western New South Wales, which is mainly carbon dioxide.

Figure 3.26: Underground black coal production by coal field



Source: Coal Services Pty Ltd and the Queensland Department of Mines and Energy.

Figure 3.27: The gas content profile of Australian underground production by coal field



Source: NGERS data.

3.8.2 Methodology

Fugitive emissions from coal mining activities are estimated using a mix of tier 3 and tier 2 methods. Estimates for underground mines are prepared using a tier 3 method. Data on measured methane emissions for individual mines are obtained from coal mining companies reporting under the National Greenhouse and Energy Reporting System (NGERS). For the 2009 year, data on measured methane emissions was available for 38 underground mines (over 90% of all underground mines).

For open cut and decommissioned mines a country-specific tier 2 approach is used with emission factors (m³ CH₄/tonne coal produced) derived from measurement data obtained for mines with similar characteristics. Time series consistency has been maintained for the underground mine emissions estimates following the introduction of NGERS data (see section 3.8.3).

Table 3.31: Summary of methods and emission factors: 1.B.1 Solid Fuels

GREENHOUSE GAS SOURCE AND SINK CATEGORIES	CO ₂		CH ₄		N ₂ O	
	Method Applied	Emission Factor	Method Applied	Emission Factor	Method Applied	Emission Factor
1. Underground mines	T3	PS	T3, T2	PS, CS	NA	NA
2. Open Cut mines	NA	NA	T2	CS	NA	NA
3. Post Mining emissions	NA	NA	T2	CS	NA	NA
4. Decommissioned Mines	NA	NA	T2/T3	CS	NA	NA

Notes: T2 = Tier2, T3 = Tier3, CS = Country-specific, PS = Plant-specific.

Activity data

Data on coal production provides activity data for the sector and are used as drivers for the estimation of emissions from mines in years where directly measured emissions data is not available. The production data for each mine are published annually in the statistical publications of:

- NSW – Coal Services Pty Ltd (formerly the Joint Coal Board)
- Queensland – Queensland Department of Mines and Energy
- Western Australia – WA Department of Industry and Resources
- Tasmania – Mineral Resources Tasmania
- South Australia – Primary Industries and Resources South Australia
- Victoria – VIC Department of Primary Industries

Underground Mining (1.B.1ai)

Emissions derived from direct measurement account for the majority of emissions from underground mines reported in the inventory. Emissions are estimated using methods set out in the *National Greenhouse and Energy (Measurement) Determination 2008*. In addition, mines are subject to State Government legislation including the *Coal Mine Health and Safety Act 2002 (NSW)* and the *Coal Mine Health and Safety Regulation 2006 (NSW)*; *Coal Mining Safety and Health Act 1999 (Qld)* and the *Coal Mining Safety and Health Regulation 2001 (Qld)* which establish mandatory monitoring and reporting regulations for mines. The *NGER Measurement Determination* builds on these existing State regulatory processes.

Coal companies reporting measured methane from underground mines under NGERS are also required to measure and report carbon dioxide emissions. This is significant as it is the first time that fugitive carbon dioxide data has been available for Australian coal mines.

Choice of emission factor

Estimates based on direct measurements were reported for 38 mines under NGERS in 2009. Emissions for the remaining 5 underground coal mines for which tier 3 data were not available have been estimated by the application of the average implied emission factor for their respective coal fields to data on coal production for those mines consistent with the decision tree in section 1.4.1. It was decided that the NGERS data demonstrated that facility specific emission factors aggregated into subgroups based on spatial correlation (ie by coal field) were sufficiently different from the national country-specific emission factor and drew on the general understanding that mines within coal fields shared common characteristics due to their shared geological history and structure. Detailed discussion as to how time series consistency has been maintained with the inclusion of NGERS data for underground mines is given in section 3.8.3.

Surface Mining (1.B.1aⁱⁱⁱ)

Black coal mine production

Measurements have also been made of CH₄ emissions from open cut coal mines in New South Wales and Queensland (Williams et al. 1993). The empirical results were used to estimate emission factors (in m³/tonne raw coal) applicable to open cut black coal mining in each State, and these are shown in Table 3.32.

Brown coal (lignite) mine production

Open cut mining of brown coal occurs in Victoria for combustion in electricity generation. Emissions are regarded as very low, however, no national emission factor data is available. Given that no IPCC default factor is available, a default methane emission factor of 0.05 m³/t for open cut brown coal mining, as used in the IEA, *Coal Industry Advisory Board, 1992* is used to estimate surface mined methane emissions of lignite/brown coal for Australia.

Open cut sub-bituminous coal mining in Western Australia

Coal mined in Western Australia is regarded as sub-bituminous. In the absence of mine specific factors, the 2006 IPCC Guideline default methane emission factor for surface mining of 1.2m³/tonne is applied. This value is also consistent with the IPCC good practice guidance.

Open cut sub-bituminous coal mining in South Australia

Open cut mining of a low rank sub-bituminous coal occurs in South Australia for combustion in electricity generation. Coal mined in South Australia has an energy content of 13.5 GJ/t. Based on the IEA fuel type classification, which classes non-agglomerating coals under 17.435 GJ/t as being lignite (IEA 2005), the methane emission factor from open cut brown coal mining of 0.05 m³/t (as used for Victorian brown coal) has been applied.

Table 3.32: CH₄ emission factors for open cut coal mining

State	EF CH ₄ m ³ /t raw coal mined	Volume-to-mass conversion factor ^(c) kg/m ³
NSW	3.2 ^(a)	0.6767
Queensland	1.2 ^(a)	0.6767
Tasmania	1.0 ^(b)	0.6767
Western Australia	1.2 ^(d)	0.6767
South Australia	0.05 ^(e)	0.6767
Victoria	0.05 ^(e)	0.6767

(a) Source: Williams et al. (1993) and confirmed by Australian Coal Association.

(b) Source: D Cain, Australian Coal Association, pers. comm. (1993) .

(c) These factors are derived by treating CH₄ as an ideal gas, i.e. 16 g (1 gmole) occupies 23.645 at 15 °C and 1 atmosphere.

(d) Source: IPCC 2006.

(e) Source: IEA 1992.

Total emissions from open cut mines are then estimated by the following algorithm:

$$E_o = (QTY_{state\ i} \cdot EF_{state\ i}) \cdot C_{oc} \dots\dots\dots(1B1_4)$$

Post Mining Activities

Emissions from post mining activities reflect the fugitive escape of gases from the coal after mining—during preparation, transportation, storage or crushing—and are based on the measurements of Williams et al. (1993) and Williams et al. (1996). In these studies the amount of gas retained in coal from gassy underground mines in NSW and Queensland, once the coal reached the surface, was analysed. Most of this gas is likely to desorb from the coal before combustion – during preparation, transportation, storage or crushing – and can therefore be classified as fugitive emissions from post mining activities. These studies related emissions E_{pm} to the quantity of black coal from underground Class A mines QTY_a an emission factor EF_{pm} and C_{pm} the volume-to-mass conversion factor for post mine emissions, which equals 0.6767 kg/m^3 :

$$E_{pm} = QTY_a \cdot EF_{pm} \cdot C_{pm} \dots\dots\dots(1B1_5)$$

The emission factor, E_{pm} , is the average of the results of the two empirical studies. Williams et al. (1993) found that the amount of gas retained was quite variable, but adopted an average gas emission factor of $1.7\text{ m}^3/\text{t}$ raw coal, of which 75% was CH_4 and 25% CO_2 . Williams et al. (1996) estimated an emission factor equal to 20% of the in-situ methane content of coal ($6.78\text{ m}^3/\text{tonne}$ in this case). It is assumed that post mining emissions are associated only with black coal mined in underground Class A (gassy) mines, and not with black coal mined in underground Class B mines, or with brown coal.

Total Emissions from Coal Mining

Total emissions from coal mining in Australia E_{cm} are estimated by summing emissions from underground black coal mining E_u , open cut black coal mining E_o and post-mining activities E_{pm} . This is expressed by the following algorithm:

$$E_{cm} = E_u + E_o + E_{pm} \dots\dots\dots(1B1_6)$$

Decommissioned Mine Emissions (1.B.1.c)

Methane emissions are also known to occur under certain conditions following closure/abandonment of coal mines – leaking to the atmosphere through fractured rock strata, open vents and seals occurs over daily to decadal timescales.

The Australian methodology is based on the approach developed in the 2006 IPCC Guidelines for National Greenhouse Gas Inventories. The decline of emissions following mine closure are modelled using emission decay curves (EDCs) for dry gassy and non-gassy mines. In addition, the EDCs are adjusted on a mine-by-mine basis according to the flooding characteristics of each mine.

Key data required for the approach include:

- mine closure history
- emissions at time of closure
- dry mine EDCs for gassy and non-gassy Australian mines
- mine void size
- mine water inflow rates.

The approach seeks to maximise the use of publicly available data and is best described as a high tier 2/3 approach. It is consistent with a Tier 3 approach in that it estimates emissions on an individual mine basis. However, other mine-specific data characteristic of higher level Tier 3 approach are absent, such as characteristics of the mined coal seam, permeability and direct measured emissions.

The EDC methodology used for estimating methane emissions from decommissioned mines can be described as:

$$E_{dm} = (E_{tdm} \cdot EF_{dm} \cdot (1 - F_{dm})) - E_{rec} \dots\dots\dots(1B1_7)$$

Where E_{dm} is the emissions (Gg methane/year) for a mine at a particular point in time.
 E_{tdm} is the annual emission rate of the mine at point of decommissioning (Gg methane/year)
 EF_{dm} is the emission factor for a mine at a point in time since decommissioning. It is derived from the EDC (formulae 1B1_8 and _9). The emission factor is dimensionless.
 F_{dm} is the fraction of mine flooded at a point in time since decommissioning.
 E_{rec} is the quantity of methane emissions avoided by recovery.

Emission Decay Curves

An emission decay curve (EDC) describes the decline in fugitive methane emissions over time following mine closure. Hyperbolic curves have been found to function best in portraying the rapid decline in emissions in first few years, followed by a slow decline over time of the remaining emissions.

Australian-specific EDCs were utilised for gassy and non-gassy mines respectively. The EDCs represent the dry mine case and have been developed from studies of long term (1982 -2006) direct gas emission measurements from Australian mines Lunarzewski (2005), and Armstrong, Lunarzewski and Creedy (2006). The EDCs are shown in Figure 3.28, and are described in the following formulae:

Gassy mines

$$EF_{dm} = (1 + A \cdot T)^b - C \dots\dots\dots(1B1_8)$$

Non-gassy mines

$$EF_{dm} = (1 + A \cdot T)^b - C \dots\dots\dots(1B1_9)$$

where EF_{dm} is the emission factor for a mine at any point in time since decommissioning (the emission factor is dimensionless).

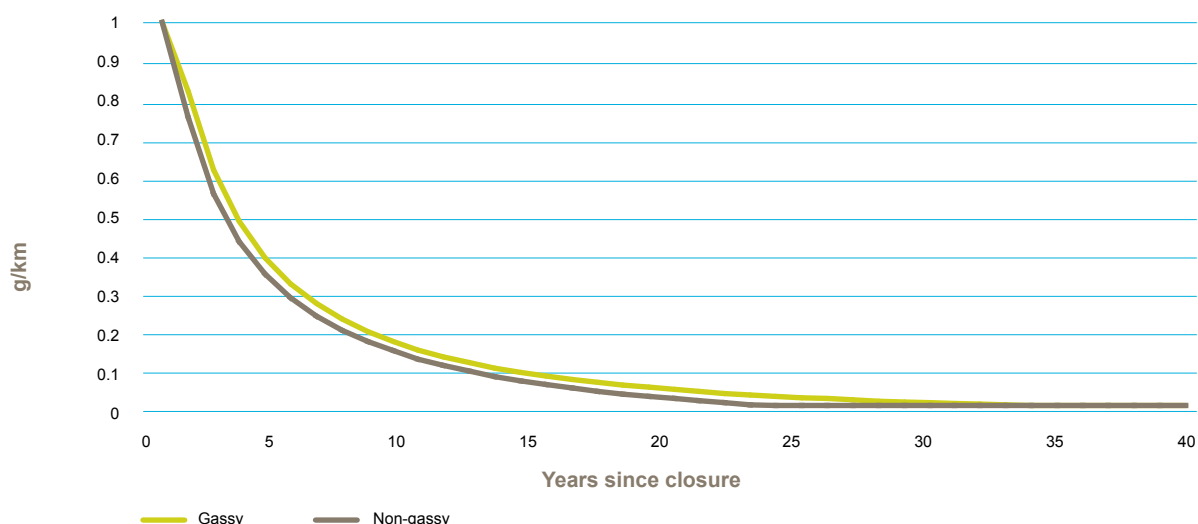
T is the time elapsed since decommissioning of mine.

A , b and C are coefficients unique to the decline curves (see table 3.33).

Table 3.33: Coefficients used in Australian emission decay curves from decommissioned mines

Mine category	Coefficients		
	A	B	C
Gassy Mines	0.23	-1.45	0.0242
Non-Gassy mines	0.35	-1.01	0.0881

Figure 3.28: Emission decline curves for gassy and non-gassy Australian decommissioned coal mines



Mine Production Data

Mine production data are obtained from:

- Coal Services Pty Ltd (formerly the Joint Coal Board) for NSW mines from 1972 to 2009; and
- Queensland Department of Natural Resources, Mines and Energy, from 1979 to 2009.

In both datasets, details were obtained for mine type (underground/opencut), annual run-of-mine production and time of closure. Only underground mines were included in the study. Opencut mines were not included in the study as they are associated with relatively low methane emissions. This approach is consistent with that presented in the 2006 IPCC Guidelines for National Greenhouse Gas Inventories.

Emissions at Closure

In order to estimate the decline of emissions over time following closure, it is first necessary to establish emissions at year zero, that is, emissions at the point prior to closure. The approach used is consistent with that used to estimate methane emissions from active underground coal mines (see 1.B.1ai). Final mine production at closure is taken as the last full year of production.

Decommissioned mines are defined as Class A (gassy) or Class B (non-gassy) based on existing classifications used to calculate previous National Greenhouse Gas Inventories. For earlier mines, for which class tends to be unknown, mines were classified according to their geological proximity to other mines for which class was known.

Adjustment of EDC for flooding mines

It is common for decommissioned mines to become flooded over time. The flooding of mines is known to result in a very rapid decline in the release of methane, thus having a substantial impact on the shape of the EDC and on overall emissions.

The approach uses emission values calculated using dry mine EDCs (formulae 1B1_8 and 1B1_9) and makes adjustments based on the proportion of the mine flooded at that time. For example, if a mine is 50% flooded 10 years post closure then the emission value derived from the EDC is adjusted at that point in time by 50%.

The following information is required in order to estimate the flooding rate of any particular mine:

- Size of the mine void volume
- Rates of mine water inflow.

Estimating mine void volume

The quantity of run-of-mine coal production removed from the mine is used as a basis for estimating the mine void volume remaining at the time of closure. Total historical mine run-of-mine coal production is converted from tonnes to cubic metres by dividing the total tonnage by 1.425, representing the specific gravity of an average Australian worked coal seam Lunarzewski (2006).

Mine water production data are difficult to obtain on a mine by mine basis, particularly for older, decommissioned mines. The approach taken is to develop a set of basin/state average mine water inflow rates based on available data.

The primary source of mine water production rates for individual mines were obtained from publicly available Environmental Impact Statements (EIS) for mining development projects. EIS provides a good coverage of ground water hydrology, providing data on mine water production rates for proposed mines, extensions, nearby existing mines and the flooding status of surrounding mines.

Water production rates for three regions were then calculated. The Southern NSW region contained mine water production rates ranging between 1 – 5.0 ML/Day and an average value of 2.5 ML/day. The Central NSW region ranged between 0.4 and 3 ML/Day and an average value of 1.2 ML/Day and Queensland ranged between 0.1 and 0.4 ML/Day and an average value of 0.2 ML/Day.

The following assumptions were necessary in estimating mine water inflow rates:

1. The mine floods at a linear rate
2. Mine water production is the same for each mine on a basin/state scale
3. Methane is produced evenly throughout the mine and flooding reduces the emissions proportionately to the void volume flooded.

Fully Flooded Mine Emissions

Once a goaf area has been fully flooded, the associated primary gas sources can no longer release gas into the workings, however, remaining free gas in the strata and desorbing gas from unflooded secondary gas sources could continue to leak to the atmosphere (ground surface) via fractured rock strata i.e. geological faults, cracks and fissures (structurally induced pathways). A constant of 2% of the emissions at the time of mine closure has been adopted to represent emissions once fully flooded Lunarzewski (2006).

3.8.3 Uncertainties And Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas.

The transition to the use of NGERs data for underground coal mines has had to be carefully managed to ensure that time series consistency has been respected. It is *good practice* to perform the splicing using more than one technique before making a final decision and to document why a particular method was chosen. The surrogate method, involving the use of coal production data and an emission factor derived from actual mine measurements, was chosen as the most appropriate splicing technique. This choice was made because run-of-mine coal production data is available for individual mines for all years and is an underlying activity data parameter that best explains emission trends.

Interpolation was considered as a complementary approach where emissions data are available from non-NGERS sources for a previous year and which could be used to provide an emission factor per unit of coal production for earlier years. In accordance with good practice (IPCC 2000) interpolated estimates were compared with surrogate data as a QA/QC check.

For a number of years, data on emissions for certain underground mines have been available from estimates published within company environmental reports or from industry reports to the Australian Greenhouse Office (2006b). This emissions data has been used for each mine for the years for which they are available. For earlier years, where such emissions data are not available, then an emission factor per unit of production for each mine was established and applied to production levels back through the time series from 1990 to the year when data on emissions first becomes available (Figure 3.29). For the years between the latest company report and the year of the NGERs data, the emission factor for each mine was

calculated by interpolating between the emission factor for the latest year for which company data was available and the emission factor based on NGERs data for the year 2009.

A small number of underground mines closed in the period 1990-2005 for which there are no mine-specific measured data available. Emissions for each year were recalculated using a basin-specific emission factor calculated from the NGERs data for 2009, multiplied by production. A similar approach has been adopted for the inclusion of emissions of carbon dioxide for all mines (Figure 3.30).

Figure 3.29: Time series consistency method for determining underground coal mine emission factors – methane

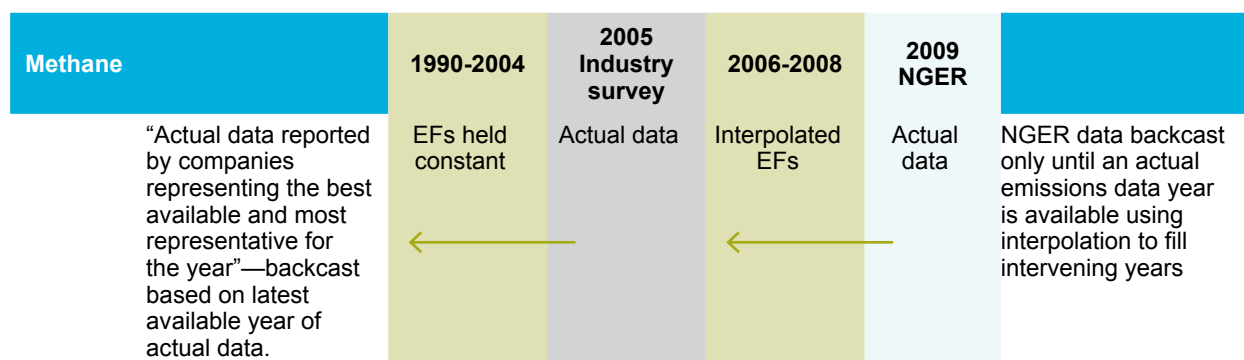
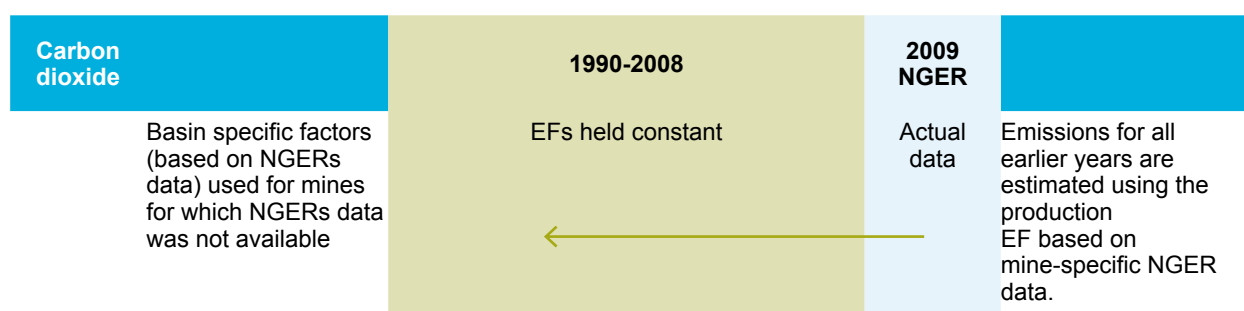


Figure 3.30: Time series consistency method for determining underground coal mine emission factors – carbon dioxide



3.8.4 Source Specific QA/QC

This source category is covered by the general QA/QC of the greenhouse gas inventory in section 1.6.

Implied emission factors

Methane emissions

Australia uses mine site direct measurement tier 3 and country specific tier 2 methods for estimating methane emissions. Australia’s implied methane emission factors have been compared statistically against emission factors reported by other countries. Overall, Australia’s implied emission factors are not significantly different to the means of the implied emission factors of all other reporting parties.

In 2009, Australia’s implied emission factor for underground mines was 7.45 kg CH₄/t compared to 8.83 kg CH₄/t (n = 20) for the 2008 mean of countries. The result of a t-test comparison of the means showed that Australia’s CH₄ from underground mining is not significantly different to that of the mean for all countries.

Australia’s implied emission factor for surface mining, in 2009, was 1.07 kg CH₄/t compared to 0.63 (n = 17) for the 2008 mean of countries and was at the high end of the distribution range. However, the result of a t-test comparison of the means showed that Australia’s CH₄ from surface mining is not significantly different to that of the mean for all reporting parties.

Carbon dioxide emissions

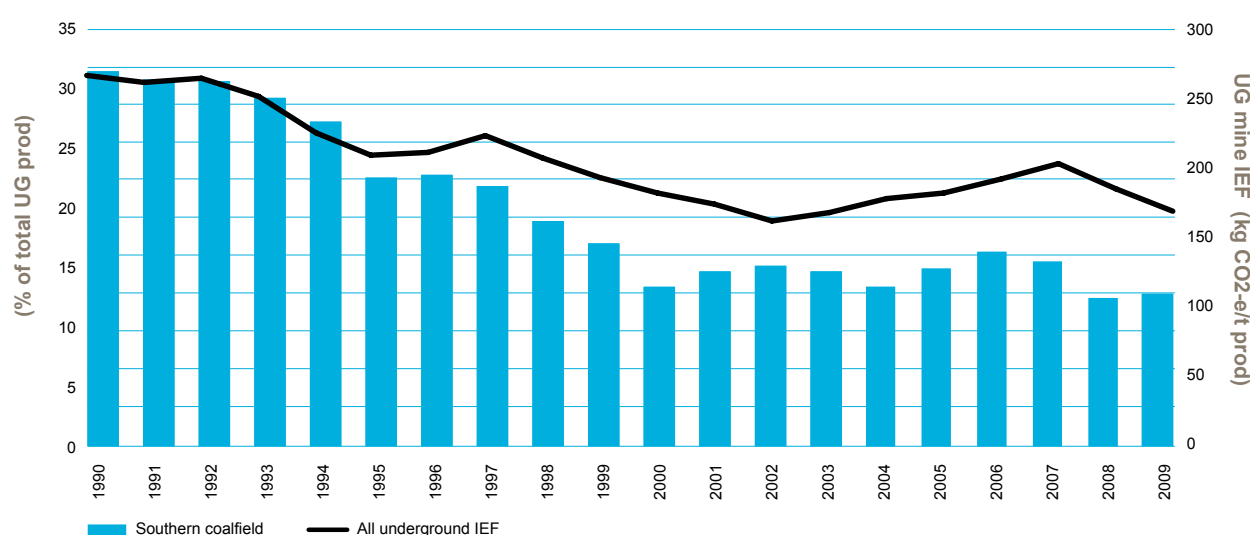
Australia reports carbon dioxide emissions from underground mines. All carbon dioxide emissions are based on direct measurement. The implied emission factor for underground mining, in 2009, was 9.87 kg CO₂/t. Statistical comparison with other countries was not possible as very few countries report carbon dioxide emissions from coal mining.

Trends in implied emission factors

Implied emission factors from coal mining activities are influenced over time by changes in the share of production from mines of varying gas content and gas type and the quantity of methane recovered. This is evident in a declining trend of the methane IEF for underground mines which reflects a relative increase in production from less gassy mine regions and compared to production from high gas coalfields. Figure 3.31 details the declining trend of the underground coal mine IEF since 1990 and the corresponding fall in production from the NSW Southern Coalfield which has the highest IEF of Australian coalfields.

The implied emission factor for all coal mining activities has also declined since 1990 reflecting the additional influence of a relative increase of open cut mine production compared to underground production. The trend in production also varies over time, reflecting the effects of opening and closure of large mines, commodity prices and global demand.

Figure 3.31: Decline of the overall underground coal mine implied emission factor compared with the fall in production from the high gas content Southern Coalfield



Source: Coal Services Pty Ltd and NGERs

3.8.5 Recalculations Since the 2008 Inventory

No recalculations have been made following the extensive recalculation made in the previous NIR as a result of the incorporation on NGERs data for underground coal mines.

Table 3.34: 1.B.1 Solid Fuels: recalculation of total CO₂-e emissions (Gg), 1990-2008

	1990	2000	2001	2002	2003	2004	2005	2006	2007	2008
1.B.1 Solid Fuels										
2010 submission	19,287	22,502	22,561	22,079	21,957	23,746	25,178	25,720	29,003	28,870
2011 submission	19,287	22,502	22,561	22,079	21,957	23,746	25,178	25,720	29,003	28,870
Change	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Percent Change	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00

3.8.6 Planned Improvements

For the first year of reporting under the NGERs underground mine operators could elect to use measured emission data to calculate their emissions rather than use the default emission factor method. The outcome was that 38 out of the 43 mine elected to utilise measured gas data for the first reporting year 2009. For 2010, under the provisions of the *NGER Measurement Determination 2008*, all operators of underground mines must estimate emissions using the higher tier measurement. Consequently, some small revisions to the time series of emissions could be expected in Australia's next submission following the reporting of this data by companies to the Greenhouse and Energy Reporting Office.

For open cut mines, NGERs provides the choice of the national default emission factor method or the use of a higher tier method for estimating gas released annually from a mine's *in situ* gas stock. The data required for the operation of this method is best obtained prior to extraction. Consequently, all operators of currently active mines have reported under NGERs using the current default emission factors for 2009. However, uptake of the higher tier method is expected over future years as the details of the method implementation are refined and as new mining areas are opened up, resulting in an increase in mine-specific emission data availability for open cut mines.

Australia is also investing in the refinement of measurement techniques associated with fugitive coal mine emissions. A comprehensive program of measurement technique development is currently underway in order to underpin future emissions estimation processes under NGERs. The program is a collaboration of government, industry, CSIRO researchers and commercial laboratories and has several streams looking at underground and open cut emission sources and laboratory measurement procedures. Outcomes from the program are anticipated to significantly increase the quality of Australia's coal mining fugitive emissions estimation through the development of more rigorous methods for the measurement of gas. The DCCEE is jointly managing the program through membership of the Australian Coal Association Research Program Fugitive Emission Steering Committee, which provides guidance and oversight of the methodology development and research projects.

Specific components of the program include development of:

- laboratory methods to permit more accurate measurement of gas in samples taken from boreholes from mine sites or in coal to reduce the minimum detectable limit from 0.3m³/tonne to below 0.1 m³/tonne;
- methods for the estimation of mine-specific gas-in-place estimates for open cut mines and a related approach for the estimation of emissions from very low gas regions;
- procedures to permit the use of new technology online measuring devices in underground mines;
- methods to measure errors and uncertainty; and
- methods for mines to estimate residual gas in coal associated with post-mining activities as an alternative to the current default emission factor.

3.9 SOURCE CATEGORY 1.B.2 – OIL AND NATURAL GAS

3.9.1 Source Category Description

The IPCC guidelines define a two level hierarchical structure for source categories related to the oil and gas industries. At the top level of the hierarchy is:

- emissions related to oil (1B2a)
- emissions relating to gas (1B2b)
- venting and flaring emissions relating to both oil and gas (1B2c).

The main emission sources included in the first two categories are leakage, evaporation and storage losses. Emissions from venting and flaring are activities that are managed as part of normal operations at field processing facilities and oil refineries.

The approach used for defining vents and leaks is provided below and has been developed with a view to completeness and consistency with American Petroleum Institute (API) – *Compendium of Greenhouse Gas Emissions Methodologies for the Oil and Gas Industry*:

- vents are emissions that are the result of process or equipment design or operational practices; and
- leaks are emissions from the unintentional equipment leaks from valves, flanges, pump seals, compressor seals, relief valves, sampling connections, process drains, open-ended lines, casing, tanks, and other leakage sources from pressurised equipment not defined as a vent.

In Australia, most onshore exploration wells are drilled to delineate new reserves in established gas fields or in the expectation of discovering new gas fields rather than the expectation of finding oil. Therefore in Australia's inventory, oil and gas exploration emissions are treated as a single category (reported under 1.B.2.b.i Natural gas exploration), which should be seen as related to subsequent production and consumption of both oil and gas.

In the case of natural gas, field processing is typically a considerably more complex process than in the case of crude oil production. Natural gas processing includes some of the same types of processes which, in the case of oil, occur at the refinery, e.g. separation into product streams, such as ethane, LPG and condensate, and removal of undesirable components of the raw gas stream, such as CO₂ and sulphur compounds.

Liquefaction of natural gas for export takes place at the North West Shelf liquefied natural gas (LNG) plant near Dampier in Western Australia and at Darwin in the Northern Territory. Combustion of raw natural gas used in the gas processing and liquefaction for energy purposes is reported under stationary energy 1.A.1.C Manufacture of Solid Fuels and Other Energy Industries. Venting of CO₂ extracted from raw gas and flaring of unusable gas produced at various stages of processing are separately accounted for under venting and flaring.

Natural gas leakage in residential and commercial sectors (1.B.2.b.v.i) is included in the methodology for the estimation of emissions for natural gas distribution. Therefore these emissions are reported under 1.B.2.b.iv Natural gas distribution.

Emission data obtained from NGERS for Natural gas leakage in industrial plants and power stations (1.B.2.b.v.ii) has been included in 1.B.2.b.ii Natural gas production/processing.

For the years 1990 to 2008, emissions for the upstream oil and natural gas categories were based on estimates published by the Australian Petroleum Production and Exploration Association (APPEA). This covered the inventory categories:

- Oil and Gas exploration
- Crude oil production
- Natural gas production and processing
- Venting
- Flaring

APPEA emission factors were largely based on oil and gas industry factors from the E and P Forum 1994. APPEA estimated flaring emissions using country-specific emission factors applied to a measured activity parameter (volume flared). Venting emissions are reported from measured data (tier 3).

NGERS facility data were available for the first time for the compilation of the 2009 inventory. The NGERS data has now replaced the APPEA data, which had ceased to be collected. The NGERS methodologies for facilities to estimate emissions were designed to be consistent with those used to compile the previous APPEA inventories. This ensured time series consistency in methodologies and emissions between the 2009 year and previous years. The emission factors are listed under each respective source category description.

Emissions from Natural Gas Distribution and Transmission are based on country-specific emission factors and activity data derived from statistics published by the Australian Gas Association and the Energy Supply Association of Australia (ESAA). Oil and natural gas production activity data are obtained from the Department of Resources, Energy and Tourism monthly series, *Australian Petroleum Statistics*.

3.9.2 Methodology

Oil (1.B.2a)

Oil and Gas Exploration (1.B.2ai and 1.B.2bi)

Emissions may occur during the process of drilling for oil and gas either during exploration or development drilling, whenever gas or liquid hydrocarbons are encountered. Short term testing activities of hydrocarbon flows and pressure may be undertaken. In the absence of collection infrastructure, which is generally the case in exploration, the hydrocarbons will be usually flared as a means of disposal. Carbon dioxide, some unburnt methane and other non-carbon dioxide gases are released as a result of the flaring. Emissions are estimated using the emission factors reported in Table 3.35. For reporting purposes, exploration emissions have been reallocation from 1.B.2.a Oil to 1.B.2.b Natural gas due to NGERS data showing that the majority of exploration emissions were associated with natural gas exploration rather than oil exploration.

Table 3.35: Oil exploration fugitive emission factors

Inventory Category	Operation/source	Emissions (t) / combustion (kt)					
		CO ₂	CH ₄	NM VOC	N ₂ O	NO _x	CO
Exploration (for both oil and gas)	Offshore testing:						
	- Gas Flared	2750	35.0	15.0	0.081	1.5	8.7
	- Oil Flared	3200	0.33	2.97	0.22	3.7	18.0
	Onshore Production testing:						
	- Gas Flared	2750	35.0	15.0	0.081	1.5	8.7
	- Liquids Flared	2750	0.33	2.97	2.2	3.7	18

Source: APPEA 2005, E and P Forum 1994 and NGER Measurement Determination 2008

Crude Oil Production (other than venting and flaring) (1.B.2aii)

Emissions of CH₄ and NMVOCs may occur during oil production, including field processing, as a result of:

- leakages at seals in flanges, valves, and other components in a variety of process equipment; and
- storage tanks and losses of gases during oil production.

Emission factors for crude oil production are shown in Table 3.36.

Table 3.36: Crude oil production fugitive emission factors

Inventory Category	Operation/source	Emissions (t) / throughput (kt)					
		CO ₂	CH ₄	NM VOC	N ₂ O	NO _x	CO
Crude oil production	Production leaks		0.057				
	Internal Floating tank		0.00004	0.0002			
	Fixed roof tank		0.0002	0.112			
	Floating tank		0.0002	0.0009			

Source: APPEA 2005, E and P Forum 1994 and NGER Measurement Determination 2008

Crude Oil Transport (1.B.2aiii)

The marine (or road/rail) transport of crude oil results in the emissions of NMVOCs, CH₄ and dissolved CO₂. The extent of emissions depends on the gas control technology employed during transfer operations, fuel properties, e.g. vapour pressure and gas composition, ambient temperatures, trip duration and the leak integrity of tanks.

Emissions associated with the marine transport of crude oil are of three types: loading, transit, and ballasting. From the use of USEPA data, it is estimated that 745 kg CH₄ is emitted per PJ of oil tankered (IPCC, 1997, Volume 3). Using the USEPA finding that CH₄ makes up 15% of the mass of total organic emissions (USEPA, 1995b), the NMVOC emission factor for marine transport is estimated to be 4,200 kg per PJ of oil tankered.

Fugitive emission estimates are reported for three categories of oil: indigenous crude oil used within Australia, exported crude oil and imported crude oil. Fugitive emissions from the cargoes of ships engaged in international trade are a component of international bunker fuels, which are excluded from national inventories.

The volume of indigenous crude oil transported by ship to Australian refineries is assumed to equal indigenous crude oil production, minus crude oil exports, minus petroleum product consumption in Victoria (where the great majority of the crude oil processed at Victorian refineries is transported via pipeline from the Bass Strait oil fields). Bass Strait oil is transported to other Australian refineries by ship, as is most of the oil from other Australian oil fields used by Australian refineries.

Crude Oil Refining and Storage (1.B.2aiv)

Crude oil is refined to numerous products via a wide variety of physical and chemical processes. During such processing, fugitive emissions of NMVOCs and CH₄ are generated. Fugitive emission sources at crude oil refineries include valves, flanges, pump and compressor seals, process drains, cooling towers, and oil/water separators.

Crude oil is stored at pipeline pump stations and refineries. During such storage, NMVOCs and CH₄ are emitted from normal processes such as tank breathing, and working and standing losses. Storage or tank losses are a complex function of a number of variables including tank characteristics, fuel properties, meteorological conditions, vapour emission control, and liquid throughput. In the absence of data at the individual refinery level, national CH₄ emissions from crude oil refining and storage may be calculated using default emission factors according to IPCC guidelines (1997). The mid range IPCC default emission factors are adopted for crude oil refining and storage, i.e. 745 kg/PJ for refining and 140 kg/PJ for storage.

Fugitive emissions of NMVOCs resulting from crude oil refining and storage have been estimated for Victoria (Carnovale et al., 1991). Based on the Victorian data it is estimated that the NMVOC emission factor associated with fugitive and tank storage/loading is 20,000 kg/PJ of oil refined.

The NGERS data has provided for the first time data on the emissions associated with the burning of refinery coke to restore the activity of the catalyst during the petroleum refining process. National CO₂ emissions from this source for 2009 were 1.7 million tonnes CO₂-e. Refineries utilised NGERS methodologies involving measurement of flue flow rates, flue gas composition and reference to the Fluid Catalytic Cracking handbook used in the petroleum refining industry.

Consistent with previous practice, and in order to maintain time series consistency, this source of emissions has continued to be included within total refinery fuel combustion. This remains consistent with practice followed by most other countries. Furthermore, the IPCC guidelines are ambiguous as to whether emissions from this source should be reported as fuel combustion or fugitive emissions.

Oil refinery flaring

The composition of refinery flare feed-gas is highly variable and depends on plant processing, process upsets and flare operation. In this inventory the composition of refinery gas directed to flares is assumed to be 30% CH₄, 30% NMVOCs and 40% H₂ (by volume). An average flare combustion efficiency of 98% is used based on studies by USEPA (1995b).

For the years 1990 to 2008, the quantity of gas flared is calculated as 0.6 percent of the total ABARES annual refinery feedstock as no detailed data has been available on refinery flaring volumes. The methodology considered the range and age of technologies of the Australian refining industry and publicly available information on annual flaring emissions from Australian facilities. These assumptions were reviewed in GHD (2006b).

Facility level data on flaring volumes have become available for the first time in 2009 through NGERS. Analysis has shown that the flared quantity based on NGERS data is consistent with the assumptions used to derive the activity data prior to 2009. Given that flaring quantities depend on facility-specific technology types and processes, as well as the episodic nature of flaring, it was decided that it was not appropriate to interpolate the NGERS activity data back through the time series.

The emissions factors for flaring are country-specific factors used consistently throughout the time series and are provided in Table 3.37 below.

Table 3.37: Emission factors for flaring of gas at oil refineries

Unit	CO ₂	CH ₄	N ₂ O	NO _x	CO	NMVOCS
Kg/t gas flared	2,695	6.8	0.081	1.5	8.7	12
Gg/PJ energy flared	47.2	0.12	0.001	0.026	0.15	0.21

Source: DCCEE estimates, following methodology of E&P Forum (1994)

Petroleum Product Distribution (1.B.2av)

The distribution of petroleum products represents a significant source of fugitive NMVOC emissions. Emission sources include motor vehicle refuelling, service station tank filling and breathing losses, major fuel-terminal storage and tank filling losses and refuelling of aircraft and other mobile sources.

The NMVOC emission factors for fuel storage tanks are a complex function of a number of variables and are shown in Table 3.38 on the basis of emissions per sales volumes of each product distributed in Australia. These emission factors are calculated from a weighted average analysis of fuel transfer and storage regulations in different regions of Australia (see Appendix 3.A.21 and 3.A.22).

Table 3.38: NMVOC emission factors for petroleum product distribution (kg/kl distributed)

Emission Source	Emission factor (kg/kl distributed)		
	Petrol	Diesel	Avgas
Motor Vehicle/Equipment Refuelling	1.40 ^a	0.084 ^b	N/A
Service Station/Premises, Storage/Transfer	0.66 ^c	0.006 ^d	N/A
Bulk Fuel Terminal, Storage/Transfer	1.08 ^c	0.009 ^d	N/A
Aircraft, Refuelling/Storage	N/A	N/A	2.69 ^e
Total all sources	3.14	0.099	2.69

Sources: (a). USEPA (1995) Uncontrolled refuelling and spillage.

(b). USEPA (1992) Uncontrolled refuelling and spillage.

(c). See Appendix Table 3.A.21 and 3.A.22.

(d). Scaled according to ratio of diesel/petrol emission rate for tank breathing and emptying as reported in USEPA (1992).

(e). Australian Environment Council (1988).

A number of assumptions were made in compiling these emission factors. Emissions from refined petroleum products in storage and in transit are assumed to be negligible, meaning that all emissions are associated with transfer and fuelling operations. Emissions associated with the normal distribution of LPG are also assumed to be negligible (Environment Protection Authority Victoria 1991; Environment Protection Authority NSW 1995). From a consideration of emission factors (USEPA, 1992) and the predominant modes of distribution of avtur and fuel oil, emissions of NMVOCs from the distribution of these fuels are estimated to be negligible.

Natural Gas (1.B.2b)

Natural Gas Production and Processing (other than venting and flaring) (1.B.2bi)

This category represents leakage emissions from natural gas production and processing, and includes emissions from the unintentional equipment leaks from valves, flanges, pump seals, compressor seals, relief valves, sampling connections, process drains, process drains, open-ended lines, casing, tanks, and other leakage sources from pressurised equipment not defined as a vent.

Emission factors for natural gas production and processing leaks are country-specific factors and are shown in Table 3.39. These factors are based on APPEA data which utilised emission factors from the E and P Forum 1994.

Table 3.39: Fugitive emission factors for natural gas

Inventory category	Operation/source	Emissions (t) / Output (kt)					
		CO ₂	CH ₄	NM VOC	N ₂ O	NO _x	CO
Venting		M	M	M	0	0	0
Flaring		2900	35	15	0.81	1.5	8.7
Natural Gas Production and processing	Fugitive emissions – general leaks		0.058	0.007	0	0	0

Source: APPEA 2005 and E and P Forum 1994. M = directly measured emissions

Natural Gas Transmission and Distribution (1.B.2bii)

Natural gas transmission

Australia has an extensive system of long distance natural gas transmission pipelines. As with oil and gas production, emissions may occur as a result of compressor starts (for which gas expansion is typically used to start gas turbine power units), blowdowns for maintenance at compressor stations, maintenance on pipelines, leakage and accidents.

The Australian high pressure gas transmission system is of relatively recent vintage (the oldest line dates from 1969), has been built to high quality standards and is well maintained. Work undertaken by the Pipeline Authority (the organisation formerly responsible for operation of the Moomba to Sydney pipeline) concluded that losses from a typical gas transmission pipeline in Australia are 0.005% of throughput.

The factor of 0.005% and the throughput data are used in conjunction with national average pipeline gas composition figures for each year, as given in Table 3.32. Throughput data are obtained from the Australian Gas Association (AGA) and the Energy Supply Association of Australia (ESAA). *IPCC Good Practice Guidance (2000)* recommends an approach where emissions are also linked to the length of pipeline rather than solely using throughput. Consistent with this approach, emissions are calculated for a reference year and emissions for other years scaled against the reference year according to the change in pipeline length.

Natural gas distribution

The boundary between natural gas transmission and distribution is generally taken to be the city gate regulator stations at which gas pressures are reduced from transmission pressures (up to about 15 MPa) to sub-transmission pressures. Most of the gas lost from gas transmissions and distribution systems is by way of leakage from the low-pressure network. The amount of leakage depends on the number and condition of joints in the pipes. The high pressure and trunk main pipes are welded steel, so flanged joints are typically only at valves and compressors. Pressures are so high that any major leaks that might occur are obvious, dangerous and quickly attended. Other causes of fugitive emissions from gas distribution systems (up to and including customer meter) are:

- third party damage (e.g. excavators)
- purging of new mains
- unburnt gas from gas compressors (if there are any on the distribution system)

- gas lost to atmosphere on start up and shut down of compressors
- regulating and relief valves.

There are no Australian data on fugitive emissions from the customer side of the meter, but these may arise from such sources as:

- leaking lines at fittings
- purging of lines during appliance installation and maintenance
- leaking appliance valves
- extinguished pilot lights without automatic cut-off
- leakage when intermittently operated appliances (e.g. cookers) are ignited and extinguished.

Emissions from the distributor side of the meter are not measured directly, but must be based on estimates of unaccounted for gas (UAG). Components of UAG include: leakage emissions, meter inaccuracies, use of gas within the system itself, theft of gas, variations in temperature and pressure and differences between billing cycles and accounting procedures between companies delivering and receiving the gas.

The ratio of emissions to UAG for Australian utilities has been estimated at 80% (Dixon 1990) and 70–80% (Hutchinson et al. 1993). A leakage component for UAG of 90% is used for 1990 (NGGI C 1994), reflecting an additional allowance for the additional emissions from the customers' side of the meter, which were not covered in the two studies. An analysis of industry data on the progressive upgrade of the gas distribution infrastructure in response to a variety of drivers, including greenhouse gas emissions concerns, concluded that a figure in the range of 50–60% is more realistic for current circumstances (Energy Strategies 2005). Accordingly, the new estimate for leakage under UAG adopts a figure of 55% from 2003 onwards.

The data sources necessary to calculate emissions from natural gas distribution are:

- estimates of UAG as a percentage of gas issued annually by gas utilities in each State, published in the Energy Supply Association of Australia series; *Electricity, Gas Australia*.
- annual gas utility sales in each State and Territory, published in the Energy Supply Association of Australia series; *Electricity, Gas Australia*; this figure is sales through the low pressure distribution system, and excludes sales made through high pressure mains to electricity generators and large industrial customers
- the composition of pipeline gas supplied in each State and Territory pipeline system (Table 3.40).

Table 3.40: Natural gas composition and emission factors, 2009

Pipeline	Longford, Melb. (Victoria)	Moomba, Sydney, Adelaide (NSW, SA)	Roma, Bris. (Qld)	Denison, Gladstone (Qld)	Dampier, Perth (WA)	Dongarra, Perth (WA)	Amadeus, Darwin (NT)	Australia (wtd average)
kg CO ₂ /GJ	0.9	0.8	0.8	0.7	1.0	1.5	0.0	0.88
kg CH ₄ /GJ	15.5	15.6	15.0	16.0	13.9	16.2	12.6	14.9
kg NMVOC / GJ	2.5	2.4	3.2	1.8	4.3	1.8	5.8	3.2
Weighted state averages:								
kg CO ₂ /GJ			0.8		1.1			
kg CH ₄ /GJ			15.1		14.3			
kg NMVOC / GJ			3.1		3.9			

Note: In previous years, emission factors for both combustion and fugitive natural gas emissions have been calculated from pipeline gas composition data, published by the Australian Gas Association. However, as from 2001 the Association no longer collect this data from its members. It has been assumed that all natural gas emission factors are unchanged from 2000.

Oil and Gas Production Venting and Flaring (1.B.2c)

Venting refers to emissions that are the result of process or equipment design or operational practices. Venting at oil and gas processing facilities is mainly associated with the release of CO₂, which is extracted from the raw gas stream in the course of gas processing. Because separation of the other components of the gas stream from the CO₂ is incomplete, the vented CO₂ contains small quantities of CH₄. The quantities of CO₂ and CH₄ vented will depend on the concentration of CO₂ in the raw gas, which varies significantly between gas fields, and on the mode of operation and efficiency of the CO₂ stripping plant. Gas processing facilities monitor the volumes of the vent gas and CO₂ and CH₄ concentrations as a part of routine plant operation. The venting of CH₄, also occurs from gas assisted pumps and cold process vents.

Flaring refers to the controlled combustion of a mixed flammable gas stream. At oil and gas processing plants, flared gas may arise from crude oil processing or natural gas processing. Where there is no market for gas separated from the wellhead production stream, the gas is reinjected or flared. With the growth in markets for natural gas and an increase in its value, some Australian petroleum production facilities now operate as combined oil and gas facilities, with both oil and gas as marketable products. At such facilities, smaller quantities of gas are flared as part of normal operation of the various processing units. Gas sent to flare is normally mostly CH₄ with smaller concentrations of other volatile hydrocarbons and is usually different in composition to pipeline gas.

Venting – Gas

From 1990 to 2008, estimates of emissions are based on APPEA 2007 data. The APPEA data consists largely of direct monitored emissions associated with control vent releases, equivalent to a Tier 3 estimation, as well as estimates of emissions from cold process vents. The NGERS approach for 2009 enhanced the methodologies available for technology types by utilising the American Petroleum Institute Compendium methodologies for vents.

Flaring – Oil and Gas Combined

Emissions reported combines flaring from both oil and gas sources. Emission factors can be found in Table 3.39, and are country-specific, sourced from the APPEA industry inventory. The NGERS emission factors are consistent with those used for the APPEA inventory, thus ensuring time series consistency.

3.9.3 Uncertainties And Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas.

Time series consistency is maintained through the use of consistent methodologies and data over time. For the upstream oil and gas categories where NGERS data are replacing APPEA data for 2009 – NGERS methodologies were designed to be consistent with those used to compile the previous APPEA inventories. This ensured time series consistency in methodologies and emissions between the 2009 year and previous years.

3.9.4 Source Specific QA/QC

This source category is covered by the general QA/QC of the greenhouse gas inventory in section 1.6. APPEA's inventories have been independently verified through the Greenhouse Challenge Program by PricewaterhouseCoopers and KPMG.

3.9.5 Recalculations Since The 2008 Inventory

Recalculations made to 1.B.2 oil and gas are detailed at the sub-category level in Table 3.41. The recalculations are due to the reallocation of exploration emissions from 1.B.2.a Oil to 1.B.2.b Natural gas. This was due to the fact that the NGERS data demonstrated the majority of exploration emissions were associated with natural gas exploration rather than oil exploration.

Table 3.41: 1.B.2 Oil and gas: recalculation of total CO₂-e emissions (Gg), 1990-2008

	1990	2000	2001	2002	2003	2004	2005	2006	2007	2008
1.B.2.a Oil (a)										
2010 submission	475	642	779	814	654	645	733	588	618	607
2011 submission	461	545	564	569	547	512	511	462	497	482
Change	-14	-97	-215	-245	-107	-133	-222	-125	-121	-125
Percent Change	-2.9	-15.1	-27.6	-30.1	-16.3	-20.7	-30.3	-21.3	-19.6	-20.7
1.B.2.b Natural gas (b)										
2010 submission	4224	2767	2813	2934	3021	3082	3019	3094	2939	3307
2011 submission	4238	2864	3028	3179	3128	3215	3241	3219	3060	3432
Change	14	97	215	245	107	133	221	125	121	125
Percent Change	0.32	3.50	7.64	8.34	3.53	4.31	7.33	4.04	4.11	3.79

Recalculation explanation:

(a) Exploration emissions have been reallocation from 1.B.2.a Oil to 1.B.2.b Natural gas due to NGERS data showing that the majority of exploration emissions were associated with natural gas exploration rather than oil exploration

(b) See above explanation regarding the reallocation of exploration emissions.

3.9.6 Planned Improvements

NGERS data will continue to be analysed for incorporation into the inventory where appropriate.

APPENDIX 3.A ADDITIONAL INFORMATION ON ACTIVITY DATA

Table 3.A.1: Non-CO₂ Emission Factors 1.A.1 Energy Industries

Fuel Type	Emission Factors (Mg/PJ)					
	CH ₄	N ₂ O	NO _x	CO	NMVOC	SO ₂
1.A.1.b Petroleum Refining (ANZSIC Group 251)						
Natural Gas	1.1	0.1	722.9	56.6	1.1	2.3
Crude Oil	3.8	0.6	1234.8	254.6	67.4	57.0
Kerosene	3.8	0.6	1234.8	254.6	67.4	57.0
ADO	3.8	0.6	1234.8	254.6	67.4	57.0
Fuel Oil	0.8	0.6	330.9	44.8	0.8	1282.1
LPG	4.0	0.6	1322.0	349.0	45.0	2.3
Naphtha	0.8	0.6	383.8	54.0	0.8	57.0
Refinery Gas and Liquids	0.8	0.6	383.8	54.0	0.8	2.3
Refinery Coke	0.8	0.6	383.8	54.0	0.8	370.0
1.A.1.c Coke Oven Operation (ANZSIC Class 2711, part)						
Black Coal	1.0	0.8	237.0	201.0	1.0	370.0
Coke Oven Gas	1.0	0.8	287.0	36.7	1.0	370.0
Fuel Oil	1.0	0.6	502.0	75.0	0.9	1282.1
Briquette Manufacture (ANZSIC Class 1103, part)						
Brown Coal	1.3	0.8	287.0	105.0	1.0	150.0
Coal Mining (ANZSIC Subdivision 11, remainder)						
Brown Coal Briquettes	1.3	0.8	287.0	105.0	1.0	150.0
Natural Gas	1.1	0.1	86.24	14.0	1.1	2.3
Automotive Gasoline	8.6	0.6	1306.0	476.0	163.0	15.0
ADO	3.6	0.6	1166.3	221.4	70.0	57.0
LPG	0.1	0.6	154.0	13.0	5.0	2.3
Petroleum products nec	0.8	0.6	383.8	54.0	0.8	57.0
Ethanol	8.6	0.6	1306.0	476.0	163.0	0
Oil and Gas Extraction (ANZSIC Subdivision 12)						
Natural Gas	7.8	0.1	176.3	42.7	2.3	2.3
Ethane	1.1	0.1	492.6	41.8	1.1	2.3
ADO	3.8	0.6	1234.8	254.6	67.4	57.0
Fuel Oil	0.8	0.6	530.9	44.8	0.8	1282.1
LPG	4.0	0.6	1322.0	349.0	45.0	2.3
Petroleum products nec	4.0	0.6	1320.0	346.8	45.5	57.0
Other Transport (Natural Gas Transmission) (ANZSIC Subdivision 65)						
Natural Gas	7.8	0.1	187.9	46.0	2.4	2.3
Gas Production and Distribution (ANZSIC Subdivision 362)						
Natural gas	8.0	0.1	190.0	46.0	2.4	2.3
LPG	4.0	0.6	1322.0	349.0	45.0	2.3

Source: Derived from Table 3.A.4.

Table 3.A.2: Non-CO₂ Emission Factors 1.A.2 Manufacturing and Construction

Fuel Type	Emission Factors (Mg/PJ)					
	CH ₄	N ₂ O	NO _x	CO	NMVOC	SO ₂
Iron and steel (ANZSIC Group 271)						
Black coal	1.1	0.8	450.5	82.2	1.0	370.0
Natural Gas	0.9	0.1	914.6	68.2	1.1	2.3
Coke Oven Gas	1.1	0.1	274.0	36.7	1.0	370.0
ADO	2.5	0.5	553.6	175.5	43.3	57.0
LPG	22.0	0.5	437.0	5465.0	409.0	2.3
1.A.2.b Non-Ferrous Metals (ANZSIC Group 272)						
Black Coal	1.2	0.8	316.9	103.6	1.0	370.0
Coke	1.2	0.8	316.9	103.6	1.0	370.0
Wood and Wood Waste	4.2	4.1	75.0	680.0	6.8	0
Natural Gas	1.1	0.1	519.2	43.6	1.1	2.3
ADO	5.2	0.6	906.7	341.0	93.0	57.0
Fuel Oil	2.0	0.6	294.0	38.5	0.8	1282.1
Naphtha	2.0	0.6	294.0	38.5	0.8	57.0
Petroleum and Coal Product Manufacturing (ANZSIC Group 252)						
Brown Coal Briquettes	1.0	0.8	488.5	77.1	1.0	150.0
Natural Gas	1.1	0.1	199.0	21.3	1.1	2.3
ADO	4.4	0.6	823.0	301.1	81.2	57.0
Fuel Oil	0.5	0.6	274.1	34.5	0.7	1282.1
Liquefied Aromatic Hydrocarbons	0.5	0.6	274.1	34.5	0.7	57.0
LPG	1.5	0.6	281.9	294.5	20.2	2.3
Basic Chemical Manufacturing (ANZSIC Group 253)						
Black Coal	1.3	0.8	287.0	105.0	1.0	370.0
Brown Coal Briquettes	1.3	0.8	287.0	105.0	1.0	150.0
Natural Gas	1.1	0.1	492.6	41.8	1.1	2.3
Ethane	1.1	0.1	482.6	41.8	1.1	2.3
ADO	1.0	0.6	505.0	77.0	1.6	57.0
Liquefied Aromatic Hydrocarbons	1.3	0.8	287.0	105.0	1.0	57.0
LPG	5.8	0.6	487.0	1303.1	39.9	2.3
Naphtha	1.0	0.6	501.7	75.0	0.9	57.0
Petroleum products nec	1.0	0.6	501.7	75.0	0.9	57.0
Chemicals, Rubber and Plastic Products (ANZSIC Groups 254–256)						
Black Coal	1.3	0.8	287.0	105.0	1.0	370.0
Brown Coal Briquettes	1.3	0.8	287.0	105.0	1.0	150.0
Natural Gas	1.2	0.1	190.6	22.4	1.1	2.3
ADO	1.1	0.6	500.2	84.6	5.6	57.0
LPG	6.8	0.6	465.8	1581.0	115.2	2.3
1.A.2.D Pulp, Paper And Print (ANZSIC Subdivisions 23 And 24)						
Black coal	1.3	0.8	287.0	105.0	1.0	370.0
Wood and Wood Waste	4.2	4.1	75.0	680.0	6.8	0
Natural Gas	1.2	0.1	58.4	13.8	1.1	2.3
ADO	5.7	0.6	1006.0	390.0	108.0	57.0
LPG	1.1	0.6	165.8	212.7	15.5	2.3
Petroleum products nec	0.3	0.6	155.6	13.5	0.6	57.0
1.A.2.E Food Processing, Beverages, Tobacco (ANZSIC subdivision 21)						

Fuel Type	Emission Factors (Mg/PJ)					
	CH ₄	N ₂ O	NO _x	CO	NMVOC	SO ₂
Black coal	1.3	0.8	207.1	105.9	1.0	370.0
Brown coal briquettes	1.3	0.8	207.1	105.9	1.0	150.0
Wood and Woodwaste	4.2	4.1	75.0	680.0	6.8	0
Bagasse	9.3	4.1	84.0	1625.0	16.3	0
Natural Gas	1.1	0.1	49.5	11.0	1.7	2.3
ADO	4.9	0.6	868.2	329.1	90.6	57.0
Fuel Oil	0.6	0.6	154.2	13.2	0.6	1282.1
LPG	1.3	0.8	207.1	105.9	1.0	57.0
Ethanol	0.6	0.6	154.2	13.2	0.6	2.3
Textile, Clothing, Footwear and Leather (ANZSIC Subdivision 22)						
Black Coal	1.3	0.8	287.0	105.0	1.0	370.0
Brown Coal Briquettes	1.3	0.8	287.0	105.0	1.0	150.0
Natural Gas	1.2	0.1	58.0	13.7	1.1	2.3
Fuel Oil	0.4	0.6	155.7	13.6	0.6	1282.1
Petroleum products nec	0.4	0.6	155.7	13.6	0.6	57.0
Other Metal Products (ANZSIC Groups 274–276)						
Natural Gas	1.1	0.1	41.0	8.5	2.2	2.3
ADO	0.8	0.6	330.9	44.8	0.8	1282.1
LPG	22.0	0.6	437.0	5465.0	409.0	2.3
Machinery and Equipment (ANZSIC Subdivision 28)						
Natural Gas	1.1	0.1	542.2	44.9	1.1	2.3
ADO	2.8	0.6	651.0	194.3	44.9	57.0
LPG	21.1	0.6	435.6	5242.0	392.2	2.3
Other Manufacturing (ANZSIC Subdivision 29)						
Natural gas	1.1	0.1	189.3	20.3	1.1	2.3
1.A.2.F(II) Construction (ANZSIC Division E)						
Natural Gas	1.0	0.1	58.0	10.0	0.3	2.3
Kerosene	1.0	0.6	160.0	15.0	0.3	57.0
ADO	5.7	0.6	1005.4	389.7	107.9	57.0
Fuel Oil	5.7	0.6	1005.4	389.7	107.9	1282.1
LPG	1.0	0.6	160.0	15.0	0.3	2.3
Glass and Glass Products (ANZSIC Group 261)						
Natural Gas	1.0	0.1	1010.0	75.0	1.1	2.3
LPG	1.0	0.6	502.0	75.0	0.6	2.3
Ceramics (ANZSIC Group 262)						
Black coal	1.0	0.8	502.0	75.0	1.0	370.0
Wood and Wood Waste	4.2	4.1	75.0	680.0	6.8	0
Natural Gas	1.0	0.1	1002.1	74.5	1.1	2.3
ADO	5.7	0.6	1006.0	390.0	108.0	57.0
Fuel Oil	1.0	0.6	493.7	73.5	0.9	1282.1
LPG	8.2	0.6	474.1	1932.7	141.6	2.3
Petroleum products nec	1.0	0.6	493.7	73.5	0.9	57.0
Cement, Lime, Plaster and Concrete (ANZSIC Group 263)						
Black coal	1.0	0.8	500.0	75.7	1.0	370.0

Fuel Type	Emission Factors (Mg/PJ)					
	CH ₄	N ₂ O	NO _x	CO	NMVOC	SO ₂
Coke	1.0	0.8	500.0	75.7	1.0	370.0
Tyres	0.9	0.6	322.8	43.5	0.9	57.0
Wood and Wood Waste	4.2	4.1	75.0	680.0	6.8	0
Natural Gas	1.0	0.1	815.5	61.8	1.1	2.3
Coke Oven Gas	1.0	0.8	500.0	75.7	1.0	370.0
ADO	5.3	0.6	946.6	359.9	98.7	57.0
Fuel Oil	0.9	0.6	946.6	359.9	98.7	57.0
Solvents	0.9	0.6	322.8	43.5	0.9	57.0
LPG	21.7	0.6	435.6	5398.1	404.0	2.3
Petroleum products nec	0.9	0.6	322.8	43.5	0.9	57.0
Non-Metallic Mineral Products nec (ANZSIC Code: 264)						
Black coal	1.1	0.8	465.5	80.0	1.0	370.0
Coke	1.1	0.8	465.5	80.0	1.0	370.0
Natural Gas	0.9	0.1	900.1	67.8	1.1	2.3
ADO	5.7	0.6	1006.0	390.0	108.0	57.0
Fuel Oil	0.4	0.6	477.1	70.6	0.9	1282.1
LPG	19.7	0.6	441.3	4884.1	365.1	2.3
Petroleum products nec	0.4	0.6	477.1	70.6	0.9	57.0
1.A.2.F(iv) Mining (non-Energy Minerals (ANZSIC subdivisions 13,14 and 15))						
Black coal	1.0	0.8	493.0	78.0	1.0	370.0
Coke	1.0	0.8	493.0	78.0	1.0	370.0
Natural Gas	1.1	0.1	86.2	14.0	1.1	2.3
Coke Oven Gas	3.9	0.1	122.6	28.0	1.6	370.0
ADO	3.7	0.6	1190.4	226.2	71.1	57.0
Fuel Oil	0.8	0.6	330.9	44.8	0.8	1282.1
LPG	1.1	0.6	330.9	44.8	0.8	2.3
Petroleum products nec	0.8	0.6	383.8	54.0	0.8	57.0

Source: Derived from Table 3.25.

Table 3.A.3: Non-CO₂ Emission Factors: Other Sectors

Fuel Type	Emission Factors (Mg/PJ)					
	CH ₄	N ₂ O	NO _x	CO	NMVOC	SO ₂
37 Water, Sewerage and Drainage						
Natural Gas	1.2	0.1	58.0	14.0	1.1	2.3
Kerosene	0.1	0.6	154.0	13.0	0.5	57.0
ADO	5.7	0.6	1006.0	390.0	108.0	57.0
65–67 Other Transport, Services and Storage (part)						
ADO	0.6	0.6	53.0	13.0	0.9	57.0
Div. F, G Wholesale and Retail Trade						
Wood and Wood Waste	3.4	4.1	75.0	330.0	5.6	0
Natural Gas	1.1	0.2	42.3	45.5	2.8	2.3
Town Gas	1.1	0.2	42.3	15.6	2.8	2.3

Fuel Type	Emission Factors (Mg/PJ)					
	CH ₄	N ₂ O	NO _x	CO	NMVOC	SO ₂
ADO	0.6	0.6	53.2	13.1	0.9	57.0
Fuel Oil	0.6	0.6	53.2	13.1	0.9	1282.1
LPG	0.6	0.6	53.2	13.1	0.9	2.3
Div. H, P, Q Accommodation, Cultural and Personal						
Wood and Wood Waste	3.4	4.1	19.0	330.0	5.6	0
Natural Gas	1.1	0.1	44.8	9.7	2.0	2.3
ADO	0.6	0.6	53.0	13.0	0.9	57.0
LPG	0.6	0.6	53.0	13.0	0.9	2.3
Div. J Communication						
Natural Gas	1.1	0.1	41.0	8.5	2.2	2.3
Kerosene	0.6	0.6	53.0	13.0	0.9	57.0
ADO	0.6	0.6	53.0	13.0	0.9	57.0
Div. K, L Finance, Insurance, Property and Business						
Natural Gas	1.1	0.1	41.0	8.5	2.2	2.3
Div. M Government Administration and Defence						
Brown Coal Briquettes	1.3	0.8	287.0	105.0	1.0	150.0
Wood and Wood Waste	4.2	4.1	57.0	680.0	6.8	0
Natural Gas	1.1	0.1	46.6	10.3	1.8	2.3
Kerosene	0.4	0.6	84.4	13.0	0.8	57.0
ADO	0.4	0.6	84.4	13.0	0.8	57.0
LPG	0.4	0.6	84.4	13.0	0.8	2.3
Div. N, O Education, health and community services						
Black Coal	1.3	0.8	287.0	105.0	1.0	370.0
Brown Coal Briquettes	1.3	0.8	287.0	105.0	1.0	150.0
Wood and Wood Waste	4.2	4.1	75.0	680.0	6.8	0
Natural Gas	1.1	0.1	44.8	10.4	1.8	2.3
Town Gas	1.1	0.1	44.8	10.4	1.8	2.3
Kerosene	0.4	0.6	53.0	13.0	0.7	57.0
ADO	0.4	0.6	53.0	13.0	0.7	57.0
LPG	0.4	0.6	53.0	13.0	0.7	2.3
Residential						
Wood and Wood Waste ^(a)						
Natural Gas	1.6	0.1	30.0	16.0	3.1	2.3
Town Gas	1.6	0.1	39.0	16.0	3.1	2.3
ADO	4.7	0.6	48.0	13.0	1.9	57.0
LPG	4.7	0.6	48.0	13.0	1.9	2.3
1.A.4.c Agriculture, Forestry & Fisheries: (ANZSIC Division A)						
Natural Gas	1.1	0.1	41.0	476.0	2.2	2.3
Gasoline	8.6	0.6	1306.0	476.0	163.0	15.0
ADO	8.6	0.6	1306.0	14.0	163.0	57.0
LPG	1.3	0.6	154.0	14.0	0.5	2.3

Source: Derived from Table 3.A.4.

(a) See Table 3 A.4 for Residential biomass emission factors.

Table 3.A.4: Derivation of non-CO₂ emission factors for stationary energy

Sector			Emission Factors(Mass/Mass or Volume or Energy)								Conversion Factors		Emission Factors (Mass/Gross Energy Use)				
													CH ₄	N ₂ O ^a	NO _x	CO	NM-VOC
	Fuel	Equipment	CH ₄	N ₂ O ^a	NO _x	CO	NMVOC	Units	Value	Units	CH ₄	N ₂ O ^a	NO _x	CO	NM-VOC		
Mg/ PJ																	
Electricity Generation/Utility																	
1	Natural Gas	Boiler[b]	4.8		8,800	640	23.24	kg/10 ⁶ m ³	38.9	MJ/m ³	0.1	0.1	226	16	0.6		
2	Residual Oil	Boiler[c]	0.034		8	0.6	0.091	kg/10 ³ L	42.9	MJ/L	0.8	0.6	186	14	2.1		
3	Distillate Oil	Boiler[d]	0.002		3	0.6	0.06	kg/10 ³ L	45.6	MJ/L	0.04	0.6	64	13	1.4		
4	Black Coal	Tangentially Fired[e]	0.02		7.2	0.25	0.04	kg/t	23.5	GJ/t	0.9	0.8	306	11	1.7		
5	Black Coal	Pulverised Wall[f]	0.02		10.85	0.25	0.04	kg/t	23.5	GJ/t	0.9	0.8	462	11	1.7		
6	Brown Coal	Tangentially Fired[g]									0.9	1.4	136	17	1.7		
4	Natural Gas	Turbine[h]	7.95		190	46	2.37	ng/J	1.0		8.0	0.1	190	46	2.4		
7	Natural Gas	Internal Combustion[i]	240		1,331	340	80	ng/J	1.0		240	0.1	1,331	340	80		
8	Fuel Oil	Internal Combustion[j]	4		1,322	349	45	ng/J	1.0		4.0	0.6	1,322	349	45		
9	Diesel	Internal Combustion[k]	4		1,322	349	45	ng/J	1.0		4.0	0.6	1,322	349	45		
Other Energy Transformation and Industrial																	
10	Natural Gas	Boiler[l]	47.8		2,240	560	44.16	kg/10 ⁶ m ³	38.9	MJ/m ³	1.2	0.1	58	14	1.1		
11	Residual Oil	Boiler[m]	0.12		6.6	0.6	0.034	kg/10 ³ L	42.9	MJ/L	2.8	0.6	154	14	0.8		
12	Distillate Oil	Boiler[n]	0.006		2.4	0.6	0.024	kg/10 ³ L	45.6	MJ/L	0.1	0.6	53	13	0.5		
13	Black Coal	Boiler[o]	0.03		6.8	5	2.5	0.025	kg/t	23.9	GJ/t	1.3	0.8	287	105	1.0	
14	Wood	Boiler[p]	0.042		0.75	6.8	0.068	kg/t	10.0	GJ/t	4.2	4.1	75	680	6.8		
15	Bagasse	Boiler [q]									10.0	4.1	84	1,625	16.3		
16	Natural Gas	Kiln[r]									1.0	0.1	1,010	75	1.1		
17	Fuel Oil	Kiln[r]									1.0	0.6	502	75	0.8		
18	Black Coal	Kiln[r]									1.0	0.8	502	75	1.0		
19	Black Coal	Coke Oven[r]									1.0	0.8	287	201	1.0		
20	Natural Gas	Dryer[r]									1.0	0.1	58	10	1.1		
21	Fuel Oil	Dryer[r]									1.0	0.6	160	15	0.8		
22	Black Coal	Dryer[r]									1.0	0.8	215	170	1.7		
Commercial																	
23	Natural Gas	Boiler[s]	43.5		1,600	330	84.48	kg/10 ⁶ m ³	38.9	MJ/m ³	1.1	0.1	41	8.5	2.2		
24	Residual Oil	Boiler[t]	0.057		6.6	0.6	0.136	kg/10 ³ L	42.9	MJ/L	1.3	0.6	154	14	3.2		
25	Distillate Oil	Boiler[u]	0.026		2.4	0.6	0.041	kg/10 ³ L	45.6	MJ/L	0.6	0.6	53	13	0.9		
26	Black Coal	Boiler[v]	0.03		3.75	3	0.025	kg/t	23.9	GJ/t	1.3	0.8	157	126	1.0		
27	Wood	Boiler[w]	0.034		0.19	3.3	0.056	kg/t	10.0	GJ/t	3.4	4.1	19	330	5.6		
Residential																	
28	Natural Gas	Heater[x]	61.2		1,500	640	118.8	kg/10 ⁶ m ³	38.9	MJ/m ³	1.6	0.1	39	16	3.1		
29	Black Coal	Hot Water Heater[y]	2.5		4.55	137.5	5	kg/t	23.9	GJ/t	105	0.8	190	5,753	209		
30	Wood	Fireplace[z]	43.5		1.3	126.3	71.0	kg/t	16.2	GJ/t	2,686	4.1	80	7,796	4,382		
31	Wood	Stove[aa]	2.4		1.4	115.4	21.9	kg/t	16.2	GJ/t	148	4.1	86	7,123	1,352		
32	LPG	Furnace[ab]	0.02		1.7	0.2	0.04	kg/10 ³ L	25.3	MJ/L	0.8	0.1	67	7.9	1.6		
33	Distillate Oil	Furnace[ac]	0.214		2.2	0.6	0.085	kg/10 ³ L	45.6	MJ/L	4.7	0.6	48	13	1.9		
General																	
34	Gas	Miscellaneous [ad]									1.1	0.1	41	8.5	2.2		
35	Oil	Miscellaneous [ad]									1.3	0.6	154	14	3.2		
36	Black Coal	Miscellaneous [ad]									1.3	0.8	157	126	1.0		

Sources: [a] IPCC (1997, Volume 3) Assume value is independent of equipment category, i.e. fuel dependent only (0.1, 0.6, 0.8 and 4.1 Mg/PJ for natural gas, oil, coal and wood combustion, respectively).

[b] USEPA (1995b) Pg 1.4-4 to 1.4-6. Uncontrolled emissions of CO, NO_x, NMVOC and CH₄ from natural gas fired utility/large industrial boilers (>29 MW).

[c] USEPA (1995b) Pg 1.3-2 to 1.3-6. Uncontrolled emissions of CO, NO_x, NMVOC and CH₄ from residual oil (No. 4 – 6) fired utility boilers (normal firing).

- [d] Distillate oil fired utility boiler data not available. Assume emissions equal those of residual oil fired utility boiler scaled by relative emissions of industrial boiler category (USEPA, 1986, Pg 1.3-2) .
- [e] USEPA (1995b) Pg 1.1-6 and 1.1-22. Uncontrolled emissions for pulverised coal fired dry bottom configuration (tangentially fired boiler).
- [f] USEPA (1995b) Pg 1.1-6 and 1.1-22. Uncontrolled emissions for pulverised coal fired dry bottom configuration.
- [g] Assume CH₄ and NMVOC emission factors identical to black coal combustion. CO and NO_x emission factors based on average of SECV data (1994). N₂O emission rate based on mid-range value as reported by IEA Coal Research (1993).
- [h] USEPA (1995b) Pg 3.1-3 and 3.1-5. Uncontrolled emissions of CO and NO_x for large stationary natural gas turbines. NMVOC and CH₄ emissions estimated from ratio of NMHC and CH₄, respectively: to Total Organic Compounds for selective catalytic reduction controlled turbines.
- [i] USEPA (1995b) Pg 3.4-3. Assume dual fuel emission factors.
- [j] USEPA (1995b) Pg 3.4-3. Assume diesel emission factors.
- [k] USEPA (1995b) Pg 3.4-3.
- [l] USEPA (1995b) Pg 1.4-4 to 1.4-6. Uncontrolled emissions of CO, NO_x, NMVOC and CH₄ from natural gas fired 'small industrial' boilers (3-29 MW).
- [m] USEPA (1995b) Pg 1.3-2 to 1.3-6. Uncontrolled emissions of CO, NO_x, NMVOC and CH₄ from residual oil fired industrial boilers.
- [n] USEPA (1995b) Pg 1.3-2 to 1.3-6. Uncontrolled emissions of CO, NO_x, NMVOC and CH₄ from distillate oil fired industrial boilers.
- [o] USEPA (1995b) Pg 1.1-6 and 1.1-22. Uncontrolled emissions for spreader stoker firing configuration.
- [p] USEPA (1995b) Pg 1.6-6 to 1.6-7. Uncontrolled emissions from wood waste combustion in stoker boiler. Assume wood moisture content of 50% as recommended by USEPA. Assume CH₄/TOC as for [z].
- [q] IPCC (1997a) data for NO_x and CO converted to gross calorific equivalent by dividing by 1.05. CH₄ and NMVOC emission rates estimated by scaling relative to wood boiler data (see [p]).
- [s] USEPA (1995b) Pg 1.4-4 to 1.4-6. Uncontrolled emissions of CO, NO_x, NMVOC and CH₄ from natural gas fired 'commercial' boilers (0.1-2.9 MW).
- [t] USEPA (1995b) Pg 1.3-2 to 1.3-6. Uncontrolled emissions of CO, NO_x, NMVOC and CH₄ from residual oil (No. 5 and 6) fired commercial boilers (0.1-2.9 MW).
- [u] USEPA (1995b) Pg 1.3-2 to 1.3-6. Uncontrolled emissions of CO, NO_x, NMVOC and CH₄ from distillate oil fired commercial boilers (0.1-2.9 MW).
- [v] USEPA (1995b) Pg 1.1-6 and 1.1-23. Uncontrolled emissions for overfeed stoker firing configuration.
- [w] USEPA (1995b) Pg 1.6-6 to 1.6-7. Uncontrolled emissions from wood waste combustion in Dutch oven boiler. Assume wood moisture content of 50% as recommended by USEPA. Assume CH₄/TOC as for [z].
- [x] USEPA (1995b) Pg 1.4-4 to 1.4-6. Uncontrolled emissions of CO, NO_x, NMVOC and CH₄ from natural gas fired residential furnaces (< 0.1 MW).
- [y] USEPA (1995b) Pg 1.1-7 and 1.1-23. Uncontrolled emissions for hand fired units.
- [z] USEPA (1995b) Pg 1.9-3. Assume CH₄ constitutes 38% of total organic emissions for fireplace (USEPA recommendation for wood combustion).
- [aa] USEPA (1995b) Pg 1.10-3. Assume CH₄ constitutes 38% of total organic emissions for fireplace (USEPA recommendation for wood combustion).
- [ab] USEPA (1995b) Assume propane combustion (commercial boiler) and CH₄ to constitute 34% of total organic emissions (as recommended by USEPA for residential gas fired furnace).
- [ac] USEPA (1995b) Pg 1.3-2 to 1.3-6. Uncontrolled emissions of CO, NO_x, NMVOC and CH₄ from distillate oil fired residential furnace (< 0.1 MW).
- [ad] Assume emission factors for commercial boilers (0.1 – 2.9 MW).

Table 3.A.5: Passenger and light commercial vehicles: CH₄, NO_x and CO emission factors split by urban/non-urban road conditions and hot/cold operation at vehicle group's average VKT

Vehicle Age	Passenger Car			LCV		
	Urban		Non-urban	Urban		Non-urban
	Hot	Cold		Hot	Cold	
EF (g/km)	EF (g/km)	EF (g/km)	EF (g/km)	EF (g/km)	EF (g/km)	EF (g/km)
CH₄						
Post 2005	0.005	0.022	0.005	0.007	0.029	0.008
2004-2005	0.015	0.034	0.011	0.026	0.057	0.021
1998 – 2003	0.022	0.051	0.015	0.047	0.083	0.028
1994 – 1997 (3-way cat)	0.081	0.080	0.054	0.098	0.167	0.056
CO						
Post 2005	0.205	2.234	0.315	0.218	2.853	0.453
2004-2005	0.757	3.956	0.887	1.459	7.402	1.286
1998 – 2003	1.879	8.503	1.534	5.429	14.627	3.430
1994 – 1997 (3-way cat)	7.663	16.206	5.195	14.802	28.031	10.544
NO_x						
Post 2005	0.052	0.133	0.045	0.145	0.328	0.124
2004-2005	0.090	0.281	0.107	0.271	0.360	0.304
1998 – 2003	0.365	0.837	0.547	1.442	1.768	1.915
1994 – 1997 (3-way cat)	0.769	1.366	1.045	1.734	1.982	2.168

Note: As deterioration rates are assumed to be 0 for N₂O the EF at the vehicles group's average VKT is the same as at 0 VKT. The cold start emission factors are reported in the table above as g/km.

Source: Orbital Australia 2011

Table 3.A.6: Passenger and light commercial vehicles: Zero kilometre CH₄ emissions factors split by urban/non-urban road conditions and hot/cold operation

Passenger Car																	
fuel type	Urban						LCV										
	Hot			Cold			Non-urban			Urban		Hot		Cold		Non-urban	
	EF (g/km)	Source	EF (g/start)	Source	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source	EF (g/start)	Source	EF (g/km)	Source	EF (g/km)	Source	
Petrol																	
post 2005	0.002		0.053		0.001		0.002		0.073		0.003						
2004-2005	0.005	Orbital Australia 2011	0.073	Orbital Australia 2011	0.002	Orbital Australia 2011	0.017	Orbital Australia 2011	0.138	Orbital Australia 2011	0.013	Orbital Australia 2011					
1998 – 2003	0.003		0.098		0.000		0.013		0.155		0.000						
1994 – 1997 (3-way cat)	0.073		0.228		0.049		0.048		0.384		0.012						
1985 – 1993 (3-way cat)	0.039	USEPA (as cited in IPCC 2006)	0.151	USEPA (as cited in IPCC 2006)	0.030	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio	0.040	Passenger car EF x USEPA (IPCC 2006) LCV to car EF ratio	0.192	USEPA (as cited in IPCC 2006)	0.030	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio					
1985 – 1997 (2-way cat)	0.112	FORS 1996	0.345		0.063		0.110		0.429		0.085						
1976 – 1985	0.125		0.434		0.065		0.140		0.487		0.087						
Pre 76	0.133	Carnovale 1991	0.461		0.112	Carnovale 1991	0.150		0.521		0.100						
LPG																	
post 2005	0.080		0.240	COPERT IV (converted to a per start EF)	0.025	COPERT IV (Highway)	0.080	Petrol LCV EF x Pass car LPG to petrol ratio	0.240	Petrol LCV EF x Pass car LPG to petrol ratio	0.025	Petrol LCV EF x Pass car LPG to petrol ratio					
2004-2005	0.080		0.240		0.025		0.080		0.240		0.025						
1998 – 2003	0.024		0.096		0.011		0.024		0.096		0.011						
1985 – 1997 (3-way cat)	0.024		0.096	Hot EF x Copert IV (IPCC 2006) cold to hot ratio	0.011	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio	0.024	Petrol EF x USEPA 2006 LPG to petrol EF ratio	0.096	Hot EF x Copert IV (IPCC 2006) cold to hot ratio	0.011	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio					
1985 – 1997 (2-way cat)	0.033	Petrol EF x USEPA 2006 LPG to petrol EF ratio	0.131	Hot EF x Copert IV (IPCC 2006) cold to hot ratio	0.014	Copert IV (IPCC 2006) non-urban to urban ratio	0.033		0.131	Hot EF x Copert IV (IPCC 2006) cold to hot ratio	0.014	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio					
1976 – 1985	0.031		0.125		0.014		0.031		0.125		0.014						
Pre 76	0.032		0.126		0.014		0.032		0.126		0.014						

Passenger Car										
fuel type	Urban						LCV			
	Urban			Non-urban			Hot		Cold	
	EF (g/km)	Source	EF (g/start)	EF (g/km)	Source	EF (g/km)	EF (g/km)	Source	EF (g/start)	Source
ADO										
post 2005	0.003	COPERT IV	0.021	0.000	COPERT IV (converted to a per start EF)	0.003	0.003	COPERT IV (converted to a per start EF)	0.021	COPERT IV (converted to a per start EF)
2004-2005	0.007		0.018	0.002		0.007	0.007		0.018	
1998 – 2003	0.001		0.003	0.000		0.001	0.001		0.003	
1985 – 1997 (3-way cat)	0.001	Petrol EF x USEPA 2006	0.003	0.000	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio	0.001	0.001	Petrol EF x USEPA 2006 diesel to petrol EF ratio	0.003	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio
1985 – 1997 (2-way cat)	0.001	diesel to petrol EF ratio	0.004	0.001		0.001	0.001		0.004	
1976 – 1985	0.001		0.004	0.001		0.001	0.001		0.004	
Pre 76	0.001		0.004	0.001		0.001	0.001		0.004	
Ethanol ^a										
post 2005	0.037	USEPA (as cited in IPCC 2006)		0.049		0.037	0.037			0.049
2004-2005	0.037	– mid-point of reported range		0.049		0.037	0.037			0.049
1998 – 2003	0.037			0.025	Hot EF x Petrol Non-urban to Hot Urban ratio	0.053	0.053	Passenger car EF x LCV to car ratio		Ethanol car hot EF x LCV non-urban to petrol hot urban ratio
1985 – 1997 (3-way cat)	0.206		NA	0.158		0.211	0.211	NA		
1985 – 1997 (2-way cat)	0.592	Post 97 EF x earlier petrol age class relativity		0.331		0.581	0.581			
1976 – 1985	0.661			0.344		0.740	0.740			
Pre 76	0.703			0.592		0.793	0.793			

(a) raw ethanol content of blended fuel

Sources (as indicated in table): FORS (1996); Carnovale et al. (1991); IPCC (2006); Orbital Australia (2011)

Table 3.A.7: Medium and heavy duty trucks and buses: Zero kilometre CH₄ emissions factors split by urban/non-urban road conditions and hot/cold operation

fuel type	Medium Duty Truck						Heavy Duty Truck						Bus			
	Urban			Non-urban			Urban			Non-urban			Urban		Non-urban	
	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source
Petrol																
Post 2002	0.078	COPERT IV (x EF reduction %)	0.062	COPERT IV (x EF reduction %)	0.078	COPERT IV (x EF reduction %)	0.062	COPERT IV (x EF reduction %)	0.078	COPERT IV (x EF reduction %)	0.062	COPERT IV (x EF reduction %)	0.062	COPERT IV (x EF reduction %)	0.062	COPERT IV (x EF reduction %)
1996 – 2002	0.140	COPERT IV	0.110	COPERT IV	0.140	COPERT IV	0.110	COPERT IV	0.140	COPERT IV	0.110	COPERT IV	0.110	COPERT IV	0.110	COPERT IV
Pre 1996	0.140		0.110		0.140		0.110		0.140		0.110		0.110		0.110	
LPG																
Post 2002	0.123		0.054	Passenger car LPG COPERT IV non-urban to urban ratio	0.123	Passenger car LPG COPERT IV non-urban to urban ratio	0.054	Passenger car LPG COPERT IV non-urban to urban ratio	0.123	Passenger car LPG COPERT IV non-urban to urban ratio	0.054	Passenger car LPG COPERT IV non-urban to urban ratio	0.067	Passenger car LPG COPERT IV non-urban to urban ratio	0.029	Passenger car LPG COPERT IV non-urban to urban ratio
1996 – 2002	0.220	USEPA 1989	0.096		0.220	USEPA 1989	0.096		0.220	USEPA 1989	0.096		0.120	USEPA 1989	0.053	
Pre 1996	0.220		0.096		0.220		0.096		0.220		0.096		0.120		0.053	
ADO																
Post 2003	0.048	COPERT IV (x EF reduction %)	0.022	Hot urban EF x COPERT IV non-urban to urban ratio	0.098	COPERT IV (x EF reduction %)	0.045	Hot urban EF x COPERT IV non-urban to urban ratio	0.017	COPERT IV (x EF reduction %)	0.017	Hot urban EF x COPERT IV non-urban to urban ratio	0.008	COPERT IV (x EF reduction %)	0.008	Hot urban EF x COPERT IV non-urban to urban ratio
1996 – 2002	0.157	USEPA 1989	0.072		0.157	USEPA 1989	0.072		0.030	USEPA 1989	0.030		0.014	USEPA 1989	0.014	
Pre 1996	0.157		0.072		0.157		0.072		0.030		0.030		0.014		0.014	

Table 3.A.8: Passenger and light commercial vehicles: Zero kilometre N₂O emissions factors split by urban/non-urban road conditions and hot/cold operation

fuel type	Passenger Car						LCV					
	Urban			Non-urban			Urban			Non-urban		
	Hot		Cold	Hot		Cold	Hot		Cold	Hot		Cold
	EF (g/km)	Source	EF (g/start)	Source	EF (g/km)	Source	EF (g/start)	Source	EF (g/start)	Source	EF (g/km)	Source
Petrol												
post 2005	0.001		0.037		0.001		0.003		0.144		0.001	
2004-2005	0.008		0.121		0.004		0.006		0.087		0.009	
1998 – 2003	0.030	Orbital Australia 2011	0.332	Orbital Australia 2011	0.020	Orbital Australia 2011	0.041	Orbital Australia 2011	0.156	Orbital Australia 2011	0.029	Orbital Australia 2011
1994 – 1997 (3-way cat)	0.037		0.231		0.034		0.023		0.137		0.013	
1985 – 1997 (3-way cat)	0.029		0.178		0.005	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio	0.048	Passenger car EF x USEPA (IPCC 2006) LCV to car EF ratio	0.297	USEPA (as cited in IPCC 2006)	0.048	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio
1985 – 1997 (2-way cat)	0.011	Weeks et al 1993	0.106	USEPA (as cited in IPCC 2006)	0.002		0.015		0.138		0.015	
1976 – 1985	0.004		0.041		0.002		0.005		0.047		0.005	
Pre 76	0.003		0.036		0.002		0.003		0.041		0.002	
LPG												
post 2005	0.005		0.027	COPERT IV (converted to a per start EF)	0.001	COPERT IV (Highway)	0.008	Petrol LCV EF x Pass car LPG to petrol ratio	0.081	Petrol LCV EF x Pass car LPG to petrol ratio	0.003	Petrol LCV EF x Pass car LPG to petrol ratio
2004-2005	0.013	COPERT IV	0.069		0.002		0.026		0.178		0.018	
1998 – 2003	0.016		0.048		0.006		0.016		0.048		0.006	
1985 – 1997 (3-way cat)	0.006		0.017		0.001	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio	0.006	Petrol EF x USEPA 2006 LPG to petrol EF ratio	0.017	Hot EF x Copert IV (IPCC 2006) cold to hot ratio	0.001	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio
1985 – 1997 (2-way cat)	0.003	Petrol EF x USEPA 2006 LPG to petrol EF ratio	0.008	Hot EF x Copert IV (IPCC 2006) cold to hot ratio	0.002		0.003		0.008		0.002	
1976 – 1985	0.003		0.008		0.000		0.003		0.008		0.000	
Pre 76	0.002		0.005		0.000		0.002		0.005		0.000	

Table 3.A.8: Passenger and light commercial vehicles: Zero kilometre N₂O emissions factors split by urban/non-urban road conditions and hot/cold operation (continued)

fuel type	Passenger Car						LCV					
	Urban			Non-urban			Urban			Non-urban		
	Hot		Cold	Hot		Cold	Hot		Cold	Hot		Cold
	EF (g/km)	Source	EF (g/start)	Source	EF (g/km)	Source	EF (g/start)	Source	EF (g/start)	Source	EF (g/km)	Source
ADO												
post 2005	0.009	COPERT IV	0.045	COPERT IV (converted to a per start EF)	0.004	COPERT IV (Highway)	0.009	COPERT IV	0.045	COPERT IV (converted to a per start EF)	0.004	COPERT IV (Highway)
2004-2005	0.004		0.045		0.006		0.004		0.045		0.006	
1998 – 2003	0.003		0.010		0.001		0.003		0.010		0.001	
1985 – 1997 (3-way cat)	0.001	Petrol EF x USEPA 2006	0.003	USEPA (as cited in IPCC 2006)	0.002	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio	0.001	Petrol EF x USEPA 2006	0.003	USEPA (as cited in IPCC 2006)	0.002	Hot urban EF x Copert IV (IPCC 2006) non-urban to urban ratio
1985 – 1997 (2-way cat)	0.001	diesel to petrol EF ratio	0.002		0.001		0.001	diesel to petrol EF ratio	0.002		0.001	
1976 – 1985	0.001		0.002		0.000		0.001		0.002		0.000	
Pre 76	0.000		0.001		0.000		0.000		0.001		0.000	
Ethanol ^a												
post 2005	0.030	USEPA (as cited in IPCC 2006)			0.015		0.049				0.049	
2004-2005	0.030	– mid-point of reported range			0.007		0.059				0.059	
1998 – 2003	0.030				0.025		0.082				0.082	
1985 – 1997 (3-way cat)	0.030			NA	0.029	Post 97 hot Ef x Petrol Non-urban to Hot Urban ratio	0.049	Passenger car EF x LCV to car ratio	NA		0.049	ethanol car hot EF x LCV non-urban to petrol hot urban ratio
1985 – 1997 (2-way cat)	0.012	Post 97 EF x earlier petrol age class relativity			0.011		0.015				0.015	
1976 – 1985	0.004				0.010		0.005				0.005	
Pre 76	0.003				0.010		0.003				0.002	

(a) raw ethanol content of blended fuel

Sources (as indicated in table): Weeks et al. (1993), IPCC (2006), Orbital Australia (2011)

Table 3.A.9: Medium and heavy duty trucks and buses: Zero kilometre N₂O emissions factors split by urban/non-urban road conditions and hot/cold operation

fuel type	Medium Duty Truck				Heavy Duty Truck				Bus			
	Urban		Non-urban		Urban		Non-urban		Urban		Non-urban	
	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source	EF (g/km)	Source
Petrol												
Post 2002	0.006		0.006		0.006		0.006		0.006		0.006	
1996 – 2002	0.006	COPERT IV	0.006	COPERT IV	0.006	COPERT IV	0.006	COPERT IV	0.006	COPERT IV	0.006	COPERT IV
Pre 1996	0.006		0.006		0.006		0.006		0.006		0.006	
LPG												
Post 2002	0.020		0.020	Hot urban EF x COPERT IV non-urban to urban ratio	0.020		0.020	Hot urban EF x COPERT IV non-urban to urban ratio	0.011		0.011	Hot urban EF x COPERT IV non-urban to urban ratio
1996 – 2002	0.020	USEPA 1989	0.020		0.020	USEPA 1989	0.020		0.011	USEPA 1989	0.011	
Pre 1996	0.020		0.020		0.020		0.020		0.011		0.011	
ADO												
Post 2002	0.030		0.030	Hot urban EF x COPERT IV non-urban to urban ratio	0.030		0.030	Hot urban EF x COPERT IV non-urban to urban ratio	0.030		0.030	Hot urban EF x COPERT IV non-urban to urban ratio
1996 – 2002	0.030	COPERT IV	0.030		0.030	COPERT IV	0.030		0.030	COPERT IV	0.030	
Pre 1996	0.025	USEPA 1989	0.025		0.025	USEPA 1989	0.025		0.025	USEPA 1989	0.025	

Table 3.A.10: Vehicle emission factors for indirect gases by year of vehicle manufacture

Fuel type	Passenger Car				LCV				Medium Duty Truck				Heavy Duty Truck				Bus	
	NO _x	CO	NM VOC	NO _x	CO	NM VOC	NO _x	CO	NO _x	CO	NM VOC	NO _x	CO	NO _x	CO	NM VOC	NO _x	CO
Petrol																		
post 2005	0.044	0.108		0.139	0.047													
2004-2005	0.075	0.399		0.275	0.669													
1998 – 2003	0.167	0.037	0.077	0.820	1.664	0.236												
1994 – 1997 (3-way cat)	0.534	6.588		1.512	10.108													
							2.52	10.87	1.04	2.52	10.87	1.04	10.87	2.52	48.61	3.91	3.47	
1985 – 1993 (3-way cat)	0.450	3.850	0.294	0.917	6.712	0.894												
1985 – 1997 (2-way cat)	0.930	5.460	0.260	1.895	9.519	0.791												
1976 – 1985	1.400	14.900	1.419	2.853	25.977	4.314												
Pre 76	2.460	24.000	2.275	5.014	41.842	6.914												
LPG																		
Post 97	0.472	2.327	0.199	0.472	2.327	0.199												
1985 – 1997 (3-way cat)	0.942	10.305	0.755	0.942	10.305	0.755												
1985 – 1997 (2-way cat)	1.947	14.614	0.669	1.947	14.614	0.669	4.83	24.00	4.21	4.83	10.87	4.21	2.76	24.00	2.41			
1976 – 1985	2.931	39.881	3.647	2.931	22.875	3.647												
Pre 76	5.150	64.238	5.846	5.150	36.846	5.846												
ADO																		
Post 97	0.250	0.116	0.062	0.250	0.116	0.062												
1985 – 1997 (3-way cat)	0.500	0.515	0.237	0.500	0.515	0.237												
1985 – 1997 (2-way cat)	1.034	0.731	0.210	1.034	0.731	0.210	5.20	6.44	1.15	5.20	24.00	1.15	4.90	2.88	1.56			
1976 – 1985	1.556	1.994	1.144	1.556	1.994	1.144												
Pre 76	2.734	3.212	1.833	2.734	3.212	1.833												

Note: For light duty vehicles hot urban emission factors are reported in the table above.

Table 3.A.11: Passenger and light commercial vehicles: non-carbon dioxide emission factor deterioration rates (g/km/km)

	Vehicle Age Class							
	Pre '79 ^c	'80-'85 ^c	'85-'93 ^{ac}	'85-'97 ^{bc}	94-97 ^d	98-03 ^d	04-05 ^d	06-current ^d
CH₄	6.35E-07	4.76E-07	8.43E-07	1.49E-06	4.58 E-08	1.44E-07	1.51E-07	1.54E-07
N₂O	0	0	0	0	0	0	0	0
CO	1.45E-04	1.27E-04	7.82E-05	1.57E-04	6.09E-06	1.515E-05	6.728E-06	4.867E-06
NO_x	0.00E+00	6.48E-06	8.91E-06	7.81E-06	1.54E-06	2.03E-06	2.41E-07	2.87E-07
NMVOC^d	9.95E-06	7.45E-06	4.42E-06	7.83E-06	4.424E-06	4.42E-06	4.42E-06	4.42E-06

Notes: (a) 3-way catalyst (b) 2-way catalyst

Sources: (c) EPA NSW 1995 (d) Orbital Australia (2011)

Table 3.A.12: Road transport: non-carbon dioxide emission factors

Source Category Sector	Fuel Type	Emission Factor (g/km)				
		CH ₄ ^a	N ₂ O ^b	NO _x ^c	CO ^c	NMVOC ^c
Medium Trucks	NG ^e	0.101	0.001	1.200	0.200	0.010
Heavy Trucks	NG ^e	0.101	0.001	1.200	0.200	0.010
Buses	NG ^e	0.101	0.001	1.200	0.200	0.010
Motorcycles	Petrol	0.150	0.002	0.210	19.270	4.580
Passenger Cars	NG ^e	0.261	0.001	0.190	0.110	0.020
Light Commercial Vehicles	NG ^e	0.261	0.001	0.190	0.110	0.020

Sources: (a) Hoekman (1992); (b) Weeks et al. (1993); (c) Carnovale et al. (1991); (d) EPA NSW (1995); (e) de Maria (1992).

Table 3.A.13: Shares used to allocate ABARES fuel consumption to unlisted categories, 2009

ANZSIC category fuel consumption reported by ABARES	General use	Military	Small marine craft	Off-road vehicles	Utility engines
Road transport automotive gasoline	96.85%	0.02%	2.39%	0.10%	0.63%
Road transport ADO	99.81%	0.19%			
Water transport ADO	75.50%	24.50%			
Water transport fuel oil	100%				
Air transport aviation gasoline	98.52%	1.48%			
Air transport aviation turbine fuel	91.40%	8.60%			

Sources: Derived from Farrington 1988, ABS 2006 and Department of Defence 2010.

Table 3.A.14: Shares of total road fuel consumption by vehicle and fuel type, 2009

Vehicle Type	Fuel Type			
	Automotive Gasoline	ADO	LPG	NG ^a
Passenger cars	0.849	0.102	0.712	0.053
Light commercial vehicles	0.142	0.180	0.252	0.053
Medium duty trucks	0.002	0.266	0.010	0.010
Heavy duty trucks	-	0.400	0.001	-
Buses	0.002	0.053	0.024	0.884
Motor cycles	0.006	-	-	-

Source: ABS 2008a. (a) Apelbaum Consulting Group 2006

Table 3.A.15: Australian petrol-fuelled vehicle stock age distribution and fuel consumption rates, 2009

Year of manufacture	Passenger cars		Light commercial Vehicles		Medium Duty Trucks		Heavy Duty Trucks		Buses	
	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)
2009 ^b	329906	0.110	31166	0.129	166	0.213	45	0.300	648	0.150
2008	677477	0.110	64084	0.129	401	0.213	102	0.300	1328	0.150
2007	732593	0.110	67512	0.129	668	0.213	232	0.300	1144	0.150
2006	695846	0.111	64884	0.131	722	0.197	104	0.300	1145	0.160
2005	727720	0.111	96583	0.131	644	0.197	66	0.300	824	0.160
2004	689292	0.111	91281	0.131	636	0.197	98	0.300	475	0.160
2003	674963	0.111	84242	0.131	475	0.197	55	0.300	942	0.160
2002	602606	0.111	67515	0.131	550	0.197	46	0.300	774	0.160
2001	579760	0.112	61550	0.135	455	0.189	35	0.312	589	0.151
2000	583898	0.113	57278	0.134	364	0.198	15	0.324	1011	0.140
1999	546065	0.113	61284	0.134	329	0.198	18	0.324	761	0.140
1998	570530	0.113	56742	0.134	434	0.198	38	0.324	850	0.140
1997	486485	0.113	46663	0.134	400	0.198	30	0.324	736	0.140
1996	395930	0.113	45299	0.134	334	0.198	24	0.324	630	0.140
1995	381545	0.113	42496	0.134	313	0.198	30	0.324	626	0.140
1994	339516	0.113	42587	0.134	317	0.198	45	0.324	532	0.140
1993	280389	0.113	32789	0.134	278	0.198	29	0.324	457	0.140
1992	242426	0.113	33133	0.134	392	0.198	27	0.324	420	0.140
1991	207386	0.114	27524	0.132	330	0.211	8	0.352	366	0.155
1990	210031	0.114	31377	0.133	423	0.225	22	0.379	383	0.184
1980 – 1989 ^a	613326	0.114	150814	0.133	6287	0.225	144	0.379	1692	0.184
1979 and earlier	225007	0.114	66177	0.133	19572	0.225	302	0.379	304	0.184

Notes: (a) Fuel consumption rates average for period 1980–89. (b) At the time of emission calculation the latest motor vehicle census was not available so the number of 2009 manufactured motor vehicles was assumed to be the same as in 2008. Assumes new cars on road for average of 6 months in the first year.

Source: DCEEE estimates derived from ABS 2009a, ABS 2008a

Table 3.A.16: Australian diesel-fuelled vehicle stock age distribution and fuel consumption rates, 2009

Passenger cars: year of manufacture	Passenger cars		Light commercial Vehicles		Medium Duty Trucks		Heavy Duty Trucks		Buses	
	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)
2009 ^b	43235	0.112	53462	0.123	10428	0.283	2120	0.556	2376	0.289
2008	90596	0.112	110345	0.123	24680	0.283	4987	0.556	4933	0.289
2007	69471	0.112	88529	0.123	29955	0.283	8921	0.556	3470	0.289
2006	54386	0.118	79023	0.123	22949	0.289	5291	0.563	3368	0.270
2005	39114	0.118	62699	0.123	23282	0.289	5504	0.563	3412	0.270
2004	33888	0.118	56222	0.123	21811	0.289	5182	0.563	2582	0.270
2003	28962	0.118	46279	0.123	16093	0.289	4394	0.563	2405	0.270
2002	24504	0.118	41838	0.123	18642	0.289	3277	0.563	2350	0.270
2001	20887	0.128	29667	0.124	13252	0.290	2437	0.551	2267	0.262
2000	19550	0.123	35066	0.125	12867	0.293	2536	0.532	3112	0.267
1999	16528	0.123	32576	0.125	13484	0.293	2747	0.532	2619	0.267
1998	15910	0.123	30144	0.125	12602	0.293	3123	0.532	2711	0.267
1997	13631	0.123	26167	0.125	10129	0.293	2301	0.532	2330	0.267
1996	11929	0.123	23610	0.125	9083	0.293	1979	0.532	2378	0.267
1995	11978	0.123	21691	0.125	10115	0.293	2287	0.532	2692	0.267
1994	12873	0.123	22034	0.125	11292	0.293	2446	0.532	2662	0.267
1993	12114	0.123	19211	0.125	8998	0.293	1523	0.532	2231	0.267
1992	14505	0.123	17668	0.125	8872	0.293	863	0.532	2231	0.267
1991	12314	0.119	13368	0.126	7606	0.298	642	0.527	1780	0.308
1990	11733	0.134	16113	0.122	11610	0.298	1391	0.526	1849	0.355
1980 – 1989 ^a	48395	0.134	64288	0.122	86366	0.298	10352	0.526	9507	0.355
1979 and earlier	949	0.134	3896	0.122	19347	0.298	3800	0.526	711	0.355

Notes: (a) Fuel consumption rates average for period 1980–89. (b) At the time of emission calculation the latest motor vehicle census was not available so the number of 2009 manufactured motor vehicles was assumed to be the same as in 2008. Assumes new cars on road for average of 6 months in the first year.

Source: DCCCE estimates derived from ABS 2009a, ABS 2008a.

Table 3.A.17: Australian LPG-fuelled vehicle stock age distribution and fuel consumption rates, 2009

Passenger cars: year of manufacture	Passenger cars		Light commercial Vehicles		Medium Duty Trucks		Heavy Duty Trucks		Buses	
	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)	Vehicle numbers	Average Fuel Consumption Rate (L/km)
2009 ^b	5500	0.168	3213	0.147	25	0.372	4	0.644	229	0.530
2008	12595	0.168	6975	0.147	59	0.372	9	0.644	475	0.530
2007	12775	0.168	5944	0.147	69	0.372	27	0.644	281	0.530
2006	15832	0.162	7558	0.151	63	0.441	14	0.600	279	0.471
2005	16309	0.162	6421	0.151	48	0.441	13	0.600	190	0.471
2004	17324	0.162	5761	0.151	45	0.441	6	0.600	147	0.471
2003	20596	0.162	6332	0.151	41	0.441	3	0.600	121	0.471
2002	19664	0.162	6694	0.151	59	0.441	3	0.600	132	0.471
2001	20785	0.160	7413	0.157	35	0.340	2	0.528	284	0.392
2000	20856	0.162	7869	0.155	41	0.276	3	0.500	176	0.362
1999	26570	0.162	7287	0.155	19	0.276	2	0.500	165	0.362
1998	23730	0.162	6480	0.155	27	0.276	6	0.500	55	0.362
1997	20083	0.162	5908	0.155	32	0.276	3	0.500	59	0.362
1996	19041	0.162	6076	0.155	20	0.276	2	0.500	53	0.362
1995	17917	0.162	5918	0.155	28	0.276	3	0.500	139	0.362
1994	17046	0.162	5892	0.155	25	0.276	0	0.500	146	0.362
1993	14260	0.162	4538	0.155	25	0.276	0	0.500	81	0.362
1992	10268	0.162	3534	0.155	47	0.276	0	0.500	61	0.362
1991	6725	0.166	3015	0.167	55	0.305	0	0.466	71	0.305
1990	7924	0.161	3609	0.163	49	0.270	0	0.432	85	0.260
1980 – 1989 ^a	26061	0.161	17215	0.163	1270	0.270	6	0.432	250	0.260
1979 and earlier	8792	0.161	9942	0.163	1816	0.270	5	0.432	25	0.260

Notes: (a) Fuel consumption rates average for period 1980–89. (b) At the time of emission calculation the latest motor vehicle census was not available so the number of 2009 manufactured motor vehicles was assumed to be the same as in 2008. Assumes new cars on road for average of 6 months in the first year.

Source: DCCEE estimates derived from ABS 2009a, ABS 2008a.

Table 3.A.18: Average rate of fuel consumption for road vehicles by vehicle and fuel type

Vehicle Type	Fuel Type		
	Automotive Gasoline (L/km)	ADO (L/km)	LPG / NG (L/km)
Passenger cars	a	a	a
Light commercial vehicles	a	a	a
Medium duty trucks	a	a	a
Heavy duty trucks	a	a	a
Buses	a	a	a
Motor Cycles	0.065	NA	NA

Source: ABS 2008a. (a) Refer to Table 3.A.15 – 3.A.17.

Table 3.A.19: Evaporative emission factors for road vehicles using automotive gasoline

Vehicle Type	Emission Factor (g/km)	
	Hot Soak and Diurnal Emissions (FH _{ij}) ^a	Running Losses (FR _{ij}) ^b
Passenger Cars^c		
Post 1985	0.38	0.9
1976–1985	0.96	0.9
Pre-1976	1.92	0.9
Light Commercial Vehicles	1.13	0.19
Medium Trucks	2.24	0.26
Heavy Trucks	2.75	0.29
Buses	2.24	0.20
Motorcycles	0.76	0.0

Sources: (a) Carnovale et al. (1991); (b) OECD (1991); (c) Calculated with an RVP of 11.0 psi

Table 3.A.20: CO₂ emission factor for coke

Year	Emission Factor (Gg/ PJ)
1990	105.9
1991	105.9
1992	106
1993	106
1994	105.9
1995	106
1996	105.9
1997	106
1998	106
1999	106
2000	106
2001	105.9
2002	106.6
2003	105.5
2004	106.9
2005	107.7
2006	108.4
2007	108.9
2008	108.9
2009	109.9

Source: Determined using a carbon balance of the coke oven process.

Table 3.A.21: NMVOC emission factors for service station storage and transfer operations

Region	Population (million) ^a	Emission factor (kg per kl distributed) ^b
Sydney Statistical Region ^c	3.67	0.16
Port Phillip Control Region ^d	3.39	0.16
Other	10.22	1.00
Australia ^e	17.28	0.66

Sources: (a) Australian Bureau of Statistics, Census 1991. (b) Filling losses and underground-tank breathing. (c) Environment Protection Authority NSW (1995). (d) Melbourne, Geelong and Westernport Regions, Environment Protection Authority Victoria (1991). (e) Population weighted average, all years 1988-1994.

Table 3.A.22: NMVOC emission factors for bulk fuel storage facilities

Region	Population (million) ^a	Emission factor (kg per kl distributed) ^b
Melbourne/Sydney Region ^c	7.06	0.48
Other ^d	10.22	1.49
Australia ^e	17.28	1.08

Sources: (a) Australian Bureau of Statistics, Census 1991. (b) Storage and working losses (c) Assume emission factors in Melbourne (Environment Protection Authority Victoria, (1991) and Sydney are similar because control regulations are identical (d) From Australian Environment Council (1988) data for regions outside Melbourne and Sydney. (e) Population weighted average, all years 1988-1994.

4. INDUSTRIAL PROCESSES

4.1 OVERVIEW

Total net emissions estimated from industrial processes were 29.6 Mt CO₂-e in 2009, or 5.4% of net national emissions (excluding LULUCF) (Table 4.1).

Table 4.1: Industrial processes sector CO₂-e emissions, 2009

Greenhouse gas source and sink categories	CO ₂ -e emissions (Gg)				
	CO ₂	CH ₄	N ₂ O	HFC/PFC/SF ₆	Total
2 INDUSTRIAL PROCESSES	19,555	62	3,383	6,618	29,617
A Mineral products	6,507	NA	NA	NA	6,507
B Chemical industry	3,180	12	3,369	NA	6,561
C Metal production	9,697	50	14	308	10,069
D Other production	170	NA	NA	NA	170
E Production of halocarbons and sulphur hexafluoride	NO	NO	NO	NO	NO
F Consumption of halocarbons and sulphur hexafluoride	NA	NA	NA	6,310	6,310

Metal production contributed 34.0% (10.1 Mt CO₂-e) of the sector's emissions, *mineral products* contributed 22.0% (6.5 Mt CO₂-e), *chemical industries* contributed 22.2% (6.6 Mt CO₂-e), the *consumption of HFCs and SF₆* contributed 21.3% (6.3 Mt CO₂-e) and *food and drink* contributed 0.6% (0.2 Mt CO₂-e).

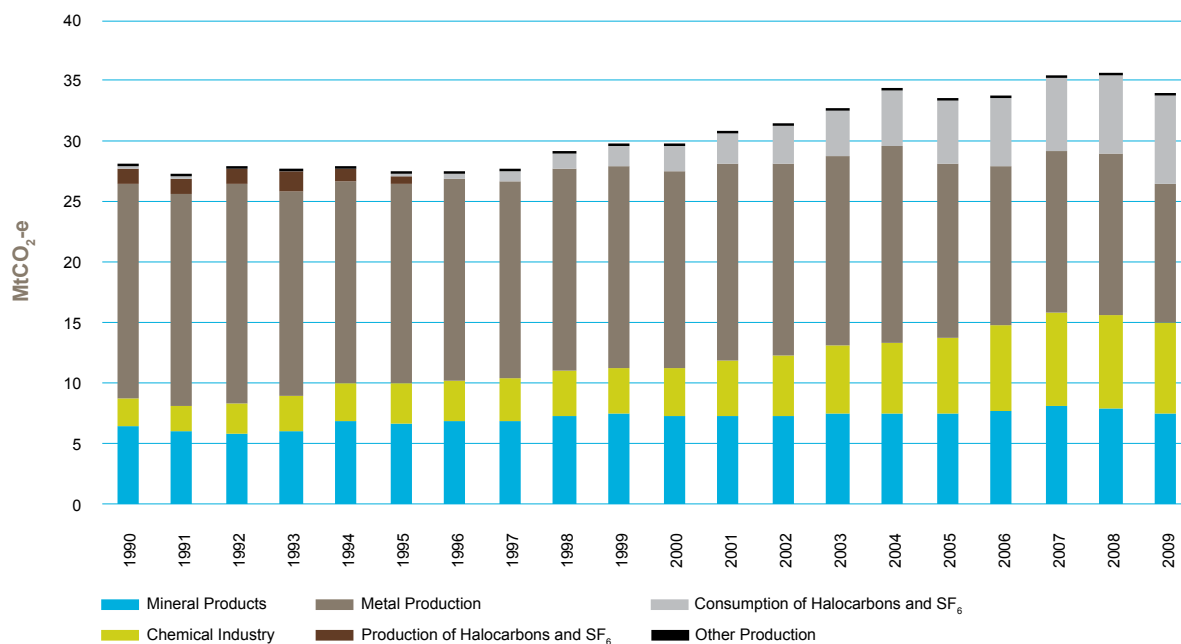
The main gas emitted by industrial processes is carbon dioxide, contributing 66.0% (19.6 Mt) of the sector's emissions in 2009. PFCs contributed 1.0% (0.3 Mt), HFCs contributed 21.1% (6.2 Mt), SF₆ contributed 0.2% (0.1 Mt), nitrous oxide contributed 11.4% (3.4 Mt), and methane 0.2% (0.1 Mt).

Trends

Net emissions from *industrial processes* increased by 21.0% (5.1 Mt CO₂-e) from 1990 to 2009, and decreased by 5.0% (1.6 Mt CO₂-e) between 2008 and 2009 (Figure 4.1). The increases in sectoral emissions observed over the longer term are principally due to growth in emissions associated with the manufacture of chemical products. The decline in emissions from 2008 to 2009 is the largest occurring in industrial processes since 1990 and reflects decreased production levels right across the sector in response to the global economic downturn.

Each source category's contribution to total emissions and to sectoral trends within the Industrial Processes sector between 1990 and 2009 is shown in Figure 4.1.

Figure 4.1: Emissions from industrial processes by subsector, 1990–2009



Clinker Production

Emissions of CO₂ for this source category are dependent on the quantity of clinker produced and this in turn is closely tied to annual growth in the Australian economy. Emissions of CO₂ from clinker production in 2009 were 3829 kt CO₂-e, a 10.6 per cent increase from 1990, while production has increased by 12.6 per cent over the same period. Improvements in industry practices have resulted in lower emissions per unit production.

From 2008 to 2009, total CO₂ emissions decreased by 0.9 per cent. This decrease reflects a decline in production of the same magnitude from 2008 to 2009. Year on year growth in emissions from clinker production is variable and matches fluctuations in clinker production.

Lime Production

Emissions of CO₂ from the production of lime vary year to year according to the quantities of commercial and in-house lime produced. The quantities of lime produced are dependent on the demand for lime within the Australian economy. Total lime production in 2009 was 1,530 kt compared with 1,758 kt in 2008 representing a decrease in production of 13.0 per cent. While this is a large decrease, similar fluctuations have taken place through the time series, for example, lime production decreased 16.7 per cent in 2000. The decline in 2000 is attributed to the fall in demand for minerals processing particularly in the gold sector. These fluctuations illustrate the sensitivity of lime production levels to demand in the resources sector.

Limestone and Dolomite Use

The total CO₂ emissions reported in this source category include emissions from the consumption of carbonates in (calcite, magnesite, dolomite, sodium bicarbonate, potassium carbonate, barium carbonate, lithium carbonate and strontium carbonate), magnesia production, zinc production, ferroalloys production, iron and steel production, ceramics and glass production, soda ash use and production and miscellaneous uses of carbonates. The trend in emissions is heavily influenced by the consumption of limestone which is consumed in greater quantities than any other carbonate. In 2009 limestone consumption had increased by 11.2 per cent from 1990. The year on year growth in limestone consumption, however, has varied from positive to negative throughout the time series with a decrease of 11.6 per cent from 2008 predominantly reflecting a decrease in carbonate consumption in iron and steel production.

Soda Ash Production and Use

Soda ash is produced in Australia by only one company, Penrice Soda Products. Soda ash is also imported into Australia in significant quantities. More than half the soda ash produced is consumed by glass manufacturers. Other important users of soda ash are manufacturers of detergents, soaps and chemicals and the metals and mining industries. Production of soda ash has remained relatively constant while imports of soda ash have experienced large fluctuations and an overall increase in quantities. Emissions for this source category are confidential and are reported under 2.A.3 limestone and dolomite use.

Road Paving with Asphalt

NMVOC emissions from road paving with asphalt were 18.9 kt in 2009 and are increasing over the long term with a rise of 25.8 per cent from 1990 and an increase of 3.2 per cent from 2008.

Chemical Production

There has been a scaling back of chemical products manufacture in 2009 reflecting in combination the effects of the international economic downturn and a gas explosion in Western Australia in October 2008 which affected natural gas supplies for ammonia production in that part of the country. These two events have resulted in a reduction in emissions of carbon dioxide in 2009.

Iron and Steel Production

Emissions per tonne of Iron and Steel produced vary according to changing quantities of reductants used. Emissions from iron and steel production in 2009 were 19.3 per cent lower than 2008. This decline in emissions in 2009 reflects reduced production levels in the iron and steel industry as a result of global economic conditions. For example, raw steel production data available on the Bluescope Steel website (<http://www.bluescopesteel.com/investors/financial/-production-history>) shows a 33 per cent decline in production levels in the 2009 financial year. Similarly, Onesteel report a 21 per cent decline in raw steel production on 2008 (http://www.onesteel.com/images/db_images/annualreports/OneSteel%20Limited%20-%20Annual%20Report%202009.pdf). In both cases, the companies have attributed these declines to the global financial crisis.

The down-turn in emissions during 2005 occurred due to the blast-furnace re-lining activities at the Whyalla steel works. There has been a general declining trend in the Iron and Steel CO₂-e implied emission factor due to the increased use of pulverised coal injection in lieu of coke. All coal used in iron and steel production whether for energy purposes or reductant is accounted for under 1.A.2 Manufacturing Industries and Construction.

Aluminium Production

Emissions from the production of aluminium were slightly lower in 2009 than 2008 owing to a reduction in anode effects resulting in a reduction in PFC emission factors. The aluminium industry was comparatively unaffected by the global financial crisis in 2009. Aluminium production data available on Rio Tinto's website (<http://www.riotinto.com/annualreport2009/production/metals.html>) for its aluminium smelters in Bell Bay, Boyne Island and Tomago shows production levels remained stable in 2009 compared with 2008.

The downward trend in CO₂-e emissions per tonne of aluminium produced since 1990 has occurred as a result of improvements in process control and the resultant reduction in PFC emissions. Any fluctuations in implied emission factors occurring in the latter part of the time-series are the result of small fluctuations in the number of anode effects in the production process occurring due to electricity supply disruptions and potline maintenance. The fall in the PFC implied emission factor between 2005 and 2007 occurred as a result of a smelter upgrade at Hydro Kurri Kurri (conversion of Potline No 1 from side-work to centre-work) and an enhanced emissions performance at the Tomago smelter (AAC 2007).

4.2 OVERVIEW OF SOURCE CATEGORY DESCRIPTION AND METHODOLOGY – INDUSTRIAL PROCESSES

The *industrial processes* sector includes emissions generated from a range of production processes involving *inter alia* the use of carbonates (i.e. limestone, dolomite and magnesite etc); carbon when used as a chemical reductant (e.g. iron and steel or aluminium production); chemical industry processes (e.g. ammonia and nitric acid production) and the production and use of synthetic gases such as halocarbons. Key categories for Australia include emissions from cement production, iron and steel production, aluminium production and the consumption of halocarbons.

For some industries, for example the iron and steel industry, reported emissions are split between the *industrial processes* sector and the *energy* sector depending on the type of process within the industry that generated the emissions.

The Australian methodology for *industrial processes* contains both country specific and IPCC default methodologies and emission factors (Table 4.2). The use of Tier 2 methods indicates a higher level of complexity, data requirements and in-principle accuracy than a Tier 1 method.

In certain sub sectors within industrial processes, activity data are commercial-in-confidence and, due to the direct relationship between activity and emissions, emissions estimates by gas species are also confidential. Where this is the case, it is necessary to aggregate subsectoral emission estimates in order to preserve confidentiality.

Emissions of CO₂ from soda ash production and use (2.A.4) (including sodium bicarbonate production), and magnesia production (2.A.5) have been aggregated with CO₂ from limestone and dolomite use (2.A.3). Carbon dioxide emissions from the use of acetylene (2.B.5) and the production of synthetic rutile and titanium dioxide (2.B.5) have been aggregated with CO₂ from the production of ammonia (2.B.1). This aggregate is reported at the chemical industry sectoral level (2.B.5 – confidential chemical industry emissions). Emissions of N₂O from the use of N₂O in anaesthesia and aerosols (3.D.3) have been aggregated with N₂O from nitric acid production (2.B.2). This aggregate is reported at the chemical industry sectoral level (2.B.5 – confidential chemical industry emissions).

Table 4.2: Summary of methods and emission factors: Industrial Processes

Greenhouse Gas Source and Sink Categories	CO ₂		CH ₄		N ₂ O		HFCs		PFCs		SF ₆	
	Method applied	EF	Method applied	EF	Method applied	EF	Method applied	EF	Method applied	EF	Method applied	EF
2. Industrial Processes												
A. Mineral Products	T2	CS, D	NA	NA	NA	NA						
1. Cement Production	T2	CS	NA	NA	NA	NA						
2. Lime Production	T2	CS	NA	NA	NA	NA						
3. Limestone and Dolomite Use	T2	CS	NA	NA	NA	NA						
4. Soda Ash Production and Use ^(a)	T2/3	CS, D	NA	NA	NA	NA						
6. Road Paving with Asphalt	NA	NA	NA	NA	NA	NA						
7. Other (<i>please specify</i>)												
B. Chemical Industry	CS,T1b, T2	CS,D	T2	D	T3,CS	CS	NA	NA	NA	NA	NA	NA
1. Ammonia Production ^(b)	T1b	CS,D	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
2. Nitric Acid Production ^(b)	NA	NA	NA	NA	T3	CS	NA	NA	NA	NA	NA	NA

Greenhouse Gas Source and Sink Categories	CO ₂		CH ₄		N ₂ O		HFCs		PFCs		SF ₆	
	Method applied	EF	Method applied	EF	Method applied	EF	Method applied	EF	Method applied	EF	Method applied	EF
5. Other (<i>please specify</i>)												
Polymers and other chemicals	NA	NA	T2	D	NA	NA	NA	NA	NA	NA	NA	NA
Synthetic Rutile ^(b)	T2	CS	NA	NA	NA	NA			NA	NA	NA	NA
Confidential emissions ^(c)	CS	CS			CS	CS						
C. Metal Production	T2,T1b	CS	T2	CS	T2	CS			T1c	CS	T2	CS
1. Iron and Steel Production	T1b	CS	T2	CS	T2	CS			NA	NA	NA	NA
3. Aluminium Production	T1b	CS	NA	NA	NA	NA			T1c	CS	NA	NA
4. SF ₆ Used in Aluminium and Magnesium Foundries	NA	NA	NA	NA	NA	NA			NA	NA	T2	CS
5. Other (<i>please specify</i>)												
Other Metals	T2	CS	NA	NA	NA	NA			NA	NA	NA	NA
D. Other Production	CS	CS										
E. Production of Halocarbons and SF₆							T1	D	NA	NA	NA	NA
1. By-product Emissions							NA	NA	NA	NA	NA	NA
Production of HCFC-22							T1	D	NA	NA	NA	NA
Other							NA	NA	NA	NA	NA	NA
F. Consumption of Halocarbons and SF₆							M	D	NA	NA	T1,T2	CS
1. Refrigeration and Air Conditioning Equipment							M	D	NA	NA	NA	NA
2. Foam Blowing							M	D	NA	NA	NA	NA
3. Fire Extinguishers							M	D	NA	NA	NA	NA
4. Aerosols/ MDIs							M	D	NA	NA	NA	NA
5. Solvents							M	D	NA	NA	NA	NA
6. Semiconductor Manufacture							NA	NA	NA	NA	NA	NA
7. Electrical Equipment							NA	NA	NA	NA	T2	CS
8. Other (<i>please specify</i>) (d)							NA	NA	NA	NA	T1	CS
G. Other	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA

Notes: EF = Emission Factor, T1 = Tier 1, T2 = Tier 2, T3 = Tier 3, CS= Country-specific, D= IPCC default, NE = not estimated, NA= not available, NO = not occurring, IE = included elsewhere.

(a) Emissions reported under 2.A.3 limestone and dolomite use; (b) Emissions reported under 2.B.5 confidential chemical industry emissions; (c) Methods for acetylene use and from use of nitrous oxide (3D), methods for other components identified separately; (d) Other uses of SF₆

Data Sources

The inventory for the *industrial processes* sector relies primarily on data collected under the National Greenhouse and Energy Reporting System. The following table summarises the data sources used in compiling the inventory for industrial processes.

Table 4.3: Summary of principal data sources for Industrial Processes 2009

Industrial Processes Sector	Activity data
2.A.1. Cement clinker	Cement Australia, Boral, Adelaide Brighton
2.A.2. Lime	Boral, Adelaide Brighton, Cement Australia, Unimin, Alcan and Queensland Alumina
2.A.3. Limestone and dolomite and other carbonates	Owens-Illinois, CSR, Amcor, Qmag, Causmag, OneSteel, BlueScope Steel, Rio-tinto, Billiton Manganese, Bradken, Sun Metals, BHP Billiton, Xstrata, Nyrstar, Incitec Pivot, Minara Resources, Fletcher Insulation, Thales Australia, Penrice. Qmag, Causmag
2.A.4. Soda ash	Penrice Soda Products
2.A.6. Bitumen	ABARES Commodity Statistics
2.B.1. Ammonia	Incitec, Orica, Wesfarmers CSBP, BHP Billiton, Queensland Nitrates, Burrup Fertilisers
2.B.2. Nitric acid	Orica, Wesfarmers CSBP, Queensland Nitrates
2.B.5. Other chemicals	Dynea W.A, Borden Chemicals, Orica, BP, Shell, Huntsman Chemicals, Dow Chemicals, Qenos, ExxonMobil, Continental Carbon, Cabot Australia, Australian Vinyl, BOC Gases, Airliquide, Caltex, Coogee Chemicals
2.B.5 Synthetic Rutile and Titanium Dioxide	Ti-west, Iluka Resources, Milenium Chemicals
2.C.1. Iron and steel	BlueScope Steel, OneSteel, ABARES
2.C.3. Aluminium	Alcoa, Rio Tinto, Hydro Kurri Kurri, Tomago Aluminium
2.C.4. SF ₆ used in metal production	CSIRO, Brisbane
2.C.5. Other metals	Billiton Manganese, BHP Billiton, Simcoa ABARES Commodity Statistics for various metals
2.D.2. Food and drink	ABS apparent consumption data, Penrice Soda Products, Air Liquide, BOC, Huntsman Chemicals, Incitec, Orica.
2.F Consumption of Halocarbons and SF ₆	Bulk import and pre-charged equipment data reported to the Department of Sustainability, Environment, Water, Population, Arts and Community under the regulations applying under the <i>Ozone Protection and Synthetic Greenhouse Gas Management Act 2003</i>

4.3 SOURCE CATEGORY 2.A MINERAL PRODUCTS

4.3.1 Cement Production

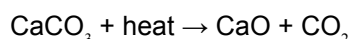
Source Category Description

Cement clinker production is a key category for Australia. Carbon dioxide is produced during the manufacture of portland clinker, which is an intermediate product in the production of cement. Carbon dioxide emissions are essentially proportional to the lime content of the clinker. On exit from the cement kiln, and after cooling, the clinker is ground to a fine powder and up to 5% (by weight) of gypsum or natural anhydrite (that is, forms of calcium sulphate) added to control the setting time of the cement. The finished product is referred to as 'portland' cement.

There are three clinker producers in Australia; Adelaide Brighton, Blue Circle Southern Cement (Boral) and Cement Australia. The production of blended cements, incorporating waste materials from other industries (e.g. slag, fly ash and silica fume), represents a significant portion (approximately 20%) of the total cement manufacturing market in Australia. According to the Cement Industry Federation (CIF 2003), the proportion of waste materials added to cement varies significantly and may range from 10% to 80% (by weight). Blending waste materials with cement significantly reduces the CO₂ emissions per unit of cement produced. According to the CIF 2009 sustainability report, the industry aims to increase the use of supplementary cementitious materials in cement and concrete to at least 29% by 2012.

Methodology

Calcium carbonate (CaCO_3) from calcium rich raw materials such as limestone, chalk and natural cement rock is heated at temperatures of approximately 1500° C in cement kilns to form lime (CaO) and carbon dioxide in a process known as calcination.



Emissions from clinker production are estimated using a tier 2 method.

$$E_{\text{cl}} = [EF_{\text{cl}} \cdot A_{\text{cl}} + EF_{\text{cl}} \cdot F_{\text{ckd}} \cdot A_{\text{ckd}} + EF_{\text{toc}} \cdot (A_{\text{cl}} + A_{\text{ckd}})] \cdot 10^{-6} \dots\dots\dots$$

Carbon dioxide emissions from clinker manufacture are estimated by the application of a country – specific emission factor EF_{cl} , in kilograms of CO_2 released per tonne of clinker produced, to the annual national clinker production A_{cl} .

The country – specific emission factor is the product of the fraction of lime used in the clinker and a constant reflecting the mass of CO_2 released per unit of lime produced. This factor was derived using the World Business Council for Sustainable Development (WBCSD 2005) methodology. Assuming CaO and MgO proportions of 0.66 and 0.015 respectively, based on Ryan and Samarin 1992, leads to an emission factor of 534 kg CO_2 per tonne of clinker.

In addition to the emissions associated with the lime used in the clinker, the methodology accounts for emissions associated with the calcination of cement kiln dust (A_{ckd}) and the quantity of total organic carbon expressed as a proportion of total clinker produced (TOC). F_{ckd} is the degree of calcination of cement kiln dust (ranging from 0% to 100%) and is assumed to be 100% in Australia such that $F_{\text{ckd}} = 1$ (following WBCSD 2005). A_{ckd} is the quantity of cement kiln dust (CKD) produced annually. The emission factor for TOC is taken from WBCSD 2005 (equivalent to 10kg CO_2 per tonne of clinker).

Choice of emission factor

No facility-specific data on emission factors were obtained under the NGER System. The country–specific carbon dioxide emission factor was applied for all facilities and for all years.

Activity data

Data for clinker production for individual facilities were obtained under the NGER System for 2009 and the reporting mechanisms of the Emissions Intensive, Trade Exposed Industries Assistance Program (EITEIs) for 2007 and 2008. Data for the period 1990-2006 were obtained by industry survey (GHD 2009a).

Table 4.4: Australian cement clinker production and emissions 1990, 2000–2009

Year	Clinker production (kt)	Cement Kiln Dust (kt)	Emissions (Gg CO_2)
1990	6,205	160	3,463
2000	6,557	99	3,621
2001	6,425	84	3,541
2002	6,354	58	3,488
2003	6,566	22	3,584
2004	6,492	42	3,555
2005	6,657	79	3,664
2006	7,076	72	3,888
2007	7,254	47	3,973
2008	7,053	48	3, 863
2009	6,986	52	3,829

Source: GHD 2009a, DCCEE Emissions Intensive Trade Exposed Industries Program 2009, DCCEE National Greenhouse and Energy Reporting Scheme 2009

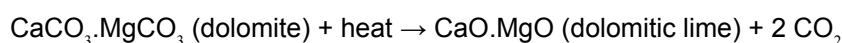
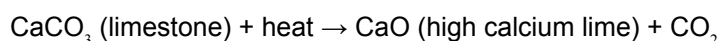
4.3.2 Lime Production (2.A.2)

Source Category Description

Lime is an important chemical having major uses in metallurgy (steel, copper, gold, aluminium and silver), other industrial applications (water softening, pH control, sewage sludge stabilisation), and construction (soil stabilisation, asphalt additive and masonry lime). The producers of commercial lime in Australia include Cement Australia, Blue Circle Southern Cement (Boral), Adelaide Brighton Cement, Unimin, and Queensland Alumina. Rio Tinto Alcan also produces in-house lime for alumina production.

Methodology

Carbon dioxide is produced when either high calcium lime (CaO) or dolomitic lime (CaO.MgO) are manufactured by the calcination of calcium rich raw materials (limestone or dolomite) in a kiln.



Emissions from lime production are estimated using a tier 2 method.

Total CO₂ emissions E_q associated with lime production A_q are estimated as the sum of emissions by facility according to:

$$E_q = \sum A_q \cdot EF_q$$

The emission factor for lime produced is estimated for each facility from a consideration of the molecular weights (56 for CaO, 44 for CO₂) and the composition of the lime products.

Choice of emission factor

Selection of emission factors was undertaken in accordance with the decision tree in section 1.4.1.

Facility-specific emission factors estimated under the EITIEs program have been used for the years 2007, 2008, where available, and extrapolated to 2009. These emission factors take into account the product specifications of the lime produced by each company at each facility as well as an estimate of the calcination of lime kiln dust in the production process. For facilities where facility-specific data was not available on the specifications of the product, the country-specific emission factor of 0.675 t CO₂/t commercial lime produced and 0.730 t CO₂/t non-commercial lime produced was applied reflecting the assumptions that the facility produced commercial lime in accordance with country-specific fractional purities (a fractional purity of 0.86 is used for commercial lime and 0.93 is used for non-commercial lime).

Time series consistency is maintained through the use of a weighted average emission factor of 0.751 t CO₂/t lime produced for the years when individual facility data are not available (1990 – 2006).

Activity data

Data on lime production have been collected under the NGER System for 2009 and the reporting mechanisms of the Emissions Intensive, Trade Exposed Industries Assistance Program (EITEIs) for 2007 and 2008.

Data for the period 1990-2006 were obtained by industry survey (GHD 2009a).

Table 4.5: Lime production emissions 1990, 2000–2009

Year	Total Lime production (kt) ^(a)	Emissions (Gg CO ₂)
1990	1,035	775
2000	1,277	957
2001	1,533	1,150
2002	1,569	1,176
2003	1,593	1,194
2004	1,624	1,217
2005	1,617	1,213
2006	1,467	1,102
2007	1,632	1,225
2008	1,758	1,320
2009	1,530	1,152

Source: GHD 2009a, DCCEE Emissions Intensive Trade Exposed Industries Program 2009, DCCEE National Greenhouse and Energy Reporting Scheme 2009

(a) includes quantities of in-house lime production

4.3.3 Limestone and Dolomite Use (2.A.3)

Source Category Description

Apart from use in clinker and lime production, limestone (CaCO₃), magnesite (MgCO₃) and dolomite (CaMg(CO₃)₂) are basic raw materials that have commercial applications in a number of industries including metallurgy (for example, iron and steel), glass manufacture, ceramics, agriculture, construction, magnesia production and environmental pollution control.

All CO₂ emissions associated with the consumption of carbonates, with the exception of the emissions reported under cement and lime production, are accounted for under Limestone and Dolomite Use. This includes emissions from the use of limestone by the iron and steel, ferroalloys, magnesia, zinc, glass, ceramics production and soda ash production industries. Emissions from the use of limestone in cement and lime production are accounted for under 2.A.1 and 2.A.2 respectively.

Companies using carbonates in their production processes include Owens-Illinois, CSR, Amcor, Qmag, Causmag, OneSteel, BlueScope Steel, Rio-tinto, Billiton Manganese, Bradken, Sun Metals, BHP Billiton, Xstrata, Nyrstar, Incitec Pivot, Minara Resources, Fletcher Insulation, Thales Australia, and Penrice.

To protect confidentiality, the emissions from the production and use of soda ash (2.A.4) have been aggregated with this source category (2.A.3).

To improve the completeness of the inventory emissions from other carbonates known to be supplied to the Australian economy have also been included in this source category (2.A.3). These include sodium bicarbonate, potassium carbonate, barium carbonate, lithium carbonate and strontium carbonate.

Methodology

A Tier 2 method is utilised for the Australian inventory. The mass of CO₂ emitted per unit of limestone EF_{ls}, dolomite EF_d and other carbonates use EF_o is estimated from a consideration of the purity of the raw materials and the stoichiometry of the chemical processes (44 for CO₂; 100 for limestone; 184 for dolomite, 84 for magnesite, 106 for soda ash and 114 for the remaining carbonates). Only the amount of carbonate material used in an application which generates CO₂ is used in the estimation of CO₂ emitted.

Total CO₂ emissions, E, are estimated by summing over each facility the quantity of limestone, A_{ls}, dolomite, A_d, and other carbonate use, A_o, multiplied by their respective country-specific fractional purities and emission factors derived from stoichiometry:

$$E = A_{ls} \cdot F_{ls} \cdot EF_{ls} + A_d \cdot F_d \cdot EF_d + A_o \cdot F_o \cdot EF_o \dots\dots\dots$$

The fractional purities are country specific and include limestone, F_{ls} , 0.90, dolomite F_d , 0.95, and for all other carbonates, 1.00. The emission factors are derived from stoichiometry and are 0.396 t CO₂/t limestone, 0.522 t CO₂/t magnesium carbonate, and 0.453 t CO₂/t dolomite.

Choice of Emission Factor

No facility-specific data on emission factors were obtained under the NGER System. Country-specific carbon dioxide fractional purities and stoichiometric emission factors were applied for all facilities and for all years.

Activity data

Limestone and dolomite consumption data have been collected under the NGER System for 2009 and the reporting mechanisms of the Emissions Intensive, Trade Exposed Industries Assistance Program (EITEIs) for 2007 and 2008.

Data for the period 1990–2006 were obtained by industry survey (GHD 2009a).

The coverage of companies for this source was expanded for this submission due to the mandatory reporting by all companies with emissions above the NGERS reporting thresholds whereas previous voluntary surveys had not identified all consumers of limestone. Where data for a particular facility collected under NGERS was not available in GHD 2009a, time series consistency was maintained by the interpolation of consumption rates reported under NGERS for 2009 to the period between the commencement date for the facility and 2008. These facilities include Bradken, Incitec Pivot, Rio Tinto, Fletcher Insulation, Thales, Sun Metals and Minara Resources.

Table 4.6: Limestone and dolomite consumption and emissions 1990, 2000–2009

Year	Limestone Use (kt)	Dolomite and Other Carbonate Use ^(a) (kt)	Total emissions from the consumption of carbonates (Gg CO ₂) ^(b)
1990	2,176	779	1,300
2000	2,800	1,183	1,740
2001	2,506	1,188	1,639
2002	2,577	1,237	1,717
2003	2,606	1,288	1,738
2004	2,557	1,253	1,705
2005	2,506	1,252	1,691
2006	2,641	1,305	1,772
2007	2,905	1,247	1,883
2008	2,736	1,300	1,808
2009	2,420	940	1,525

Source: EnerGreen Consulting 2009, DCCEE Emissions Intensive Trade Exposed Industries Program 2009, DCCEE National Greenhouse and Energy Reporting Scheme 2009.

(a) Includes magnesite, soda ash, barium carbonate, lithium carbonate, potassium carbonate, strontium carbonate and sodium bicarbonate,

(b) Includes emissions from soda ash production and use

4.3.4 Soda Ash Production and Use (2.A.4)

Source Category Description

A Tier 3 method is utilised for the Australian inventory. Soda ash (sodium carbonate, Na₂CO) is used as a raw material in a large number of industries including glass manufacture, soap and detergents, pulp and paper manufacture and water treatment. Soda ash is produced by one company, Penrice Soda Products, located in South Australia, using the Solvay process. Emissions of carbon dioxide are generated from both

the consumption and production of soda ash. To protect confidentiality, these emissions are aggregated with emissions from limestone and dolomite use reported under 2.A.3.

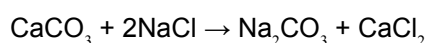
Emissions from the production of soda ash include emissions from the coke used as a reductant. This quantity of coke is deducted from the energy sector as it is a non-energy use of coke and ensures there is no double-counting. Limestone is also consumed in the manufacture of soda ash and both the emissions from the calcination of limestone and the coke used as a reductant are accounted for under limestone and dolomite use (2.A.3).

Sodium bicarbonate (NaHCO_3) is also produced in the Solvay process for soda ash production. When heated or reacted with a weak acid, sodium bicarbonate generates CO_2 . Emissions from the use of sodium bicarbonate in applications where CO_2 is generated have been included in the inventory under food and drink (2.D.2).

Methodology

Soda Ash Production

In the Solvay process, sodium chloride brine, limestone, coke and ammonia are the raw materials in a series of reactions leading to the production of soda ash, sodium bicarbonate and waste products containing calcium carbonate. Ammonia, however, is recycled and only a small amount is lost. The series of reactions involved in the Solvay process may be simply expressed as:



The CO_2 generated in pyrolysis processes is captured, and directed to Solvay precipitating towers for consumption in a mixture of brine (aqueous NaCl) and ammonia. The Solvay process itself is in theory stoichiometrically neutral in relation to CO_2 gas (that is, generation equals uptake), however, in practice a greater amount of CO_2 is generated than can be absorbed in order to optimise the production process.

Emissions from soda ash production are estimated using a tier 2 method.

The estimation of the CO_2 emissions from a stand alone soda ash plant should be based on an overall balance of CO_2 around the whole chemical process. To estimate the excess CO_2 generated during production the carbon in the products and waste materials is deducted from the carbon in the raw materials leaving the excess carbon which is assumed to be entirely converted to CO_2 gas.

$$E_s = [\sum_f \text{CC}_f \cdot A_f + \text{CC}_l \cdot A_l - \sum_p \text{CC}_p \cdot A_p - \sum_w \text{CC}_w \cdot A_w] \cdot 3.664$$

Where: E_s is the emissions of carbon dioxide from the production of soda ash and sodium bicarbonate

CC_f is the carbon content of the fuel consumed

A_f is the mass of fuel consumed (coke)

CC_l is the carbon content of the limestone consumed

A_l is the mass of limestone consumed

CC_p is the carbon content of a product

A_p is the mass of product (soda ash and sodium bicarbonate)

CC_w is the carbon content of the waste products

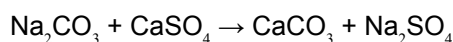
A_w is the mass of waste product (brine mud)

In the first step of the Solvay process limestone is calcined to form lime which is then mixed with water to produce slaked lime for the ammonia recovery step. Any limestone that is not calcined is removed as waste (backstone and grits) from the process and this is deducted from the mass of limestone consumed A_l in the emissions estimate.

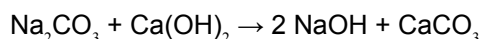
A relatively small amount of waste material containing carbon in the form of calcium carbonate is also deducted from the carbon in the raw materials. The calcium carbonate waste is produced during a brine purification process where calcium and magnesium salts are removed from the brine feedstock. The purification of the brine is achieved through a reaction of soda ash and sodium hydroxide with the calcium and magnesium salts in the brine forming the solids, calcium carbonate and magnesium

hydroxide. Calcium carbonate is also formed in the manufacture of the sodium hydroxide used in these reactions.

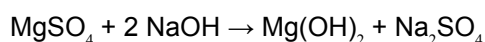
Soda ash is taken from the product stream and diverted to the brine purification process where it reacts with the calcium salts (calcium sulphate) to form calcium carbonate and sodium sulphate:



Sodium hydroxide is manufactured using soda ash (also diverted from the product stream) and slaked lime with calcium carbonate as a waste by-product:



The sodium hydroxide manufactured is then fed into the brine purification process where it reacts with the magnesium salts (magnesium sulphate) to form magnesium hydroxide and sodium sulphate.



In this way the CO₂ absorbed into the soda ash product is then diverted for use in the brine purification process and the manufacture of sodium hydroxide is converted into calcium carbonate. The carbon in the calcium carbonate formed in these reactions is deducted from the raw materials in the calculation of the emissions estimate. The soda ash product used in the brine purification process and manufacture of sodium hydroxide is essentially a non-emissive use of soda ash and the amount used is not included in the total soda ash produced for sale.

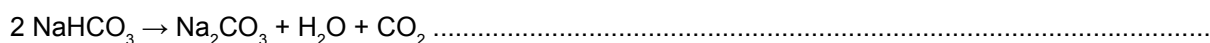
Soda Ash Consumption

A Tier 2 method is utilised for the Australian inventory. Carbon dioxide emissions are associated with the use of soda ash where it is assumed that for each mole of soda ash use, one mole of CO₂ is emitted. The mass of CO₂ emitted from the use of soda ash E_{sau} may be estimated from a consideration of the consumption data A_{sau} and the stoichiometry of the chemical process (where 44.01 is the molecular weight of CO₂ and 105.99 is the molecular weight of Na₂CO₃).

$$E_{\text{sau}} = 0.415 \text{ kg/tonne Na}_2\text{CO}_3 \cdot \sum A_{\text{sau}} \dots\dots\dots$$

Sodium Bicarbonate Consumption

Sodium bicarbonate (NaHCO₃) is also produced in the manufacture of soda ash using the Solvay process. Sodium bicarbonate has a wide range of applications some of which result in the release of CO₂. When sodium bicarbonate is heated or reacted with a weak acid CO₂ is released. Uses of sodium bicarbonate in which CO₂ is generated include leavening agents, pharmaceuticals, stock feed buffer and effervescent salts and beverages. Energreen Consulting 2009 indicates that the proportion of sodium bicarbonate consumption resulting in emissions of CO₂ is 80 per cent This proportion is used to estimate the amount of CO₂ emissions from consumption of sodium bicarbonate. It is assumed that the sodium bicarbonate thermally decomposes in the following reaction:



The mass of CO₂ emitted from the use of sodium bicarbonate E_{sbu} is estimated using consumption data A_{sbu} , the proportion resulting in emissions and the stoichiometry of the chemical process (where 44.01 is the molecular weight of CO₂ and 84.01 is the molecular weight of NaHCO₃).

$$E_{\text{sbu}} = 0.8 \cdot A_{\text{sbu}} \cdot 0.262 \text{ kg/tonne NaHCO}_3 \dots\dots\dots$$

Choice of emission factor

Soda Ash Production

The selection of emission factors was undertaken in accordance with the decision tree in section 1.4.1.

The emission factors for limestone consumption and coke consumption are facility-specific and obtained under NGERs for 2009 and under EITEIs for 2007 and 2008. As there is only one producer, complete coverage for the sector was achieved.

Time series consistency for the entire period 1990-2006 is maintained by the application of the facility-specific factors obtained for the period 2007-2008 to years when no facility data are available.

Soda Ash Consumption

No data on emission factors was obtained under NGERs. The emission factor for soda ash consumption is given by stoichiometry considerations and is the same as the IPCC default.

Activity data

Soda Ash Production

Data on limestone and coke consumption for the purpose of soda ash production were collected under the National Greenhouse and Energy Reporting Scheme for 2009 and the reporting mechanisms of the Emissions Intensive, Trade Exposed Industries Assistance Program (EITEIs) for 2007 and 2008.

Data for limestone and coke consumption for the period 1990-2006 were derived from data for soda ash production obtained by industry survey (Energreen 2009a). Time series consistency was maintained by the application of constant factors of limestone and coke consumption per unit of soda ash production estimated from data available for the period 2007-2009.

Soda Ash Consumption

Data on soda ash consumption were collected under the National Greenhouse and Energy Reporting Scheme for 2009 and the reporting mechanisms of the Emissions Intensive, Trade Exposed Industries Assistance Program (EITEIs) for 2007 and 2008. Data for soda ash consumption for the period 1990-2006 were obtained by industry survey (Energreen 2009a) and data on soda ash imports taken from ABS 2010.

Table 4.7: Soda ash use and emissions

Year	Soda ash use (kt)	Emissions (Gg CO ₂)
1990	450	186
2009	C	C

Source: EnerGreen Consulting 2009, DCCEE National Greenhouse and Energy Reporting Scheme 2009.

4.3.5 Asphalt Roofing (2.A.5)

There is negligible use of asphalt for roofing purposes in Australia.

4.3.6 Road Paving with Asphalt (2.A.6)

Source Category Description

Cutback bitumen is the most common form of primer used in Australia to protect roads from excessive wear. Cutback bitumen primers and primer binders are manufactured from refined bitumen which are 'cutback' (i.e. blended) with petroleum solvents. NMVOC emissions occur during the mixing of bitumen batches, stockpiling, application and curing of the road surface.

Methodology

According to Treadrea (1995), for a system in equilibrium where the quantity of NMVOC used is constant each year and the average temperature conditions do not vary significantly from year to year, the quantity of flux and cutter lost to the atmosphere will be approximated by the quantity used each year.

It is assumed that the quantity of fluxed bitumen is negligible; the fraction of total bitumen consumption used in cutback bitumen is approximately 42 per cent (Australian Asphalt Pavement Association, pers. comm., 1995); and, the quantity of cutter added to the bitumen used in cutback bitumen is equal to 5.4 per cent (Treadrea 1995). Bitumen data are sourced from *Australian Energy Statistics* (ABARE 2010a).

4.3.8 Uncertainties And Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas.

Time series consistency for all sources has been maintained in accordance with the principles established in section 1.4.1

Activity data obtained under NGERS was compared with activity data obtained from the EITEIs program for each facility and with data obtained from GHD and Energreen consulting to ensure the consistent classification of sources and consistency of data.

Where facilities were newly identified from NGERS data as emitting facilities, in category 2.A.3, activity data was interpolated to the facility's commencement date assuming that consumption of limestone and dolomite in previous years was equal to the consumption of limestone and dolomite in 2009 for the each of the new facilities.

Where facility-specific emission factors were identified from NGERS data for particular facilities, in category 2.A.2 and 2.A.4, the observed emission factors were interpolated using a national weighted average emission factor for all years 1990-2006.

4.3.9 Source Specific QA/QC

This source category is covered by the general QA/QC of the greenhouse gas inventory in Section 1.6. Additional source specific quality control checks were undertaken to assess completeness and international comparability.

In order to maintain continuity in the compilation of industrial processes emissions estimates, the DCCEE engaged the external consultant previously used to collect activity data and emission factor information to undertake a quality control assessment of the full time-series of activity data, emission factors and emissions estimates. This work is of particular importance in industrial processes where confidentiality of historical activity data pose some challenges for the assessment of time-series consistency.

Reconciliation between sources of carbonate supply and use in the Australian economy are undertaken to ensure completeness (see Table 4.8). This reconciliation includes limestone used in soda ash production as well as consideration of dolomite, soda ash use, magnesite and other carbonates (barium, lithium, potassium, strontium and sodium bicarbonate).

Table 4.8: Reconciliation of limestone, dolomite, soda ash, magnesite and other carbonates supply and use in the Australian economy, 2009

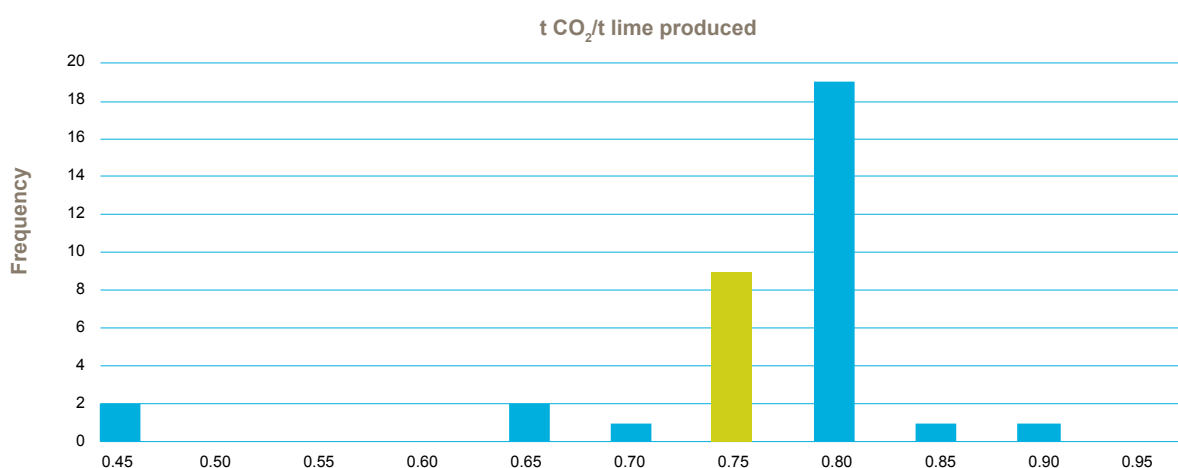
	Raw material ^(d) (kt)	Emissions (Gg CO ₂)	Carbon (kt)
<i>Use</i>			
2.A.1 Cement – clinker	9,457	3,829	1,044
2.A.2 Lime production	2,910	1,152	314
2.A.3 Limestone and Dolomite use	3,940	1,525	416
5.G. Agricultural Liming	2,683	1,073	293
Change in stocks, statistical discrepancy, and residual available for non-pyro processes	-139	–	–
Total Use ^(a)	18,438	7,579	2,067
<i>Supply</i>			
Production	17,914		
Imports	579		
Exports	54		
Total supply ^(b)	18,438		

(a) DCCEE; (b) ABS 2010 (c) Clinker emissions excluding those from the calcination of magnesium carbonates, (d) Includes tonnes of limestone, dolomite, soda ash, magnesite and other carbonates.

Comparisons of implied emission factors and activity data with international data sources are conducted systematically for the Australian inventory.

As the emission factor for lime was subject to revision as a result of the application of facility-level data obtained under EITEIs, the implied emission factor per unit of lime production is reported with the distribution of the implied emission factors of other Annex-1 parties. (Figure 4.2)

Figure 4.2: Lime production implied emission factors for Annex I countries and Australia

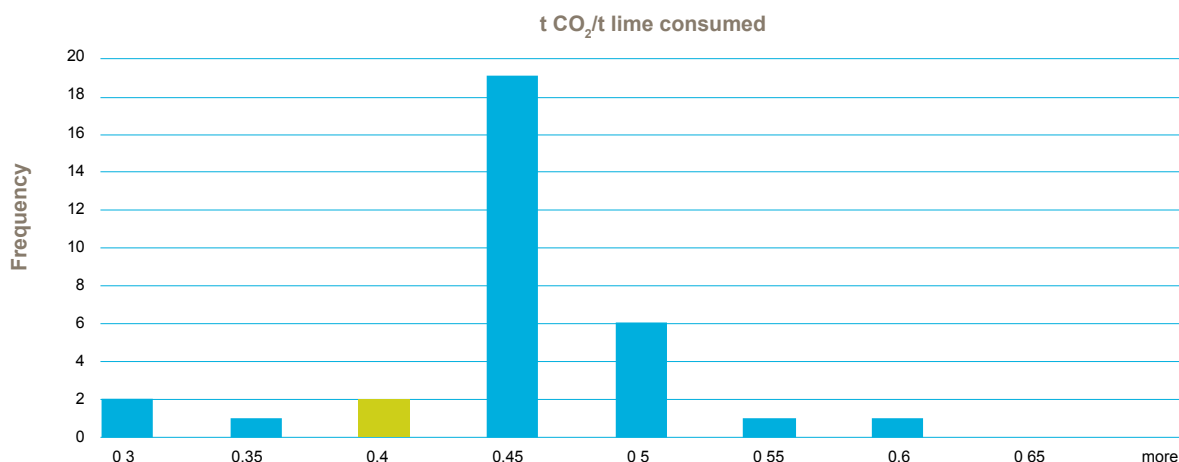


Australia's implied emission factor (IEF) for lime production at the national level ranges between 0.749 t CO₂/tonne of lime produced and 0.753 t CO₂/tonne of lime produced. The IEF fluctuates year on year according to the relative contributions of product from each facility with their own particular product specifications reflecting use of different types of carbonates as well as the relative proportions of commercial and in-house lime produced.

Statistical analysis indicates that the implied emission factor for lime production for Australia is not significantly different to the factors reported by other Annex I parties. Australia's implied emission factor is in line with the Good Practice Guidance default emission factor of 0.75 t CO₂/t high calcium quicklime produced.

The implied emission factor for limestone and dolomite use (2.A.3) for Australia is also reported with the distribution of IEF values for other Annex I countries. Results are shown in Figure 4.3.

Figure 4.3: Limestone and dolomite implied emission factors for Annex I countries and Australia



Australia's limestone and dolomite IEF ranges between 0.410 t CO₂/t carbonate consumed and 0.422 t CO₂/t carbonate consumed. With the availability of facility level data, the national IEF fluctuates according to changes in the relative proportions of each carbonate consumed by individual facilities from year on year.

Statistical analysis indicates that the implied emission factor for limestone and dolomite use for Australia is not significantly different to the factors reported by other Annex I parties. Australia's implied emission factor is below the IPCC default of 0.440 t CO₂/t limestone and 0.477 t CO₂/t dolomite. The revised 1996 IPCC Guidelines suggests the use of a fractional purity of 1 in the absence of country-specific information. In Australia's case, fractional purities of 0.9 for limestone and 0.95 for dolomite are used.

International comparison of activity data is also undertaken. Reported clinker production is consistent with cement production for Australia reported by the United Nations given the high level of use of supplementary cementitious materials (fly ash and granulated blast furnace slag) in Australian cement.

The Mineral Products sector was reviewed independently by an international expert (Tsaranu) in 2007. The review was undertaken applying the same principles governing regular UNFCCC inventory desktop reviews. A number of minor refinements were made to the Mineral Products chapter in response to recommendations made in this review.

4.3.10 Recalculations Since The 2008 Inventory

Recalculations have occurred in the mineral products sector as a result of the use of NGERS and EITEIs activity and emission factor data for the first time. At the sub-sectoral level, the recalculations have occurred in the following areas:

- Clinker production – use of EITEIs activity data for the inventory years 2007 and 2008
- Lime production – use of EITEIs activity data for the inventory years 2007 and 2008 and use of EITEIs emission factors for years 1990-2008
- Other Carbonates (Limestone and Dolomite use) – use of EITEIs carbonate consumption data for the inventory years 2007 and 2008, inclusion of additional carbonates users captured under NGERS for the whole time-series
- Soda ash use – use of EITEIs activity data for the inventory years 2007 and 2008.
- Soda ash production – Use of facility-specific coke oven coke CO₂ emission factor for 1990-2008.

Table 4.9: 2.A.1 Clinker production: recalculation of CO₂-e emissions (Gg), 1990-2008

	2010 Submission	2011 Submission	Change	Change
	Gg CO ₂ -e	Gg CO ₂ -e	Gg CO ₂ -e	%
2.A.1 Clinker Production				
1990	3,463	3,463	–	0.0%
1991	3,183	3,183	–	0.0%
1992	2,923	2,923	–	0.0%
1993	3,005	3,005	–	0.0%
1994	3,484	3,484	–	0.0%
1995	3,358	3,358	–	0.0%
1996	3,244	3,244	–	0.0%
1997	3,185	3,185	–	0.0%
1998	3,475	3,475	–	0.0%
1999	3,520	3,520	–	0.0%
2000	3,621	3,621	–	0.0%
2001	3,541	3,541	–	0.0%
2002	3,488	3,488	–	0.0%
2003	3,584	3,584	–	0.0%
2004	3,555	3,555	–	0.0%
2005	3,664	3,664	–	0.0%
2006	3,888	3,888	–	0.0%
2007	3,983	3,972	11	-0.3%
2008	3,895	3,863	32	-0.8%

Table 4.10 2.A.2 Lime production: recalculation of CO₂-e emissions (Gg), 1990-2008

	2010 Submission	2011 Submission	Change	Change
	Gg CO ₂ -e	Gg CO ₂ -e	Gg CO ₂ -e	%
2.A.2 Lime Production				
1990	706	775	70	9.9%
1991	736	810	73	9.9%
1992	771	848	77	9.9%
1993	805	885	80	10.0%
1994	876	964	88	10.1%
1995	861	948	87	10.1%
1996	990	1,092	101	10.2%
1997	1,009	1,112	103	10.2%
1998	1,155	1,275	120	10.4%
1999	1,043	1,150	107	10.3%
2000	870	957	87	10.0%
2001	1,041	1,150	108	10.4%
2002	1,066	1,176	110	10.3%
2003	1,083	1,194	112	10.3%
2004	1,102	1,217	115	10.4%
2005	1,095	1,213	119	10.8%
2006	991	1,102	111	11.2%
2007	1,008	1,225	217	21.5%
2008	1,118	1,320	202	18.1%

Table 4.11: 2.A.3 Limestone and dolomite use: recalculation of CO₂-e emissions (Gg), 1990-2008

	2010 Submission	2011 Submission	Change	Change
	Gg CO ₂ -e	Gg CO ₂ -e	Gg CO ₂ -e	%
2.A.3 Limestone and Dolomite Use				
1990	1,241	1,300	59	4.8%
1991	1,156	1,216	59	5.1%
1992	1,195	1,254	59	5.0%
1993	1,304	1,363	59	4.5%
1994	1,543	1,603	59	3.8%
1995	1,516	1,575	59	3.9%
1996	1,580	1,639	59	3.7%
1997	1,689	1,748	59	3.5%
1998	1,637	1,694	57	3.5%
1999	1,579	1,857	278	17.6%
2000	1,461	1,740	278	19.0%
2001	1,358	1,639	281	20.7%
2002	1,437	1,717	280	19.5%
2003	1,457	1,738	281	19.3%
2004	1,425	1,705	280	19.6%
2005	1,412	1,691	279	19.8%
2006	1,493	1,772	278	18.6%
2007	1,489	1,883	394	26.4%
2008	1,495	1,808	312	20.9%

4.3.11 Planned Improvements

The methodology used for the estimation of emissions from lime production will be refined to include allowance for the loss of lime kiln dust (LKD) in the production process and a specific emission factor for the production of magnesian and dolomitic lime. Currently, there is no provision for these aspects under NGERs even though the refined approach has been approved under the EITEIs program. For this submission, an implied emission factor derived from EITEIs data has been used in the interim to account for the loss of LKD and the production of magnesian and dolomitic lime. The method approved under EITEIs is expected to be formalised in the NGERs Measurement Determination to enable the collection of LKD and magnesian and dolomitic lime data for future submissions.

4.4 SOURCE CATEGORY 2.B CHEMICAL INDUSTRY

4.4.1 Ammonia Production (2.B.1)

Source Category Description

The overall manufacturing process for ammonia production involves a series of stages to remove impurities such as sulphur, carbon monoxide, carbon dioxide and water from the natural gas feedstock and the generation and reaction of hydrogen and nitrogen. The multi stage process involved in ammonia production (from natural gas feedstock) results in the industrial process emissions of CO₂, NMVOC, and CO in addition to ammonia and sulphur compounds.

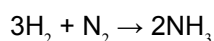
Carbon dioxide emissions from ammonia reflect the use of natural gas for both energy and feedstock uses. In Australia's inventory, only emissions from the use of natural gas as a feedstock are reported in the industrial processes sector. An appropriate deduction has been made in natural gas consumption in the *stationary energy* sector to remove the possibility of double-counting.

A portion of carbon dioxide emissions arising from the production of ammonia are principally recovered for use in the production of urea and food and drink products. Emissions from the production and use of urea are included with the emissions from ammonia in accordance with *good practice*. Emissions from the use of carbon dioxide derived from ammonia production in the food and drink industry are reported under 2.D.2 *food and drink*. The emissions from the ammonia category are aggregated with emissions from the use of acetylene and the production of synthetic rutile and titanium dioxide and included under 2.B.5 *confidential chemical industry emissions*. In 2009, ammonia production contributed around two-thirds of total CO₂ emissions under 2.B Chemicals.

Ammonia is produced by six producers in Australia; Incitec, Orica, Wesfarmers, BHP-Billiton, Queensland Nitrates and Burrup fertilisers. All companies provided natural gas consumption and CO₂ recovery data (where appropriate) for this Inventory under the NGER System. Historical ammonia production data are confidential.

Methodology

A Tier 1b method is utilised for the Australian inventory. Ammonia is manufactured by the catalytic steam reforming of natural gas. Hydrogen from the reformed natural gas and nitrogen from air are compressed at reduced temperatures to form ammonia,



The overall manufacturing process for ammonia production involves a series of stages to remove impurities such as sulphur, carbon monoxide, carbon dioxide and water from the natural gas feedstock and the generation and reaction of hydrogen and nitrogen.

The manufacture of ammonia from the catalytic steam reforming of natural gas is documented to result in emissions of CO₂, NMVOC and CO. While the CO₂ equivalent emissions associated with the use of natural gas are accounted for, data on emissions of NMVOC and CO are not currently available. It is assumed that carbon in natural gas feedstock is converted entirely to CO₂.

The general method for deriving emissions relates a country-specific emission factor EF_i (reported in Table 3.2) to plant specific natural gas consumption data A_i:

$$E_a = \sum A_i \cdot EF_i - R_{\text{food and drink}}$$

R_{food and drink} is CO₂ captured and sold for use in the food and drink industry. Carbon dioxide is captured and used in either the production of urea or the manufacture of food and drink products. The CO₂ recovered for use in urea production is not deducted from CO₂ emissions from ammonia production.

The quantity of CO₂ recovered for use in food and drink applications has been estimated based on information obtained from NGERS reporting facilities for 2010 for which reporters were required to report the recovery of CO₂ specifically for urea production (it is assumed that any CO₂ not recovered and used in urea production is sold to the food and drink industry).

Choice of emission factor

A facility-specific emission factor for one facility reported under the NGER System was used for 2009. This particular facility has reported a facility-specific natural gas emission factor throughout the time-series.

For the remaining six facilities, no facility-specific emission factor information was available. Therefore the default emission factor for the consumption of natural gas as listed in table 3.2 (combined with oxidation factors in table 3.3) was used.

Activity data

Data on fuel consumption, ammonia production and CO₂ capture were obtained under the NGER System for 2009. Data for consumption of fuels were derived from data on production for the period 1990-2008 provided by Energreen 2009 and constant consumption to production factors in order to ensure time series consistency.

4.4.2 Nitric Acid Production (2.B.2)

Source Category Description

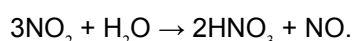
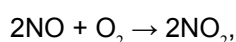
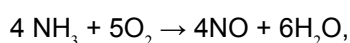
The manufacture of nitric acid (HNO₃) generates N₂O as a by-product of the high temperature catalytic oxidation of ammonia (NH₃). Nitric acid is used as a raw material mainly in the manufacture of nitrogenous agricultural fertiliser.

Nitric acid is produced by three producers in Australia; Wesfarmers, Orica and Queensland Nitrates.

Emissions for the nitric acid category are reported as 'included elsewhere' where the estimates are aggregated with emissions from the use of N₂O in anaesthesia and aerosols and included under 2.B.5 *confidential chemical industry emissions*.

Methodology

A Tier 3 method is utilised for the Australian inventory. Nitric acid production involves three distinct chemical reactions. These are summarised as follows,



Nitric oxide (NO), an intermediate in the manufacture of nitric acid, is documented to readily decompose to N₂O and nitrogen dioxide (NO₂) at high pressures for temperatures in the range of 30 to 50°C.

Facility-specific emission factors for N₂O from nitric acid production EF_n are based on periodic measurements of the off-gas emitted at nitric acid production plants in the Australia. These emission factors are confidential.

The emissions of N₂O, E_n, from the manufacture of nitric acid production A_n is calculated according to:

$$E_n = A_n \cdot EF_n$$

Choice of emission factor

The selection of emission factors was undertaken in accordance with the decision tree in section 1.4.1.

The emission factors for nitric acid production are facility-specific and obtained under the NGER System for 2009. For earlier years, incomplete data on facility-specific emission factors were available from Energreen 2009. Where facility-specific factors were not available, no information about the factors applicable to the remaining facilities were inferred from the Energreen data on the assumption that factors applicable to each facility are technology-specific and independent of each other. In these cases, IPCC good practice default factors were applied in accordance with information available on the applicable technologies (Energreen 2009).

Time series consistency is maintained by the interpolation of the available facility – specific emission factors to the most recent year for which data were available.

Activity data

Data on nitric acid production for individual facilities were collected under the NGER System for 2009.

Data for nitric acid production for the period 1990-2008 were provided by Energreen 2009.

4.4.3 Adipic Acid Production (2.B.3)

There is no adipic acid production occurring in Australia.

4.4.4 Carbide Production (2.B.4)

Silicon carbide and calcium carbide are not produced in Australia. Minor quantities of acetylene are produced from imported calcium carbide and used in welding applications. Data are reported by one company, BOC. Emissions for this category are reported as 'included elsewhere' where the estimates have been aggregated with emissions from the production of ammonia and synthetic rutile and titanium dioxide and included in 2.B.5 *confidential chemical industry emissions*.

4.4.5 Other (2.B.5) Synthetic Rutile and Titanium Dioxide

Source Category Description

Rutile (titanium dioxide) is naturally occurring in Australia. Synthetic rutile can be produced from naturally occurring ilmenite using coal reductant. The rutile is then refined using petroleum coke reductant to produce Titanium Dioxide (TiO₂).

Titanium dioxide is a white pigment which is used in paint manufacture, paper, plastics, rubber, ceramics, fabrics, floor covering, printing ink, and other miscellaneous uses). Titanium dioxide products are referred to generically as titanium dioxide unless there is a need to make a distinction between the products.

Synthetic rutile is produced in Australia by Iluka Resources and Ti-West whilst TiO₂ is produced by Ti-west and Millenium Chemicals.

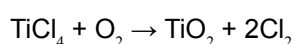
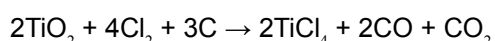
The use of coal and petroleum coke as reductants in the synthetic rutile and TiO₂ production processes are accounted for in the industrial processes sectors. These reductant quantities have been removed from the stationary energy sector to eliminate the possibility of a double-count.

Emissions data for synthetic rutile are reported as 'included elsewhere' where the estimates are aggregated with emissions from the production of ammonia and the use of acetylene and included under 2.B.5 *confidential chemical industry emissions*.

Methodology

A Tier 2 method is utilised for the Australian inventory. The processes that are used in the production of TiO₂ in Australia that lead to process greenhouse gas emissions are synthetic rutile production using the Becher process, and rutile TiO₂ production via the chloride route.

The Becher process reduces the iron oxide in ilmenite to metallic iron and then reoxidises it to iron oxide, and in the process separates out the titanium dioxide as synthetic rutile of about 91 to 93 per cent purity. Rutile TiO₂ is produced through the carbothermal chlorination of rutile ore or synthetic rutile to produce titanium tetrachloride (TiCl₄) and oxidation of the TiCl₄ vapours to TiO₂ according to the following reactions (Kirk-Othmer, 1999; p.2017):



Based on stoichiometry and assuming complete conversion of the input C to CO₂ through further conversion of CO in excess air, the CO₂ emission factor cannot be less than 0.826 tonnes of CO₂ per tonne of TiO₂ (based on 1.5 moles of CO₂ per mole of TiO₂).

Emissions from rutile and TiO_2 respectively may be calculated by:

$$\text{CO}_2 \text{ Emissions} = \sum \text{EF}_i \cdot A_i$$

Where EF_i is the emission factor for fuel type i and A_i is the quantity of fuel type i consumed as a reductant.

Choice of emission factor

No facility-specific information on emission factors from NGERS has been used in this inventory. Country-specific emission factors are applied to the quantities of black coal and petroleum coke consumed in the synthetic rutile and titanium dioxide production processes.

Activity data

Data on synthetic rutile and TiO_2 production, black coal and petroleum coke consumption were obtained under NGERS from the three manufacturers, Illuka, TiWest and Millenium Chemicals. For the inventory years 2007 and 2008, activity data collected under EITEIs has been used.

Data for consumption of coal and petroleum coke were derived from data on production for the period 1990-2006 provided by Energreen 2009 and constant consumption to production factors in order to ensure time series consistency.

4.4.6 Other (2.B.5) Polymers And Other Chemicals

Source Category Description

The manufacture of organic chemicals results in process emissions of NMVOC. Other gases such as CO_2 , CH_4 , N_2O , NO_x and CO may also be generated depending on the manufacturing process.

Complete time series of emissions of CH_4 and NMVOCs are included in the inventory for methanol, butadiene, carbon black, ethyl benzene, ethylene, ethylene oxide, formaldehyde, HDPE, LDPE, LLDPE, propylene, polypropylene, polystyrene, styrene, polyvinyl chloride, and styrene butadiene rubber. Disaggregated production and emissions data for these sources are confidential. Emissions estimates are aggregated at the polymers and other chemicals source category level.

There are approximately 15 companies producing a large range of polymers and other chemicals in Australia. Companies include Dynea W.A, Borden Chemicals, Orica, BP, Shell, Huntsman Chemicals, Dow Chemicals, Qenos, ExxonMobil, Continental Carbon, Koppers, Australian Vinyl, BOC Gases, Airliquide, Caltex, and Nuplex.

Methanol is produced by one plant owned by Coogee Chemicals which has been operating since 1994 with an annual production capacity of 80 kt (Coogee Chemicals website: http://www.coogee.com.au/op_meth.html).

Dichloroethylene is used to produce vinyl chloride monomer (VCM) which is used to produce polyvinyl chloride (PVC) resin. All PVC resin manufactured in Australia is produced from imported VCM. (<http://www.vinyl.org.au/Manufacturingprocess>). Dichloroethylene production does not occur in Australia.

Methodology

A Tier 2 method is utilised for the Australian inventory. Emissions from miscellaneous organic chemical manufacture are dependent on the level of activity and extent of emission control and estimated according to equation:

$$E_{ij} = (A_j \times EF_{ij})/106$$

where: E_{ij} is the process emission (Gg per year) of gas i from industrial sub-sector j

A_j is the amount of activity (production or consumption) of material in industrial sector j (tonnes per year unless)

EF_{ij} is the emission factor associated with gas i per unit of activity in industrial sector j (kg per tonne) – see table 4.12

The divisor 106 is a factor for converting kg to Gg (ktonnes) (1,000,000kg = 1 Gg)

Table 4.12: Emission factors for organic chemicals

Subsector	CO ₂ (kg/tonne)	CH ₄ (kg/tonne)	NM VOC (kg/tonne)
Acetylene ^(a)	3 384 kg CO ₂ per tonne C ₂ H ₂ used		
Butadiene			1.5
Carbon black		0.11	0.5
Ethyl benzene			0.03
Ethylene		0.03	0.25-1.5
Ethylene oxide			0.069
Formaldehyde			9.2
HDPE			1.5
LDPE and LLDPE			1.5
Methanol ^(b)		0.002	
Propylene			1.5
Polypropylene			1.5
Polystyrene ^(b)			0.1 – 5.4
Styrene ^(b)		4	18
Styrene butadiene rubber		1.5	1.5
Polyvinyl chloride		8.5	8.5

Source: EnerGreen 2009. (a) Based on stoichiometry. (b) IPCC 1997.

4.4.7 Uncertainties And Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas.

Activity data obtained under NGERS was compared with activity data obtained from the EITEIS program for each facility and with data obtained from GHD and Energreen consulting to ensure the consistent classification of sources and consistency of data.

No facilities were newly identified from NGERS data as emitting facilities for this category.

Where facility-specific emission factors were identified from NGERS data for particular facilities, in category 2.B.2, the reported emission factors for 2007, 2008 and 2009 were interpolated for each facility to the most recent year for which data were available.

4.4.8 Source Specific QA/QC

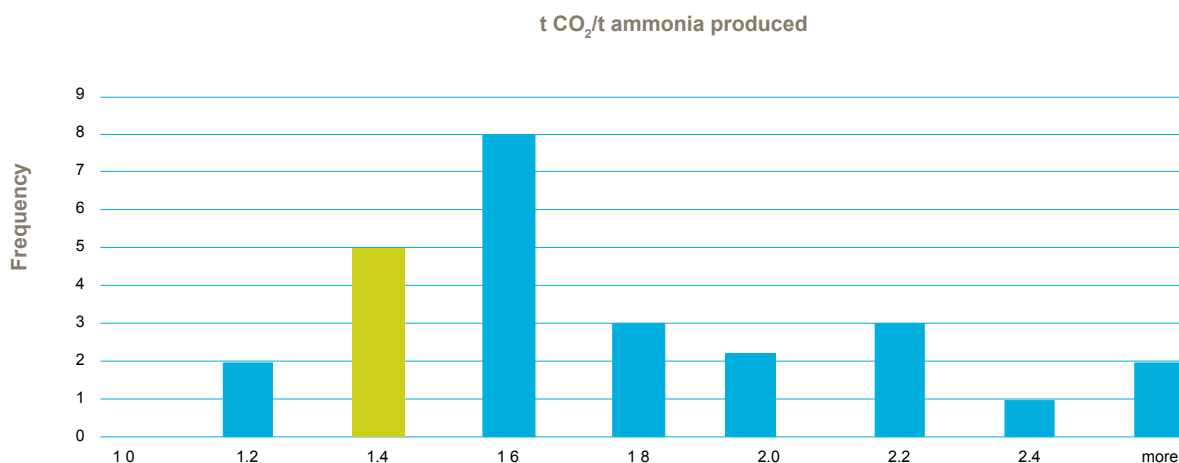
This source category is covered by the general QA/QC of the greenhouse gas inventory in Section 1.6. Additional source specific quality control checks were undertaken to assess international comparability.

In order to maintain continuity in the compilation of industrial processes emissions estimates, the DCCEE engaged the external consultant previously used to collect activity data and emission factor information to undertake a quality control assessment of the full time-series of activity data, emission factors and emissions estimates. This work is of particular importance in industrial processes where confidentiality of historical activity data pose some challenges for the assessment of time-series consistency.

The implied emission factor per unit of production for Australia's inventory was compared with the implied emission factors for other Annex I parties in the cases of ammonia and nitric acid production. The factors for Australia were found to be not significantly different to the factors reported by other Annex I parties. The results of this comparison are presented below.

The quantity of CO₂ generated per tonne of ammonia produced has been compared with that of Annex I parties reporting emissions from ammonia production. The results of this comparison are shown in Figure 4.4.

Figure 4.4: Ammonia implied emission factors for Annex I countries and Australia



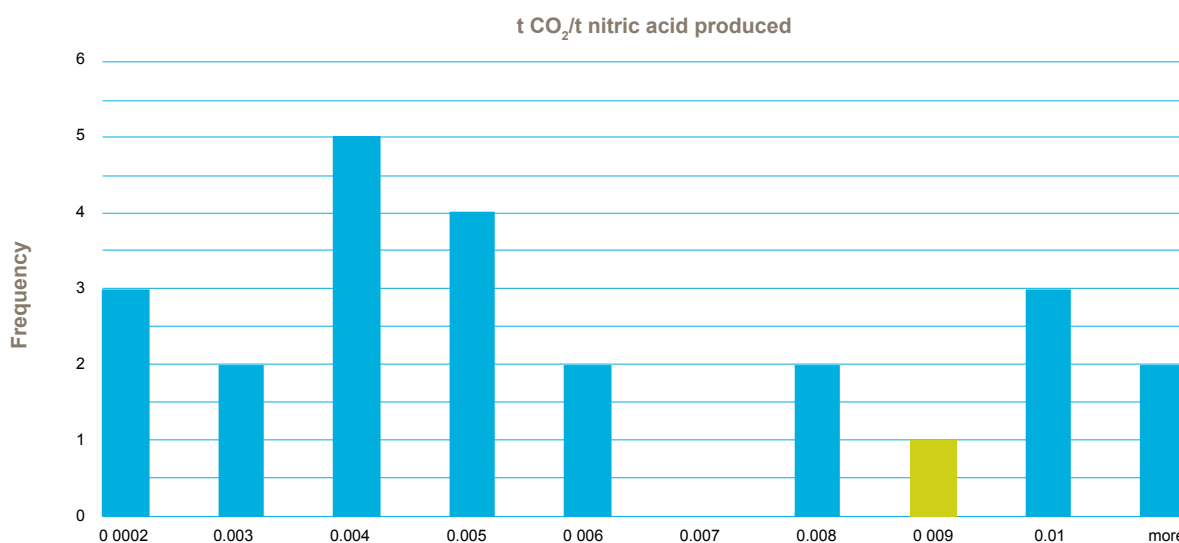
The implied emission factor for ammonia production for Australia ranges between 1.384 t CO₂ per tonne of ammonia produced and 1.544 t CO₂ per tonne of ammonia produced. The implied emission factor fluctuates year on year according to fluctuations in ammonia production levels of individual facilities.

In general, Australia's implied emission factor is broadly consistent with the default values listed in the Revised 1996 IPCC Guidelines of 1.5 – 1.6. The 2006 IPCC Guidelines lists a range of default “total fuel requirements” (including natural gas consumed for energy purposes) by production process whereas under NGERs, facilities must report feedstock and fuel use separately. This specific IP / non-IP split in activity data partly explains the difference between Australia's implied emission factor and the IPCC defaults. The specific ammonia production technology mix in Australia will also cause differences between parties and the default IPCC values.

Statistical analysis indicates that the implied emission factor for ammonia production for Australia is not significantly different to the factors reported by other Annex I parties.

The quantity of N₂O emitted per tonne of nitric acid produced has also been compared with that for Annex I parties. The results of this comparison are shown in Figure 4.5.

Figure 4.5: Nitric acid implied emission factors for Annex I countries and Australia



The implied emission factor for nitric acid production for Australia ranges between 0.006 t N₂O per tonne of nitric acid produced and 0.010 t N₂O per tonne of nitric acid produced. The implied emission factor fluctuates year on year according to fluctuations in nitric acid production levels at individual facilities as emissions at individual facilities are highly technology-specific with three main types of production plants and differing levels of abatement technology in place.

Statistical analysis indicates that the implied emission factor for nitric acid production for Australia is not significantly different to the factors reported by other Annex I parties.

The *chemical products* category was reviewed independently by an international expert (Tsaranu) in 2007. The review was undertaken applying the same principles governing regular UNFCCC inventory desktop reviews. A number of minor refinements were made to the *chemical products* chapter in response to recommendations made in this review.

4.4.9 Recalculations Since The 2008 Inventory

Recalculations have occurred in the Chemical Products sector as a result of the use of NGERS and EITEIs activity and emission factor data for the first time. At the sub-sectoral level, the recalculations have occurred in the following areas:

- ammonia production: revision to ammonia production natural gas consumption and CO₂ recovery data – NGERS data has enabled a better split between the consumption of natural gas as feedstock and natural gas consumed for other purposes. Deductions of natural gas consumption made from stationary energy have been amended accordingly;
- nitric acid: the use of facility-specific emission factors; and
- synthetic rutile and titanium dioxide: use of activity data collected under EITEIs for 2007 and 2008.

The impact of the recalculations on the time-series of emissions estimates from the chemical industry is shown in table 4.13.

Table 4.13: 2.B Chemicals: recalculation of total CO₂-e emissions (Gg), 1990-2008

	2010 Submission	2011 Submission	Change	Change
	Gg CO ₂ -e	Gg CO ₂ -e	Gg CO ₂ -e	%
2.B Chemicals				
1990	2,187	2,053	- 134	-6.1%
1991	2,033	1,885	- 148	-7.3%
1992	2,369	2,218	- 151	-6.4%
1993	2,796	2,593	- 204	-7.3%
1994	2,859	2,612	- 247	-8.6%
1995	3,027	2,785	- 242	-8.0%
1996	3,196	2,968	- 228	-7.1%
1997	3,200	3,003	- 197	-6.1%
1998	3,521	3,264	- 257	-7.3%
1999	3,434	3,211	- 223	-6.5%
2000	3,713	3,475	- 237	-6.4%
2001	4,156	4,098	- 57	-1.4%
2002	4,308	4,272	- 36	-0.8%
2003	5,079	4,920	- 160	-3.1%
2004	5,306	5,143	- 164	-3.1%
2005	5,488	5,504	16	0.3%
2006	7,103	6,158	- 945	-13.3%
2007	7,691	6,780	- 910	-11.8%
2008	7,207	6,672	- 535	-7.4%

4.4.10 Planned Improvements

Further work is planned to clarify the types of manufacturing and abatement technologies used in the nitric acid industry in Australia and how these have changed since 1990. This information will be used to further assess the emissions trends observed in nitric acid production.

As a result of the in-country review of Australia's 2008 Inventory, the Expert Review Team encouraged Australia to explore the possibility of disaggregating emissions from ammonia production. There remain confidentiality concerns in this sector as Australia plans to explore ways of disaggregating ammonia data while still protecting other confidential chemical industry data aggregated with ammonia. Further industry consultation will be required before the whole time-series of emissions could be published. In the interim, Australia has provided further implied emission factor information to assist in the assessment of this sector.

4.5 SOURCE CATEGORY 2.C METAL PRODUCTION

4.5.1 Iron And Steel Production (2.C.1)

Source Category Description

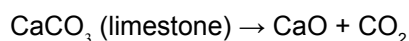
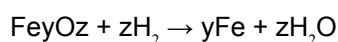
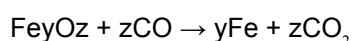
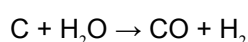
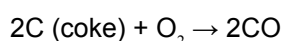
Iron and Steel production is a key source in the Australian inventory. Emission sources relate to the in-house production of metallurgical coke and lime, the use of limestone and dolomite as flux in iron, steel and ferro-alloy production and fugitive gas leaks associated with the distribution of coke oven gas and other products within industrial premises. In-house lime production as well as limestone and dolomite use is accounted for under 2.A.2 and 2.A.3.

Metallurgical coke is an essential material in iron and steel production where it serves a number of major functions including the provision of a porous support for furnace ingredients, as a combustion ingredient producing the reducing atmosphere required for ore refinement and as a chemical reductant. Emissions from the use of coke as a reductant are reported in this category. Emissions from the production of coke are reported under category 1.A.1 while the emissions generated by the combustion of coke oven gas to produce energy are reported under the stationary energy category 1.A.2.

There are two major producers of iron and steel in Australia; Onesteel and Blue Scope. Integrated iron and steel production occurs in New South Wales and South Australia. A new plant in Western Australia has commenced operation. This plant is operated by HIs melt. The plant uses non-coking coal to produce pig iron. The coal reductant used in the Hismelt process and pulverised coal used in other iron and steel production processes are accounted for in 1.A.1. A hot briquetted iron (HBI) plant that used natural gas as a reductant in Western Australia closed in 2005.

Methodology

A Tier 1b method is utilised for CO₂ and Tier 2 for non-CO₂ in the Australian inventory. The manufacture of iron involves the high temperature reduction of iron-bearing materials in a blast furnace. The blast furnace is essentially a large chemical reactor charged with iron ore, coke and limestone/dolomite to produce hot metal or 'pig iron' which is converted into steel typically by injecting oxygen gas through a charge of scrap and the molten iron. During the process, lime is added to remove impurities and provide a slag of the desired basicity. The chemical reactions that occur in the blast furnace to produce molten iron (Fe as shown in the equations) may be summarised as follows,



Coke

The emissions from the use of coke as a reductant are estimated according to equations 3.1 and 3.2 reported in chapter 3, using emission factors reported in the following tables.

Table 4.14: Carbon dioxide emission factors for iron and steel

Fuel Type	P Oxidation Factor (%)	F Emission Factor (Gg/PJ) ^(c)
Coke	98 ^(a)	109.9
Natural Gas	99.5 ^(b)	51.4

(a) IPCC (1997) default value. (b) IPCC (1997) default value. (c) the CO₂ emission factor for coke is derived from a carbon balance calculation conducted for the coke oven process. The natural gas emission factor is provided by the Australian Gas Association.

Table 4.15: Non-carbon dioxide emission factors for iron and steel

Fuel Type	F: Emission Factors (Mg/PJ)					
	CH ₄	N ₂ O	CO	NO _x	NMVOC	SO ₂
Coke	1.07	0.80	82.19	450.47	1.00	370
Natural Gas	0.94	0.09	68.20	914.60	1.10	2.3

The raw steel produced contains carbon, the ultimate source of which is fossil carbon from the coal input to coke ovens. Since steel is a long-lived product, this is a form of carbon sequestration. On advice from the industry (Energreen Consulting, pers. comm. 2007), the average carbon content of raw steel is estimated to be 0.17%.

Fugitive Emissions

In addition to the estimation of emissions from the use of coke and gas as reductants, a process emission factor is established for methane from integrated iron and steel production (0.44 kg CH₄/tonne of crude steel produced) to reflect mainly sources of fugitive emissions. The estimated methane emission factor is based on experimental data and engineering calculations conducted at the plant owned by BlueScope Steel by BHP (pers. comm. 2000) for its major Australian integrated iron and steelworks. Process emission sources considered include the in-plant distribution of coke oven gas and natural gas, leakage from coke ovens and the bleeding of unflared blast furnace gas to the atmosphere. By comparison with fugitive emissions from the in-plant distribution of coke oven gas, emissions of methane associated with leakage from coke ovens and the bleeding of unflared gas from blast furnaces are estimated to be of minor significance.

Activity data

Activity data for coke consumption in the production of iron and steel are obtained from ABARES energy statistics (ABARE 2010a). For 1990 – 2008, crude steel production has been sourced directly from companies (Energreen 2009). The crude steel production for 2009 was derived by indexing the crude steel production in 2008 to the changes in coke consumption in 2009.

Table 4.16: Australian steel production, coke consumption and emissions 1990, 2000–2009

Year	Steel production (kt)	Coke consumption (kt)	Hot Briquetted Iron production (kt)	Natural Gas consumption (PJ)	Total emissions from iron and steel production (Gg CO ₂ -e)
1990	6,223	3,232	NO	NO	9,099
2000	6,345	3,407	558	6	9,883
2001	6,027	3,341	1,223	22	9,144
2002	5,933	2,730	1,142	23	8,911
2003	6,282	3,611	1,670	34	8,671
2004	6,312	2,681	1,592	32	9,261
2005	5,977	3,696	NO	NO	7,356
2006	6,560	2,530	NO	NO	7,293
2007	6,600	3,548	NO	NO	7,433
2008	6,597	2,604	NO	NO	7,539
2009	5,274	3,407	NO	NO	6,084

Source: ABARES, DCCEE.

4.5.2 Ferroalloys Production (2.C.2)

Source Category Description

Emissions from the consumption of fossil fuels when used as reductants, or when used to produce carbon anodes on-site, or as carbon anodes are estimated under this category.

There is one company producing ferroalloys in Australia consuming black coal, coking coal, coke oven coke, petroleum coke and limestone in the process.

The availability of NGERS data on reductant consumption in the production of ferroalloys has enabled reductant emissions from this source to be estimated for the first time in this submission. These emissions are reported under 2.C.4 'Other Metals' to protect confidentiality of data. An equivalent deduction has been made in stationary energy to ensure there is no double counting or omission of emissions. The use of limestone in the production of ferroalloys is reported under 2.A.3 'Limestone and Dolomite Use'.

Methodology

Emissions from the consumption of reductants in the production of ferro-alloy metals have been estimated using a tier 2 method. Emissions from the use of reductants in the production of ferroalloys are estimated by the application of a country-specific emission factors in Table 3.2 and the oxidation factors in Table 3.3 to the quantity of each reductant used.

Choice of emission factor

Emission factors have been selected in accordance with the decision tree in section 1.4.1. No information on facility-specific emission factors were available under the NGER System. Time series consistency has been maintained by the application of values for emission factors for 2009 for the period 1990-2008.

Activity data

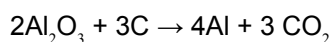
Data on fuel consumed as reductants for the purpose of production of ferro-alloy metals have been collected under the NGER System for 2009. For the years 1990-2008, this level of fuel consumption has been assumed to be constant.

4.5.3 Aluminium Production (2.C.3)

Source Category Description

Aluminium is a key source in the Australian inventory. Emissions from the consumption of fuels in the production of carbon anodes on-site, or as carbon anodes, are estimated for this source. Additional perfluorocarbon emissions resulting from process upsets are also reported under this category.

Aluminium is produced by the electrolysis of alumina in a series of complex electrode reactions. The overall reaction results in aluminium being produced at the cathode and carbon dioxide at the anode:



The electrolysis process is conducted in carbon-lined steel pots containing high purity carbon anodes. The cell electrolyte consists of a molten bath of cryolite (Na_3AlF_6) to which varying proportions of aluminium fluoride, calcium fluoride or lithium fluoride may be added to lower the melting point, decrease the density of the electrolyte and improve energy efficiency.

Carbon dioxide is primarily formed by the chemical reaction of oxygen (produced in the electrolysis process) with the carbon anode. During the electrolysis of alumina to aluminium, some of the CO_2 formed at the anode may be reduced to CO by a secondary reaction involving particles of aluminium or sodium. Grjotheim and Welch (1980) report that for a typical 150kAmp pre-baked cell, the anode gas consists of 70–85% CO_2 with the balance (15–30%) as CO. Measurements conducted by the ADC at several Australian smelters indicate that approximately 10% of the anode gas (by weight) consists of CO. On contact with air, the majority of the CO in anode gas is burnt to CO_2 immediately above the electrolyte.

The perfluorinated carbon compounds (PFC), tetrafluoromethane (CF₄) and hexafluoroethane (C₂F₆) are powerful greenhouse gases which are generated during the so-called anode effect in the production of aluminium. The anode effect is characterised by an increase in cell voltage as a result of the cryolite bath becoming deficient in alumina.

In Australia, bauxite is refined to alumina in Western Australia (WA), Queensland (Qld) and the Northern Territory (NT). The in-house production of lime at alumina refineries in Qld and NT represents an industrial process source of CO₂ emissions, which are accounted for under 2.A.2.

There are four companies operating aluminium smelters in Australia; Alcoa, Tomago Aluminium Rio Tinto and Hydro Kurri Kurri.

Methodology

Carbon dioxide emitted during the consumption of carbon anodes is reported as if all the carbon is oxidised to CO₂. Emissions from the production of carbon anodes for use in aluminium production are estimated on the basis of the quantities of coal tar, petroleum coke and coke oven coke consumed in the production process and plant-specific emission factors. Carbon dioxide emissions are derived using the equation:

$$E_{al} = A_i \cdot EC_i \cdot EF_i$$

Where

A_i is the quantity of fuel type i consumed in the production of anodes

EC_i is the energy content of each fuel type i

EF_i is the CO₂ emission factor for each fuel type i

Facility specific PFC emission factors have been estimated in accordance with accepted international measurement protocols (International Aluminium Institute (2006), *The Aluminium Sector Greenhouse Gas Protocol, Addendum to the WRI/WBCSD GHG Protocol*, US EPA, International Aluminium Institute, *Protocol for Measurement of Tetrafluoromethane (CF₄) and Hexafluoroethane (C₂F₆) Emissions from Primary Aluminium Production*).

Choice of emission factor

Carbon dioxide emission factors have been applied to the quantities of fuels used in the production of anodes. One NGERs reporting facility has derived facility-specific CO₂ emission factors for coal tar and petroleum coke. It was assumed that the fuel specifications measured at this facility were equally applicable to all facilities.

The facility-specific fuel consumption emission factors for anode production are confidential, however, the implied total CO₂ emission factor per unit of aluminium produced is shown in table 4.17 and confirms that these values are within the historical range of IEFs and not significantly different to the mean of the values reported between 1990 and 2008.

In the case of emissions of perfluorocarbons, facility-specific emission factors at all facilities have been estimated and sourced from NGERs for 2009. National average factors for previous years have been supplied by the Australian Aluminium Council based on collected information on individual facility factors.

Activity data

Data on coke oven coke, petroleum coke and coal tar consumption for the purpose of production of aluminium have been collected under the NGER System for 2009. For the years 1990-2008, coal tar and petroleum coke consumption are derived from the carbon in the reported emissions and the typical composition of carbon anodes used in the aluminium production process.

Data on aluminium for the purposes of estimating emissions of PFCs has been obtained under the NGER System for 2009 and ABARES Commodity Statistics (various years) for 1990-2008.

Table 4.17: Emission factors: kg per tonne of aluminium production 1990, 2000–2009

Year	CO ₂ ^(a)	CF ₄	C ₂ F ₆
1990	1,636	0.416	0.054
2000	1,587	0.082	0.011
2001	1,603	0.112	0.015
2002	1,663	0.106	0.014
2003	1,638	0.101	0.013
2004	1,606	0.102	0.013
2005	1,611	0.106	0.014
2006	1,586	0.040	0.005
2007	1,608	0.033	0.004
2008	1,591	0.025	0.003
2009	1,583	0.020	0.002

Source: NGERs, Beyond Neutral 2008, GHD 2009a.

(a) implied emission factor including production and consumption of anodes.

The carbon anode consumed in aluminium smelting is approximately three per cent sulphur by weight. Based on the assumption that 413 kg of carbon from the carbon anode is oxidised (consumed) for each tonne of aluminium produced, this implies that approximately 12.77 kg of sulphur and 25.54 kg of sulphur dioxide are oxidised per tonne of aluminium produced.

Table 4.18: Aluminium: production and emissions 1990, 2000–2009

Year	Aluminium production (kt) ^(a)	Emissions ^(b) (Gg CO ₂ -e)
1990	1,235	5,971
2000	1,742	3,867
2001	1,788	4,411
2002	1,809	4,489
2003	1,855	4,482
2004	1,877	4,484
2005	1,890	4,581
2006	1,912	3,621
2007	1,954	3,641
2008	1,965	3,507
2009	1,980	3,444

Source: (a) ABARE 2009b/NGERS; (b) Beyond Neutral 2008, GHD 2009a;

4.5.4 SF₆ Used In Aluminium And Magnesium Foundries (2.C.4)

The inventory includes experimental quantities of SF₆ used between 1996 and 2000 as a cover gas in magnesium foundries preparatory to the development of a commercial magnesium casting plant (which was not, ultimately, commercially viable). The data on SF₆ use for this experimental foundry was supplied by CSIRO.

4.5.5 Other Metals Production (2.C.5)

Source Category Description

In Australia the other metals source category includes emissions from the production of copper, lead, nickel, silver and zinc. There are 10 major companies involved in the production of other metals in Australia. In Australia, the major zinc refinery, in Hobart, uses an electrolytic process, which is non-emissive. The major lead refinery, at Port Pirie, which also refines a small amount of zinc, uses blast furnace technology while another smaller lead-zinc refinery, at Cockle Creek, uses the Imperial Smelting Furnace technology.

Carbon dioxide emissions from the use of fossil fuels as reductants, or in the production of carbon anodes on-site, or as carbon anodes in these refineries are reported under this category. An equivalent deduction has been made from fuel consumption in stationary energy to ensure there is no double-count of fuels in the inventory.

Carbon dioxide emissions from the consumption of limestone in the production of other metals are reported under 2.A.3.

Australia's metal ores are predominantly sulphide ores leading to the generation of SO₂ as a by-product of metal production. SO₂ emissions from metal production are reported under this category.

Methodology

Emissions from the consumption of reductants in the production of other metals have been estimated using a tier 2 method. Emissions are estimated using country-specific energy contents and CO₂ emission factors for relevant fuels or, in certain cases, based on facility-specific emission factors.

Ore composition and stoichiometric relationships have been used to derive sulphur dioxide emission estimates for copper, lead, nickel, zinc, and silver. The general approach is illustrated using the example of zinc. Zinc occurs either as sulphide ores (ZnS) or carbonate ores (ZnCO₃). Australia's zinc production is predominantly from sulphide ores. The objective of the refining process to obtain primary refined zinc is to break the compound ore down by separating the sulphur from the zinc. Based on atomic and molecular weights, 0.980 tonnes of SO₂ will be released per tonne of primary refined zinc. Emission factors for other metals, based on stoichiometry relationships, are given in the Table 4.19.

Table 4.19: Sulphur dioxide emission factors for refined metals

Metal	Tonnes SO ₂ per tonne of refined metal
Copper	2.0
Lead	0.3
Nickel	1.1
Silver	0.3
Zinc	1.0

Choice of emission factor

Emission factors have been selected in accordance with the decision tree in section 1.6.1.

In the case of one company, a facility-specific CO₂ emission factor has been used for the consumption of petroleum coke. This emission factor is confidential. In all other cases, the factors are taken from Table 3.2.

Time series consistency has been maintained by the application of values for emission factors for 2009 for the period 1990-2008.

Activity data

Data on fuel consumed as reductants for the purpose of production of other metals have been collected under the NGER System for 2009. For the years 1990-2008, this level of fuel consumption has been assumed to be constant.

4.5.6 Uncertainties And Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas.

Activity data obtained under NGERS was compared with activity data obtained from the EITEIs program for each facility and with data obtained from GHD and Energreen consulting to ensure the consistent classification of sources and consistency of data.

Where facilities were newly identified from NGERS data as emitting facilities for a category – in 2.C.2 – estimates of fuel consumption were interpolated through the time period from the most recent year for which data was available to the year of commencement of the facility based on metal production estimates.

Where facility-specific emission factors were identified from NGERS data for particular facilities, in category 2.C.4, the reported emission factors for 2007, 2008 and 2009 were interpolated for each facility between 2006 and the most recent year for which data were available.

4.5.7 Source Specific QA/QC

This source category is covered by the general QA/QC of the greenhouse gas inventory in Section 1.6. Additional source specific quality control checks were undertaken to assess international comparability.

The Metal Products sector was reviewed independently by an international expert (Tsaranu) in 2007. The review was undertaken applying the same principles governing regular UNFCCC inventory desktop reviews. Small refinements were made to the iron and steel non-CO₂ methodology and general refinements made to the metal products chapter in response to recommendations made in this review.

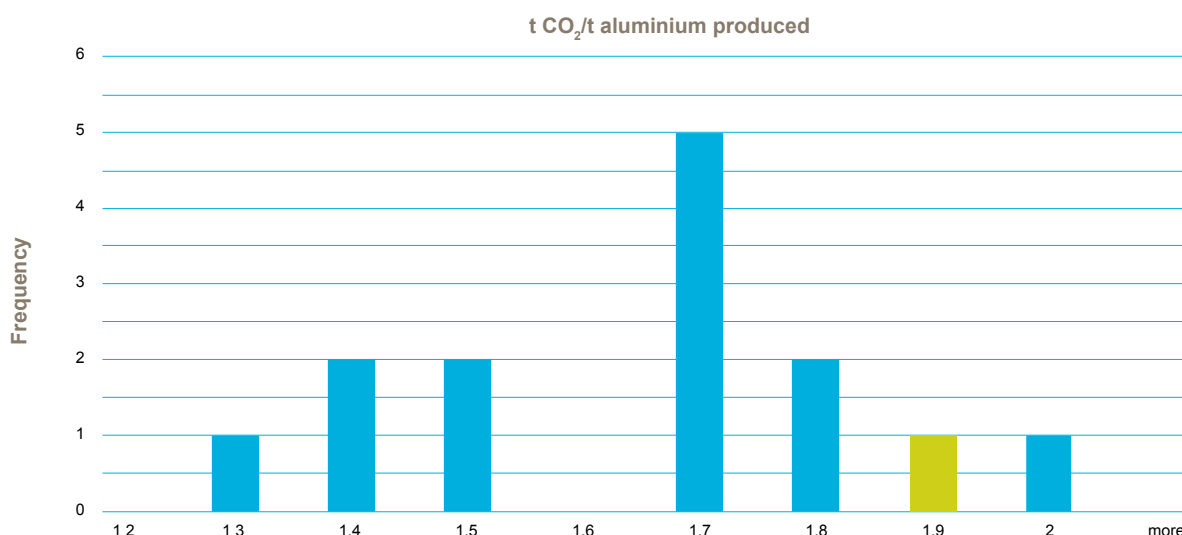
Aluminium

Emissions of PFCs by the Australian aluminium industry are a key category under both the level and trends analyses. Consequently, additional analysis has been performed to provide a comparison of Australian emission trends with those worldwide. The results of the comparison show that the trend in emissions per unit of production in Australia is very close to that observed worldwide. The decline in PFC emissions per unit of aluminium production in Australia since 1990 has mirrored the decline internationally (75%), whereas the International Aluminium Institute (2005) reports a decline of 73% between 1990 and 2003 worldwide. Emissions per unit of production reported by Australia are lower than the global averages, reflecting relatively modern plant and efficient operation, although this difference has narrowed slightly over time.

Monitoring of PFC concentrations occurs at the Cape Grim Baseline Air Pollution Station in Tasmania. Analysis of the observed atmospheric data has been undertaken by the CSIRO and compared to the emissions estimates in the inventory. Estimates of CF₄ and C₂F₆ emissions based on the measured data are in good agreement with Inventory estimates for 2009 (CSIRO 2011).

The quantity of CO₂ per tonne of aluminium produced has been compared with that from other Annex I parties reporting emissions from this source. The results of this comparison are shown in Figure 4.6.

Figure 4.6: Aluminium production implied emission factors for Annex I countries and Australia



The CO₂ implied emission factor for aluminium production for Australia ranges between 1.541 t CO₂/ t aluminium produced and 1.746 t CO₂/ t aluminium produced. Implied emission factors fluctuate observed year on year according to the quantities of carbon-based fuels used to produce anodes.

Statistical analysis indicates that the implied emission factor for aluminium production for Australia is not significantly different to the factors reported by other Annex I parties.

In order to maintain continuity in the compilation of industrial processes emissions estimates, the DCCEE engaged the external consultant previously used to collect activity data and emission factor information to undertake a quality control assessment of the full time-series of activity data, emission factors and emissions estimates. This work is of particular importance in industrial processes where confidentiality of historical activity data pose some challenges for the assessment of time-series consistency.

4.5.8 Recalculations Since The 2008 Inventory

Recalculations have been undertaken in metals production as a result of the re-allocation of emissions from the consumption of reductants in the production of ferroalloys and other metals.

Table 4.20: 2.C Metal Production: recalculation of total CO₂-e emissions (Gg), 1990-2008

	2010 Submission	2011 Submission	Change	Change
	Gg CO ₂ -e	Gg CO ₂ -e	Gg CO ₂ -e	%
2.C Metals				
1990	15,070	15,603	533	3.5%
1991	14,832	15,366	533	3.6%
1992	15,376	15,909	533	3.5%
1993	14,158	14,691	533	3.8%
1994	14,154	14,687	533	3.8%
1995	13,856	14,389	533	3.8%
1996	13,953	14,486	533	3.8%
1997	13,742	14,275	533	3.9%
1998	14,052	14,585	533	3.8%
1999	14,173	14,706	533	3.8%
2000	13,753	14,287	533	3.9%
2001	13,555	14,089	533	3.9%
2002	13,400	13,934	535	4.0%
2003	13,153	13,685	532	4.0%
2004	13,745	14,280	535	3.9%
2005	11,938	12,475	537	4.5%
2006	10,913	11,451	538	4.9%
2007	11,075	11,614	539	4.9%
2008	11,046	11,585	539	4.9%

4.5.9 Planned Improvements

Historical activity data for the consumption of reductants in the production of ferroalloys and other metals will be reviewed and updated where better information is available.

Further work is planned to investigate the possibility of re-allocating coal used as a reducing agent in the iron and steel industry from stationary energy to industrial processes. While NGERS has provided much useful information on metals production, further analysis of the data will be required to ensure the appropriate quantities of coal are deducted from stationary energy and reported under iron and steel production.

4.6 SOURCE CATEGORY 2.D OTHER PRODUCTION

4.6.1 Source Category Description

The supply of CO₂ gas for use in the food and drink industry is provided from three main sources in Australia. Three ammonia producers sell a proportion of the CO₂ generated as a by-product of the ammonia production process to the food and drink industry. Gas is also obtained from two natural CO₂ wells located at Caroline in South Australia (commissioned in 1967) and Boggy Creek in Victoria (commissioned in 1995). The third source is by product CO₂ from an ethylene oxide plant located in Botany in New South Wales.

In the case of the CO₂ wells and the ethylene oxide plant, some CO₂ sold is also used for medical and other purposes (such as use in fire extinguishers). However, all CO₂ sold by these operators is reported under 2.D.2 *Food and drink*.

A small source of CO₂ emissions also derives from the use of sodium bicarbonate in food production. These emissions are also reported under 2.D. Sodium bicarbonate is a by-product of the production of soda ash.

The manufacture of beer, wine, alcoholic spirits, and bread involve the use of fermentation processes. The IPCC (1997) indicate the fermentation of sugar by industry is not considered to be a net source of CO₂ emissions, consistent with the IPCC guidelines, Australia does not estimate CO₂ emissions from this source. NMVOC emissions from food and drink production, however, are included in the inventory. Production data for meat and poultry, beer and wine are obtained from ABS. Production data for sugar are obtained from ABARE (2009b).

4.6.2 Methodology

Emissions of CO₂ from food and drink are derived based on the assumption that all CO₂ gas used is emitted in the year of production.

Carbon dioxide generated in the production of ammonia and then captured for consumption in the food and drink industry is described in the method for the estimation of emissions from ammonia production (2.B.1). The quantity of CO₂ supplied from the two gas wells is derived based on published production capacity. The quantity of CO₂ supplied from the ethylene oxide plant is derived based on the production capacity of the plant and a CO₂ emission factor of 0.45 tonnes of CO₂ per tonne of ethylene oxide produced taken from the Netherlands National Inventory Report (no IPCC default factor is provided and the Netherlands is the only party to report emissions from this source). It is assumed that all CO₂ generated is sold for use in food and drink production.

The method for the calculation of emissions from the use of sodium bicarbonate is provided with the method for the estimation of emissions from soda ash (2.A.4).

Emissions of NMVOCs from food and drink production are based on tier 2 methods and IPCC default emission factors. Generally the methods involve multiplying the product activity level data (the amount of material produced or consumed) by an associated emission factor per unit of production or consumption. The NMVOC emission factors used are as follows:

- Beer 0.035 (kg NMVOC/hl beverage produced);
- Red Wine 0.08 (kg NMVOC/hl beverage produced);
- White Wine 0.035 (kg NMVOC/hl beverage produced);
- Bread 1.66 (kg NMVOC/t food produced);
- Sugar 10 (kg NMVOC/t food produced);
- Meat and Poultry 0.3 (kg NMVOC/t food produced).

4.6.3 Uncertainties And Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas. Time series consistency is ensured by use of consistent models, model parameters and datasets for the calculations of emissions estimates. Where changes to emission factors or methodologies occur, a full time series recalculation is undertaken.

4.6.4 Source Specific QA/QC

This source category is covered by the general QA/QC of the greenhouse gas inventory in Section 1.6.

4.6.5 Recalculations Since The 2008 Inventory

Recalculations have occurred in 2.D as a result of changes to quantities of CO₂ recovered from ammonia production.

Table 4.21: 2.D Food and Drink: recalculation of total CO₂-e emissions (Gg), 1990-2008

	2010 Submission	2011 Submission	Change	Change
	Gg CO ₂ -e	Gg CO ₂ -e	Gg CO ₂ -e	%
2.D Other Production				
1990	64	90	26	40.0%
1991	64	92	28	43.4%
1992	64	95	31	48.3%
1993	64	97	34	52.9%
1994	64	100	35	55.0%
1995	109	146	37	34.5%
1996	106	152	45	42.8%
1997	107	150	43	40.3%
1998	103	148	45	43.3%
1999	102	150	47	46.3%
2000	103	152	50	48.3%
2001	103	155	52	50.8%
2002	102	157	55	53.3%
2003	102	159	57	55.8%
2004	102	173	71	69.2%
2005	102	175	74	72.2%
2006	102	169	67	65.8%
2007	102	157	55	54.4%
2008	102	172	70	68.1%

4.6.6 Planned Improvements

Improvements to the estimation of the quantity of CO₂ sold by ammonia companies for food and drink are expected through refinements to the reporting provisions under NGERs – which require ammonia companies to report how much recovered CO₂ is used in urea production for the inventory year 2010.

4.7 SOURCE CATEGORY 2.E PRODUCTION OF HALOCARBONS AND SF₆

4.7.1 Source Category Description

HCFC-22 was produced in Australia from 1990 to 1995.

4.7.2 Methodology

For the production of HCFC-22, the IPCC tier 1 default methodology and an IPCC default emission factor of 0.04 t of HFC-23 per tonne of HCFC-22 has been used. This factor is at the upper limit of the default range and is considered appropriate given the age of the facility at the time of production.

4.8 SOURCE CATEGORY 2.F CONSUMPTION OF HALOCARBONS AND SF₆

4.8.1 Source Category Description

This sub-sector comprises emissions of synthetic gases from the use of halocarbons in refrigeration and air conditioning, foam blowing, fire extinguishers, aerosols/metered dose inhalers, solvents and SF₆ in electrical equipment and other miscellaneous applications.

The methodology used for compiling emissions estimates from this range of sources relates emissions to the stock of halocarbon (HFC) gases in various equipment end-use categories. Where equipment stock data are available (in the case of domestic refrigeration and air conditioning, motor vehicle air conditioning and metered dose inhalers), information on the age and lifetimes of the capital stock of appliances have been used to estimate emissions on a bottom up basis. Where these stock data are not available, a top-down approach has been used.

The method relies primarily on inputs of data on HFC imports (an estimate of potential emissions – there is no export or local production of HFCs in Australia) reported to the Department of Sustainability, Environment, Water, Population and Communities (DSEWPC) under the *Ozone Protection and Synthetic Greenhouse Gas Management Act, 2003*. As part of the licensing conditions specified in the *Act*, quantities of gas imported in bulk and in pre-charged equipment are reported to DSEWPC and these data are used for emissions estimation.

4.8.2 Methodology

Consistent with IPCC good practice, the methodology uses specified equations to estimate HFC emissions for each equipment type for three separate processes a) initial losses that occur at the initial charging of the equipment; b) emissions from leakages during the life of the equipment and c) the emissions from the disposal of the equipment. Initial losses occur when an amount of bulk imported gas (Mb_{ijkt}) is allocated to a specific equipment type j . Emissions during the life of the equipment depend, in the first year, on the amount of imported bulk gas allocated to the equipment type j and the amount of gas in imports of precharged equipment of type j (Mpc_{ijkt}) and, for every year thereafter, on the opening stock of gas in the equipment type (S_{ijkt-1}) plus any replenishments of gas (R_{ijkt}) in the equipment type that may have occurred in that year. Emissions at disposal depend upon the closing stock of gas of vintage k in year t (S_{ijkt}), the proportion of the capital stock retiring in each year, αK_{jkt} , and the quantity of gas recovered for destruction, D_{ijkt} .

The following equations set out the general process for estimating emissions of HFCs:

$$E_{ijkt} = Mb_{ijkt} * IL_{ijkt} + (S_{ijkt-1} + Mb_{ijkt} + Mpc_{ijkt} + R_{ijkt}) * (EF_{ij}) + (\alpha K_{jkt} * S_{ijkt} - D_{ijkt})$$

$$S_{ijkt} = S_{ijkt-1} + Mb_{ijkt} + Mpc_{ijkt} + R_{ijkt} - E_{ijkt} - D_{ijkt}$$

$$R_{ijkt} = \sum_{t-1, t-z} E_{ijkt}$$

$$D_{ijktbase} = \alpha K_{jkt} * S_{ijkt} * DF_{ijk}$$

$$D_{ijkt} = D_{ijktbase} / \sum_j \sum_k D_{ijktbase} * DTOT_t$$

and

$$E_t = \sum_i \sum_j \sum_k E_{ijkt}$$

Where: E_t is the sum of emissions of all gases of type i from all equipment types j and vintages k in year t

E_{ijkt} is the emissions of gas i from equipment type j and vintage k in year t

S_{ijkt-1} is the opening stock of gas i from equipment type j and vintage k in year t

S_{ijkt} is the closing stock of gas i from equipment type j and vintage k in year t

Mb_{ijkt} is the quantity of bulk import of gas i allocated to equipment type j for vintage k if $k = \text{year } t$, else = 0;

Mpc_{ijkt} is the quantity of gas i in imports of pre-charged equipment type j for vintage k if $k = \text{year } t$, else = 0;

R_{ijkt} is the amount of replenishment of the stock of gas i for equipment type j and vintage k in year t

EF_{ijkt} is leakage rate of gas i from equipment type j and vintage k in year t (in the first year of operation, EF is divided by 2 – assuming equipment is in operation for an average of 6 months)

IL_{ijkt} is the initial loss rate of gas i from equipment type j and vintage k in year t

αK_{jkt} is the proportion of the capital stock of equipment type j and vintage k retired in year t

$\sum_{t-z, t} E_{ijkt}$ is the sum of initial and annual emissions from $t-z$ to t where t is the current year and z is the number of years between replenishments

D_{ijkt} is the amount of gas i destroyed from equipment type j and vintage k in year t

DF_{ijkt} is the base destruction factor for gas i destroyed from equipment type j and vintage k in year t

$D_{ijktbase}$ is estimated base amount of gas i destroyed from equipment type j and vintage k in year t

$DTOT_t$ is the actual total gas destroyed reported by Refrigerant Reclaim Australia.

The initial loss rate (IL_{ijkt}) and the leakage rates of HFC gas during the life of the equipment (EF_{ijkt}) applied to each vintage of each equipment type are IPCC 2006 defaults (the mid-point of specified ranges). The IPCC 2006 default factors have been chosen as they reflect the most recently available knowledge on refrigerant losses from refrigeration and air conditioning equipment most commonly in use in Australia.

The amount of gas allocated to the replenishment of the stock of HFC gas and for each equipment type and vintage during the year (R_{ijkt}) is equal to the amount of gas leaked over the life of the equipment to that point and the frequency of replenishment undertaken by the operators of the equipment. Little information is available on this use of bulk imports of gas. Nonetheless, it is assumed that all commercial refrigeration and air-conditioning and fire protection systems are well maintained and subject to regular gas replenishment every 2 years of operation. Light vehicles are assumed to undergo a single gas re-charge at the mid-point of each unit's life. Sensitivity testing of the impact of these assumptions on emissions is provided in the QA/QC section. Lifetime emissions are not affected by these assumptions, while the time profile of emissions is considered to be not significantly sensitive to these assumptions.

Average equipment lifetimes are IPCC defaults. A constant proportion of the equipment stock (αK_{jk}) is assumed to be disposed over a period of time centred around the mid point of the average equipment lifetime. For example, the disposal of the refrigerator and air conditioning stocks is assumed to occur over a period from age five to a final date that ensures that the mid point is centred on the average age of equipment life.

Disposal losses reflect the residual charge or closing stock of gas in the equipment at the time of disposal (S_{ijkt}) and gas recovery for destruction undertaken at time of disposal. Data ($DTOT_t$) on recovery for destruction are supplied by Refrigerant Reclaim Australia (RRA), the sole entity responsible for the recovery and destruction of refrigerants in Australia. The RRA data are used to re-calibrate the recovery for individual equipment types estimated using base destruction factors or IPCC default disposal recovery rates (DF_{ijkt}). In effect, total recovery for destruction for the inventory as a whole is determined by data supplied by RRA, while the allocation of that total amount of destruction to the various equipment types is effectively determined by the relative IPCC default destruction rates.

Table 4.22: Halocarbons: key assumptions concerning average equipment life, initial and annual losses and replenishment rates, by equipment type

End Use Category	Average equipment life (a,b)	Loss on initial charge (a)	Annual loss (a)	Replenishment (c)
	Years	%	%	
Commercial refrigeration				
Stand-alone commercial applications	12.5	1.75	8.0	Full replenishment every 2 years
Medium and large commercial applications	11	1.75	22.5	Full replenishment every 2 years
Industrial commercial applications	22.5	1.75	16.0	Full replenishment every 2 years
Domestic refrigeration	15	0.6	0.3	No replenishment
Transport refrigeration	7.5	5.1	32.5	Full replenishment every 2 years
Light vehicle air conditioning	12	0.4	15.0	Full replenishment at 6 years
Heavy vehicle air conditioning	12.5	0.4	15.0	Full replenishment every 2 years
Domestic stationary air conditioning				
Refrigerated portable air conditioners	15	0.6	5.5	No replenishment
Split system air conditioners	15	0.6	5.5	No replenishment
Packaged air conditioners	15	0.6	5.5	No replenishment
Commercial air conditioners	22.5	5.1	8.5	Full replenishment every 2 years
Foams (closed cell)	20	10.0	4.5	No replenishment
Aerosols	2	0.0	50.0	No replenishment
Fire	10	0.4	5.0	Full replenishment every 2 years
Metered Dose Inhalers	2	0.0	50.0	No replenishment

(a) IPCC 2006, (b) Burnbank 2002, (c) DCCCE

Activity data: HFC gas imported into Australia in 2009

Data on imports of HFC gases are reported to the Department of Sustainability, Environment, Water, Population and Communities (DSEWPC) under licensing arrangements operating under the *Ozone Protection and Synthetic Greenhouse Gas Management Act, 2003*. Imports of bulk gas are allocated initially to individual end uses on the basis of a consideration of the amount of gas required for domestic production and replenishment/servicing and retrofitting for the sources which are estimated on a bottom-up basis (gas demand in domestic refrigeration, packaged, split and refrigerated portable air-conditioning and light vehicle air conditioning). After this initial gas demand is satisfied, the residual bulk gas is allocated to the remaining end use categories in proportion to the information on use supplied by DSEWPC and as reported by licensees under the Act. The sensitivity of these allocations on emissions estimates has been tested and the results are reported in the QA/QC section. The results show that lifetime emissions are not affected by these assumptions, and that the time profile of emissions – whilst impacted – is not considered sensitive to these assumptions.

Quantities of gas imported in bulk and contained in pre-charged equipment by end-use category are shown in Table 4.23.

Table 4.23: End-use allocation of imports of bulk and pre-charged HFC gas 2009

End Use Breakdown	Bulk Imports (Mt CO ₂ -e)	Pre-charged imports (Mt CO ₂ -e)	Total (Mt CO ₂ -e)
Refrigeration	5.093	0.360	5.453
Transport refrigeration	0.386	0.016	0.401
Commercial refrigeration	4.707	0.217	4.924
Domestic refrigeration and freezers	0.000	0.127	0.127
Stationary air-conditioning	0.400	2.558	2.957
Chillers	0.039	0.177	0.216
Refrigerated portable	0.000	0.088	0.088
Split systems	0.000	2.214	2.214
Packaged systems	0.360	0.078	0.439
Mobile air-conditioning	1.094	0.530	1.625
Cars	0.865	0.470	1.335
Trucks	0.229	0.060	0.290
Foam	0.187	-	0.187
Aerosols/solvents	0.162	-	0.162
Fire equipment	0.158	-	0.158
Metered dose inhalers		0.176	0.176
TOTAL	7.094	3.623	10.717

Source: Australian Government DSEWPC, DCCEE.

Backcasting

Collection of data on HFC imports under the *Act* commenced in the 2005 financial year. There are no data available on the import of HFCs for years prior to 2005. It is therefore necessary to backcast import data to enable an estimate of the bank of gas and associated emissions. For each of the end-use categories information on the transition from the use of CFC refrigerants to HFC refrigerants provided in Burnbank 2002 has been used to determine a time-series of HFC imports up to 2005 when actual import data are available.

Breakdown of gas species (i)

The bulk gas import data collected under the *Act* are disaggregated into HFC 134a, 'Other HFCs' and 'Exotic HFCs'. The 'Other' gas category comprises a known group of gases but is reported as an aggregated CO₂-e value, whilst the constituent gases in the 'Exotic' gas category are not reported but are provided as an aggregated CO₂-e value. Gas imported in pre-charged equipment is disaggregated into the following equipment types:

Stationary air conditioners

- Chillers – High pressure
- Chillers – Low pressure
- Air conditioner – Other
- Packaged – Air cooled
- Packaged – Water cooled
- Packaged – Window

Refrigerated portable

- Split system – Multi head/ variable refrigerant flow
- Split system – Single head

Refrigeration

- Commercial refrigerated cabinets
- Domestic refrigerator and equipment
- Other commercial refrigeration
- Portable refrigerators (commercial)
- Self powered refrigerator (transport)
- Vehicle powered refrigerator

Mobile air conditioners

- Vehicles weighing less than 3.5 tonnes Gross Vehicle Mass (cars and light commercial vehicles)
- Vehicles weighing more than 3.5 tonnes Gross Vehicle Mass (heavy vehicles)

The pre-charged equipment data are also disaggregated by the refrigerant they contain. The refrigerants that are reported are as follows: HFC-125; HFC-134a; HFC-404a; HFC-407a; HFC-407b; HFC-407c; HFC-410a; HFC-413a; HFC-417a; HFC-507a. The speciated gases in pre-charged equipment are used to disaggregate the final emissions estimates in each end use category into individual HFC species for reporting in the CRF tables.

Overview of the stocks of gas in operating equipment

The allocation of total gas imports to individual end use categories determines the relative sizes of gas stocks contained in equipment and the time profile of gas losses from the stock. Figure 4.7 shows the growth in the stock of synthetic gas in operating equipment. The chart shows significant growth in gas contained in commercial refrigeration systems, motor vehicle air conditioners and split system air conditioners. The general growth in the stock of gas in operating equipment reflects the transition from CFC to HFC refrigerant use associated with the Montreal Protocol controls on CFC use. In addition to the transitional trend, the recent strong growth in commercial refrigeration systems reflects similar growth in Australia's economy, whilst the growth in motor vehicle air conditioning and residential split systems reflects declines in relative prices of imported residential air conditioning systems as well as a transition in the vehicle fleet to more modern air conditioned vehicles.

The total stock and emissions from the consumption of halocarbons is shown in table 4.24.

Figure 4.7: Growth in the bank of HFC gas in operating equipment 1990–2009 (Mt CO₂-e)

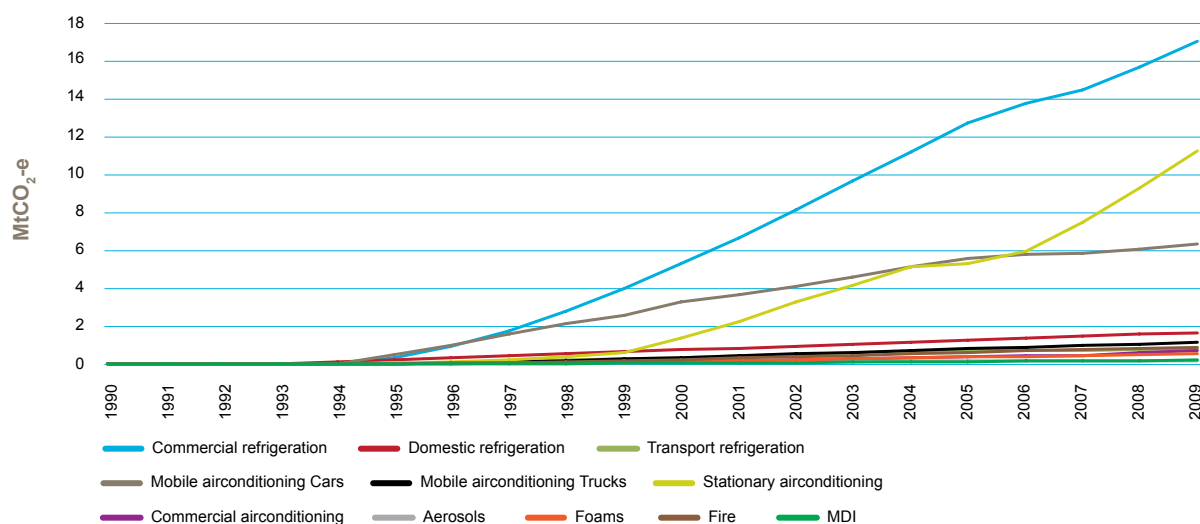


Table 4.24: Halocarbons: estimated stock and emissions: all equipment types

Year	Stock of gas (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1990	0.00	0.00
1991	0.00	0.00
1992	0.00	0.00
1993	0.00	0.00
1994	0.10	0.00
1995	1.14	0.11
1996	2.55	0.32
1997	4.37	0.61
1998	6.48	0.94
1999	8.72	1.33
2000	11.92	1.75
2001	15.04	2.26
2002	18.40	2.75
2003	21.76	3.31
2004	25.36	3.99
2005	28.02	4.47
2006	30.25	4.85
2007	32.90	5.33
2008	36.53	5.71
2009	40.64	6.25

Refrigeration and air conditioning (2.F.1)

The refrigeration and air-conditioning sector accounts for the majority of HFC consumption in Australia. Emissions from any piece of equipment include both the amount of chemical leaked during initial charging of equipment and the amount emitted during service life. Emissions also occur at equipment disposal. The disposal emission equation assumes that a certain percentage of the chemical charge will be emitted to the atmosphere when that vintage is discarded. Disposal emissions are thus a function of the quantity of chemical contained in the retiring equipment and the proportion of chemical released at disposal. The rate at which equipment is retired is based on IPCC default average service-lives for the various types of equipment.

Domestic Refrigeration and freezers

A bottom-up capital stock model has been used to determine a time-series for the stock of gas contained in domestic refrigeration and freezers. The estimates are based on data on the number of households and the numbers of domestic fridge freezers found in each household in Australia (ABS 2008b and ABS 2008c) and pre-charged equipment import data provided by DSEWPC.

Average charges per unit for domestic refrigerators are based on the pre-charged equipment data collected under the *Act* and were 0.163 kg in 2009. It is assumed that all new equipment and imports from 1994 onwards contain HFC refrigerants (Burnbank 2002). Service life emissions are derived using the IPCC default leakage rates.

Unit disposals are based on an average lifetime of 15 years with the first units in each vintage retiring after 5 years (Burnbank 2002). Under these assumptions the last units in each vintage are retired after 25 years.

The stock of HFCs in operating equipment is estimated to be 1.65 Mt CO₂-e in 2009. Domestic production of household refrigerators no longer takes place in Australia with the last producer Fisher and Paykel completing the relocation of their remaining production facility to Thailand in August 2009⁵. Gas imported in pre-charged equipment amounted to 0.13 Mt CO₂-e in 2009. It is assumed that no replenishment of gas losses from domestic refrigerators takes place as the units contain small well-sealed charges of gas. Upon disposal, a base assumption of 8 per cent of retiring gas in each year is recovered for destruction. Actual recovery for destruction is estimated by calibrating the results of these assumptions with data on actual total recovery obtained for destruction from RRA.

Table 4.25 shows the capital stocks, total number of Australian households, HFC stock and emissions from domestic refrigeration from 1994 to 2009. Comparison of refrigerator stocks and household numbers shows that there has been a trend toward the use of more than one refrigerator in the home. The average number of refrigerators per household has increased from 1.3 in 1994 to 1.4 in 2009.

Unit disposals are based on an average lifetime of 15 years with 5% of units in each vintage retiring after 5 years (Burnbank 2002). Under these assumptions the last units in each vintage are retired after 25 years. The base assumption is made that 70% of each unit's residual charge is emitted upon disposal (although this is calibrated for actual data on total recovery for destruction).

Table 4.25: Halocarbons: estimated stock and emissions: domestic refrigerator/freezers

Year	Domestic refrigerator stock ^(a)	Number of households ^(b)	Stock of gas (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1994	8,382,254	6,445,945	0.104	0.001
1995	8,578,471	6,690,307	0.209	0.001
1996	8,774,688	6,762,115	0.316	0.001
1997	8,970,905	6,910,143	0.424	0.002
1998	9,167,123	7,015,213	0.534	0.002
1999	9,363,340	7,126,529	0.640	0.006
2000	9,538,827	7,249,911	0.738	0.010
2001	9,714,313	7,393,042	0.834	0.013
2002	9,937,512	7,505,674	0.935	0.017
2003	10,226,951	7,645,366	1.045	0.021
2004	10,518,356	7,783,687	1.153	0.028
2005	10,811,949	7,920,842	1.262	0.029
2006	11,045,172	8,058,248	1.366	0.031
2007	11,267,700	8,186,752	1.493	0.037
2008	11,492,144	8,315,589	1.580	0.040
2009	11,670,685	8,444,778	1.652	0.040

Sources: (a) ABS 2008b; (b) ABS 2008c

Domestic air conditioning

Stationary air conditioning comprises refrigerated portable, split and packaged systems. Emissions from this sub category are estimated on a bottom-up basis using equipment population estimates based on numbers of households and white-goods data provided in ABS 2008c, and pre-charged equipment import data provided by DSEWPC. Tables 4.26, 4.27 and 4.28 show the capital stocks, HFC stocks and emissions from the three types of air conditioning equipment from 1995 to 2009.

The stock of HFCs in operating equipment is estimated to be 11.22 Mt CO₂-e in 2009. Imports of bulk gas for domestic production and servicing/replenishment were 0.36 Mt CO₂-e whilst 2.38 Mt CO₂-e arrived in pre-charged equipment in 2009.

⁵ <http://www.fisherpaykel.com/global/investors/Investors-PDFs/Annual%20Reports/Annual%20Review%20Year%20Ended%2031%20March%202010.pdf>

IPCC default leakage rates are applied to each gas vintage. Quantities of residual gas disposed in each vintage are based on the IPCC average equipment life of 15 years. The first disposals of gas are assumed to occur after 5 years of operation continuing until the last units are retired after 25 years. Upon disposal, a base assumption of 5 per cent of retiring gas in each year is recovered for destruction. Actual recovery for destruction is estimated by calibrating the results of these assumptions with data on actual total recovery obtained for destruction from RRA.

Table 4.26: Halocarbons: estimated stock and emissions: split system stationary airconditioners

Year	Split system air conditioner stocks ^(a)	Stock of gas (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1995	415,706	0.01	0.00
1996	444,085	0.03	0.00
1997	472,464	0.06	0.00
1998	500,843	0.11	0.01
1999	529,222	0.17	0.01
2000	935,562	0.90	0.04
2001	1,341,902	1.71	0.08
2002	1,748,243	2.61	0.13
2003	2,075,944	3.37	0.18
2004	2,403,645	4.20	0.23
2005	2,731,346	4.39	0.27
2006	3,062,064	4.95	0.32
2007	3,392,782	6.28	0.41
2008	3,723,500	7.86	0.52
2009	3,852,727	9.43	0.62

Sources: (a) ABS 2008c

Table 4.27: Halocarbons: estimated stock and emissions: packaged air conditioners

Year	Packaged air conditioner stocks ^(a)	Stock of gas (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1995	1,582,177	0.02	0.00
1996	1,637,956	0.06	0.00
1997	1,693,734	0.13	0.01
1998	1,749,512	0.25	0.01
1999	1,805,290	0.40	0.02
2000	1,782,640	0.46	0.03
2001	1,759,989	0.54	0.03
2002	1,737,339	0.63	0.04
2003	1,723,761	0.76	0.05
2004	1,710,182	0.90	0.06
2005	1,696,604	0.90	0.07
2006	1,652,936	0.99	0.07
2007	1,609,268	1.08	0.08
2008	1,565,600	1.22	0.09
2009	1,619,935	1.55	0.11

Sources: (a) ABS 2008c

Table 4.28: Halocarbons: estimated stock and emissions: refrigerated portable air conditioners

Year	Refrigerated portable system stocks ^(a)	Stock of gas (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1995	160,971	0.00	0.00
1996	155,350	0.00	0.00
1997	149,730	0.00	0.00
1998	144,109	0.00	0.00
1999	138,488	0.00	0.00
2000	141,998	0.00	0.00
2001	145,508	0.00	0.00
2002	149,019	0.00	0.00
2003	177,029	0.01	0.00
2004	205,040	0.01	0.00
2005	233,050	0.01	0.00
2006	215,967	0.02	0.00
2007	198,883	0.11	0.00
2008	181,800	0.17	0.01
2009	188,109	0.24	0.01

Sources: (a) ABS 2008c

Mobile air-conditioning (Passenger Cars)

Emissions from the use of air conditioners in passenger cars and light commercial vehicles (vehicles under 3.5 tonnes gross vehicle mass) are also estimated on a bottom-up basis. Data on the stock of motor vehicles obtained from the ABS Motor Vehicle Census (ABS 2009a) have been used to construct a capital stock model. In Table 4.29 the stock of light vehicles, the stock of HFC gas contained in motor vehicle air-conditioners and the associated emissions are reported. It is assumed that all new units manufactured from 1995 onwards contain HFC-134a.

Table 4.29: Halocarbons: estimated stock and emissions: light vehicle air conditioners

Year	Light vehicle stocks ^(a)	Stock of gas in operating equipment	Emissions
		(Mt CO ₂ -e)	(Mt CO ₂ -e)
1995	10,008,978	0.50	0.04
1996	10,398,727	0.98	0.13
1997	10,603,895	1.55	0.21
1998	10,831,891	2.12	0.30
1999	11,159,167	2.57	0.39
2000	11,510,386	3.24	0.45
2001	11,827,630	3.64	0.53
2002	12,122,155	4.09	0.60
2003	12,469,273	4.57	0.68
2004	13,024,269	5.15	0.85
2005	13,172,992	5.56	0.85
2006	13,572,803	5.76	0.90
2007	13,860,265	5.86	0.95
2008	14,224,192	6.05	0.97
2009	14,615,751	6.35	0.99

Source: (a) ABS 2009a

The stock of gas has been compiled using the ABS data on light vehicle stocks, assumptions about proportions of each vintage with air-conditioning and an average charge per unit of 0.81 kg of HFC-134a (derived from import data). Assumptions needed on the percentage of pre-1995 vehicles retrofitted with HFC-134a units to estimate an addition to the stock of gas have been taken from Burnbank 2002.

Equipment disposals are based on the IPCC default average life-span of 12 years with the first units of each vintage retiring after 5 years of operation. Under these assumptions, the last units of each vintage would be retired after 19 years. It is assumed that between 1995 and 2000, an initial base assumption that 40% of the remaining charge contained in disposed units is recovered for destruction. From 2000 onwards, the assumed base rate of recovery is assumed to grow at 1% per year. The quantity of gas not recovered is emitted to the atmosphere at disposal. These assumptions are consistent with Burnbank 2002. Actual recovery for destruction is estimated by calibrating the results of these assumptions with data on actual total recovery obtained for destruction from RRA.

Mobile air conditioning (heavy vehicles)

This source category comprises emissions from air conditioning units in vehicles over 3.5 tonnes gross vehicle mass.

The quantities of imported gas are allocated to heavy vehicle air conditioning on the basis of pre-charged equipment as reported by DSEWPC and a proportion of bulk gas adjusted for gas demand in domestic refrigeration and air conditioning and mobile air conditioning. Once the gas required for loss replenishment needs is satisfied, the remaining bulk gas is allocated to charging new locally produced units.

The bank of HFCs in operating equipment is estimated to be 1.13 Mt CO₂-e in 2009. Imports of bulk gas for domestic production and servicing/replenishment were 0.23 Mt CO₂-e whilst 0.06 Mt CO₂-e arrived in pre-charged equipment.

IPCC default leakage rates are applied to each gas vintage. Quantities of residual gas disposed in each vintage are based on the IPCC average equipment life of 12.5 years and the assumption that gas losses are replenished after every 2 years of a unit's life. The first disposals of gas occur after 5 years of operation continuing until the last units are retired after 20 years. Upon disposal, a base calculation of 8 per cent of retiring gas in each year is recovered for destruction. Actual recovery for destruction is estimated by calibrating the results of these assumptions with data on actual total recovery obtained for destruction from RRA.

Table 4.30: Halocarbons: estimated stock and emissions: heavy vehicle air conditioners

Year	Imports of gas (Mt CO ₂ -e)	Stock of gas in operating equipment (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1995	0.02	0.02	0.00
1996	0.04	0.06	0.01
1997	0.06	0.11	0.01
1998	0.09	0.17	0.02
1999	0.11	0.25	0.03
2000	0.13	0.33	0.05
2001	0.15	0.42	0.06
2002	0.17	0.51	0.08
2003	0.19	0.61	0.09
2004	0.22	0.71	0.11
2005	0.24	0.81	0.13
2006	0.23	0.89	0.15
2007	0.24	0.96	0.16
2008	0.28	1.05	0.18
2009	0.29	1.13	0.19

Source: DSEWPC – HFC import data collected under the *Ozone Protection and Synthetic Greenhouse Gas Management Act* (2003)

Transport refrigeration

Transport refrigeration comprises vehicle and self-powered refrigeration units used in commercial vehicles.

Quantities of imported gas are allocated to transport refrigeration on the basis of pre-charged equipment as reported by DSEWPC and a proportion of bulk gas adjusted for gas demand in domestic refrigeration and air conditioning and mobile air conditioning. Once the gas demand for loss replenishment is satisfied, the remaining bulk gas is allocated to charging new locally produced units.

The stock of HFCs in operating equipment is estimated to be 0.79 Mt CO₂-e in 2009. Imports of bulk gas for domestic production and servicing/replenishment were 0.39 Mt CO₂-e whilst 0.02 Mt CO₂-e arrived in pre-charged equipment in 2009.

IPCC default leakage rates are applied to each gas vintage. Quantities of residual gas disposed in each vintage are based on the IPCC average equipment life of 7.5 years and the assumption that gas losses are replenished after every 2 years of a unit's life up to the year of disposal. It is assumed that the first disposals of gas occur after 5 years of operation and continue until the last units are retired after 10 years. Upon disposal, a base assumption of 9% of retiring gas is assumed to be recovered for destruction.

Table 4.31: Halocarbons: estimated stock and emissions: transport refrigeration

Year	Imports of gas (Mt CO ₂ -e)	Stock of gas in operating equipment (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1995	0.03	0.02	0.01
1996	0.06	0.07	0.02
1997	0.09	0.13	0.04
1998	0.13	0.19	0.06
1999	0.16	0.27	0.08
2000	0.19	0.35	0.11
2001	0.22	0.42	0.14
2002	0.25	0.50	0.17
2003	0.28	0.57	0.20
2004	0.31	0.65	0.23
2005	0.35	0.71	0.28
2006	0.32	0.73	0.28
2007	0.32	0.71	0.31
2008	0.37	0.76	0.30
2009	0.40	0.79	0.35

Source: DSEWPC – HFC import data collected under the *Ozone Protection and Synthetic Greenhouse Gas Management Act* (2003)

Commercial refrigeration

Commercial refrigeration comprises stand-alone, medium and large and industrial refrigeration units and is the most significant user of synthetic gases in Australia.

The quantities of imported gas are allocated to commercial refrigeration on the basis of pre-charged equipment as reported by DSEWPC and a proportion of bulk gas adjusted for gas demand in domestic refrigeration and air conditioning and mobile air conditioning. Once the gas required for loss replenishment needs is satisfied, the remaining bulk gas is allocated to charging new locally produced units.

The stock of HFCs in operating equipment is estimated to be 16.99 Mt CO₂-e in 2009. Imports of bulk gas allocated for domestic production and servicing/replenishment were 4.71 Mt CO₂-e whilst 0.22 Mt CO₂-e arrived in pre-charged equipment in 2009.

IPCC default leakage rates are applied to each gas vintage from 1995 onwards. Quantities of residual gas disposed in each vintage are based on the IPCC average equipment life of 12.5 years for stand-alone units, 11 years for medium and large applications and 22.5 years for industrial systems and the DCCEE assumption that gas losses are replenished after every 2 years of a unit's life. It is assumed that the first disposals of gas occur after 5 years of operation and continue until the last units are retired after 20 years for stand-alone units, 17 years for medium and large applications and 40 years for industrial systems. Upon disposal, a base assumption of 10% of retiring gas is recovered for destruction.

Table 4.32: Halocarbons: estimated stock and emissions: commercial refrigeration

Year	Imports of gas (Mt CO ₂ -e)	Stock of gas (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1995	0.37	0.33	0.04
1996	0.74	0.94	0.13
1997	1.11	1.77	0.28
1998	1.48	2.80	0.46
1999	1.85	3.98	0.67
2000	2.23	5.28	0.91
2001	2.60	6.67	1.20
2002	2.97	8.13	1.47
2003	3.34	9.63	1.80
2004	3.71	11.18	2.13
2005	4.08	12.73	2.45
2006	3.80	13.76	2.68
2007	3.74	14.47	2.93
2008	4.38	15.63	3.11
2009	4.92	16.99	3.39

Source: DSEWPC – HFC import data collected under the *Ozone Protection and Synthetic Greenhouse Gas Management Act* (2003)

Commercial air conditioning

Commercial air conditioning covers the use of chiller units used in commercial buildings.

Quantities of imported gas are allocated to commercial refrigeration on the basis of pre-charged equipment as reported by DSEWPC and a proportion of bulk gas adjusted for gas demand in domestic refrigeration and air conditioning and mobile air conditioning. Once the gas demand for loss replenishment is satisfied, the remaining bulk gas is allocated to charging new locally produced units.

The bank of HFCs in operating equipment is estimated to be 0.72 Mt CO₂-e in 2009. Imports of bulk gas allocated for domestic production and servicing/replenishment were 0.04 Mt CO₂-e whilst 0.18 Mt CO₂-e arrived in pre-charged equipment in 2009.

IPCC default leakage rates are applied to each gas vintage. Quantities of residual gas disposed in each vintage are based on the IPCC average equipment life of 22.5 years and the assumption that gas losses are replenished after every 2 years of a unit's life up to the year of disposal. The first disposals of gas occur after 5 years of operation continuing until the last units are retired after 40 years. Upon disposal, a base assumption of 16 per cent of retiring gas is recovered for destruction.

Table 4.33: Halocarbons: estimated stock and emissions: commercial air conditioners

Year	Imports of gas (Mt CO ₂ -e)	Stock of gas in operating equipment (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1995	0.01	0.01	0.00
1996	0.02	0.02	0.00
1997	0.03	0.04	0.00
1998	0.03	0.07	0.01
1999	0.04	0.10	0.01
2000	0.05	0.14	0.01
2001	0.06	0.19	0.02
2002	0.07	0.23	0.02
2003	0.08	0.28	0.03
2004	0.09	0.34	0.03
2005	0.09	0.39	0.04
2006	0.07	0.42	0.04
2007	0.08	0.46	0.04
2008	0.18	0.58	0.06
2009	0.22	0.72	0.07

Source: DSEWPC – HFC import data collected under the *Ozone Protection and Synthetic Greenhouse Gas Management Act* (2003)

Foam Blowing (2.F.2)

The quantities of imported gas are allocated to foam on the basis of a proportion of bulk gas adjusted for gas demand in domestic refrigeration and air conditioning and mobile air conditioning.

The bank of HFCs contained in foams is estimated to be 0.54 Mt CO₂-e in 2009. Imports of bulk gas for domestic production were 0.19 Mt CO₂-e.

IPCC default leakage rates are applied to each gas vintage. Quantities of residual gas disposed in each vintage are based on the IPCC average equipment life of 20 years. The first disposals of gas occur after 5 years of operation continuing until the last units are retired after 35 years. There is no recovery or replenishment assumed in foams.

Foams are given emission profiles depending on the foam type (open cell or closed cell). Open cell foams are assumed to be 100 per cent emissive in the year of manufacture. Closed cell foams are assumed to emit a portion of their total HFC content upon manufacture, a portion at a constant rate over the lifetime of the foam, and a portion at disposal.

Table 4.34: Halocarbons: estimated stock and emissions: foam

Year	Imports of gas (Mt CO ₂ -e)	Stock of gas in operating equipment (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1995	0.01	0.01	0.01
1996	0.03	0.02	0.02
1997	0.04	0.04	0.03
1998	0.06	0.06	0.03
1999	0.07	0.09	0.04
2000	0.09	0.12	0.05
2001	0.10	0.16	0.06
2002	0.12	0.21	0.07
2003	0.13	0.25	0.08
2004	0.15	0.30	0.10
2005	0.16	0.36	0.11
2006	0.15	0.40	0.10
2007	0.14	0.44	0.10
2008	0.17	0.49	0.12
2009	0.19	0.54	0.13

Source: DSEWPC – HFC import data collected under the *Ozone Protection and Synthetic Greenhouse Gas Management Act* (2003)

Fire Extinguishers (2.F.3)

The quantities of imported gas are allocated to fire extinguishers on the basis of pre-charged equipment as reported by DSEWPC and a proportion of bulk gas adjusted for gas demand in domestic refrigeration and air conditioning and mobile air conditioning. Once the gas required for loss replenishment needs is satisfied, the remaining bulk gas is allocated to charging new locally produced units.

The bank of HFCs contained in fire protection equipment is estimated to be 0.89 Mt CO₂-e in 2009. Imports of bulk gas for domestic production were 0.16 Mt CO₂-e.

IPCC default leakage rates are applied to each gas vintage from 1995 onwards. Quantities of residual gas disposed in each vintage are based on the IPCC average equipment life of 10 years and the assumption that gas losses are replenished after every 2 years of a unit's life. The first disposals of gas occur after 5 years of operation continuing until the last units are retired after 15 years. Upon disposal, approximately 26 per cent of retiring gas in each year is recovered for destruction.

The UNFCCC expert review of Australia's 2008 submission recommended that the completeness of the *industrial processes* estimates be improved by inclusion of estimates of emissions from PFC use in fire extinguishers. In response, the Australian Fire Protection Association (FPA) was consulted and they confirmed that the ozone depleting or synthetic greenhouse fire fighting gases most common in Australia are: FE 227 (HFC 227ea), FM 200 (HFC 227ea), NAF-S-III (HCFC Blend A) and NAF-P-III (HCFC Blend C). The use of other gases is considered quite rare. On this basis, PFC use in fire extinguishers is considered to be 'Not Occurring'.

Table 4.35: Halocarbons: estimated stock and emissions: fire protection equipment

Year	Imports of gas (Mt CO ₂ -e)	Stock of gas in operating equipment (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1995	0.01	0.01	0.00
1996	0.02	0.04	0.00
1997	0.04	0.07	0.00
1998	0.05	0.11	0.00
1999	0.06	0.17	0.01
2000	0.07	0.23	0.01
2001	0.09	0.30	0.01
2002	0.10	0.37	0.02
2003	0.11	0.45	0.02
2004	0.12	0.54	0.03
2005	0.13	0.62	0.03
2006	0.12	0.69	0.04
2007	0.12	0.74	0.04
2008	0.14	0.81	0.05
2009	0.16	0.89	0.04

Source: DSEWPC – HFC import data collected under the *Ozone Protection and Synthetic Greenhouse Gas Management Act* (2003)

Aerosols/Metered Dose Inhalers and Solvents (2.F.4 and 2.F.5)

Emissions from these sectors come from two sources: product use and fugitive emissions associated with product manufacture. Emissions from solvent and aerosol product use can be assumed to be 100 per cent of the charge size (e.g. 100 percent of consumption).

The quantities of bulk gas imported into Australia and allocated for use in aerosols is based on the proportion of reported end use adjusted for gas requirements in domestic refrigerator and air conditioning and mobile air conditioning. There are no imports of pre-charged equipment in Australia and no replenishment is assumed to occur. Therefore all gas imported in bulk goes into charging domestically produced stock.

The complete charge of gas from an aerosol application is assumed to be lost over 3 years with any residual charge being completely emitted upon disposal.

There is no domestic production of MDIs in Australia. Imports of metered dose inhalers containing HFCs are not covered by the *Ozone Protection and Synthetic Greenhouse Gas Management Act* (2003) so that data on HFC consumption of metered dose inhalers cannot be derived from this source. Consequently, emissions of HFCs from the use of metered dose inhalers are estimated on a bottom up basis. Estimates of the imports of gas contained in metered dose inhalers is based on information supplied by DSEWPC on the number of MDIs imported into Australia in 2009 and a per-capita based estimation of imports up to that year. Assumptions about the penetration of HFC propellants in imported MDIs are based on information in Burnbank 2002. It is assumed that each imported unit is pre-charged with 18 grams of HFC-134a based on information supplied from DSEWPC (Annie Gabriel, pers comm).

Emissions from MDIs are estimated according to the same assumptions used for aerosols.

The bank of HFCs contained in metered dose inhalers is estimated to be 0.19 Mt CO₂-e in 2009. Imports of bulk gas for domestic production were 0.18 Mt CO₂-e. Table 4.36 shows the growth in imports and the bank of HFC in metered dose inhalers along with the associated emissions from this bank.

Table 4.36 Halocarbons: estimated stock and emissions: metered dose inhalers

Year	Imports of gas (Mt CO ₂ -e)	Stock of gas in operating equipment (Mt CO ₂ -e)	Emissions (Mt CO ₂ -e)
1998	0.01	0.01	0.00
1999	0.02	0.02	0.01
2000	0.04	0.04	0.02
2001	0.05	0.05	0.04
2002	0.07	0.07	0.05
2003	0.08	0.08	0.06
2004	0.09	0.10	0.08
2005	0.11	0.12	0.09
2006	0.12	0.13	0.11
2007	0.14	0.15	0.12
2008	0.16	0.17	0.14
2009	0.18	0.19	0.16

Source: DSEWPC

Semiconductor Manufacture (2.F.6)

There is no semiconductor manufacture activity occurring in Australia.

Electrical Equipment (2.F.7)

Quantities of SF₆ contained within the electricity supply and distribution network in Australia (including switch gear used in large industrial facilities such as in aluminium) have been obtained under NGERS for 2009. It is assumed that this stock of gas leaks at an annual rate of 0.5 per cent (IPCC 2000).

In total, the quantity of SF₆ contained in operating equipment was reported as 8.4 Mt CO₂-e in 2009. As the size and extent of the electricity supply network remain stable year on year, this stock has been held constant for the entire time-series.

Other uses of SF₆

Emissions of SF₆ from other applications have been estimated for the first time in 2009. An estimate of SF₆ emissions from other applications including eye surgery, tracer gas studies, magnesium casting, plumbing services, tyre manufacture and industrial machinery equipment has been made on the basis of a per-capita emissions value derived from the National Inventory of New Zealand. An average per-capita emission rate of 0.001 tonne of SF₆ per person per year has been applied to Australia's total population to derive a time-series of emissions from this source.

4.8.3 Uncertainties And Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas. Time series consistency is ensured by use of consistent models, model parameters and datasets for the calculations of emissions estimates. Where changes to emission factors or methodologies occur, a full time series recalculation is undertaken.

4.8.4 Source Specific QA/QC

This source category is covered by the general QA/QC of the greenhouse gas inventory in Section 1.6. Data are obtained by the DSEWPC from companies under licensing arrangements established under the *Ozone Protection and Synthetic Greenhouse Gas Management Act* (2003) and is subject to verification against known published sources (the Australian Bureau of Statistics data on imports of HFC-134a).

The Consumption of Halocarbons and SF₆ sector has been reviewed independently by an international expert (Tsaranu 2007). The review was undertaken applying the same principles governing regular UNFCCC inventory desktop reviews. The emissions model was reviewed previously by Burnbank consulting. The outputs of the domestic refrigeration and mobile air-conditioning components of the model have been cross-checked against those reported in Burnbank 2002 with close agreement between the two sets of estimates.

Mass balances

An additional comprehensive review of this source was undertaken during 2009 in which HFC balances were completed to ensure that:

- all imported gas in bulk and pre-charged equipment is assigned to an appropriate end-use category, and
- stock changes and emissions and gas destruction were fully tracked and accounted for.

The results of these allocation and stock balances are presented in tables 4.37 and 4.38.

Table 4.37: Halocarbons: balance sheet – allocations of imported gas (Mt CO₂-e)

Gas Imported	Backcast import data											Import data reported by DSEWPC				
	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009
Bulk gas imported	0.07	1.04	1.57	2.19	2.72	3.16	3.76	3.98	4.56	5.14	5.88	6.02	5.40	5.14	6.08	7.09
Gas imported in pre-charged equipment	0.04	0.11	0.17	0.25	0.32	0.40	1.17	1.37	1.55	1.53	1.70	1.14	1.73	2.90	3.33	3.45
Total gas imported	0.10	1.14	1.74	2.44	3.04	3.56	4.93	5.35	6.11	6.68	7.59	7.15	7.13	8.04	9.41	10.54
Allocations to end use	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	
Transport refrigeration	0.00	0.03	0.06	0.09	0.13	0.16	0.19	0.22	0.25	0.28	0.31	0.35	0.32	0.32	0.37	0.40
Commercial refrigeration	0.00	0.37	0.74	1.11	1.48	1.85	2.23	2.60	2.97	3.34	3.71	4.08	3.80	3.74	4.38	4.92
Domestic refrigeration and freezers	0.10	0.11	0.11	0.11	0.11	0.11	0.11	0.11	0.12	0.14	0.14	0.15	0.15	0.17	0.14	0.13
Chillers	0.00	0.01	0.02	0.03	0.03	0.04	0.05	0.06	0.07	0.08	0.09	0.09	0.07	0.08	0.18	0.22
Refrigerated portable	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.02	0.09	0.06	0.09
Split systems	0.00	0.01	0.02	0.04	0.05	0.07	0.76	0.89	1.02	0.95	1.06	0.47	0.89	1.75	2.12	2.21
Packaged systems	0.00	0.02	0.05	0.08	0.12	0.17	0.09	0.11	0.13	0.17	0.20	0.08	0.16	0.17	0.24	0.44
Cars	0.00	0.54	0.62	0.80	0.87	0.84	1.13	0.93	1.05	1.17	1.47	1.27	1.10	1.08	1.18	1.33
Trucks	0.00	0.02	0.04	0.06	0.09	0.11	0.13	0.15	0.17	0.19	0.22	0.24	0.23	0.24	0.28	0.29
Foam	0.00	0.01	0.03	0.04	0.06	0.07	0.09	0.10	0.12	0.13	0.15	0.16	0.15	0.14	0.17	0.19
Aerosols/Solvents	0.00	0.01	0.03	0.04	0.05	0.06	0.08	0.09	0.10	0.11	0.13	0.14	0.13	0.12	0.15	0.16
Fire equipment	0.00	0.01	0.02	0.04	0.05	0.06	0.07	0.09	0.10	0.11	0.12	0.13	0.12	0.12	0.14	0.16
Total gas allocated	0.10	1.14	1.74	2.44	3.04	3.56	4.93	5.35	6.11	6.68	7.59	7.15	7.13	8.04	9.41	10.54
Balance against total gas imported	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00

Table 4.38: Halocarbons: Supply – use balance sheet (Mt CO₂-e)

	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009
Gas supply	0.10	1.14	1.74	2.44	3.04	3.56	4.93	5.35	6.11	6.68	7.59	7.15	7.13	8.04	9.41	10.54
Pre-charged Imports	0.04	0.11	0.17	0.25	0.32	0.40	1.17	1.37	1.55	1.53	1.70	1.14	1.73	2.90	3.33	3.45
Bulk gas used in production & retrofit	0.07	1.04	1.45	1.99	2.27	2.56	2.54	2.53	2.59	2.96	3.22	3.22	2.17	2.07	2.37	3.43
Bulk gas used in replenishment	0.00	0.00	0.12	0.20	0.45	0.60	1.22	1.45	1.97	2.18	2.67	2.80	3.23	3.07	3.71	3.67
Gas use/losses	0.00	0.11	0.33	0.62	0.94	1.33	1.74	2.25	2.76	3.33	4.01	4.51	4.92	5.41	5.79	6.45
Emissions	0.00	0.11	0.32	0.61	0.94	1.32	1.73	2.23	2.70	3.25	3.91	4.37	4.74	5.21	5.57	6.09
Recovery for destruction	0.00	0.00	0.01	0.01	0.00	0.01	0.01	0.03	0.05	0.09	0.10	0.13	0.18	0.20	0.22	0.35
Stock change	0.00	0.11	0.33	0.62	0.94	1.33	1.74	2.25	2.76	3.33	4.01	4.51	4.92	5.41	5.79	4.10
Balance	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00

Sensitivity testing

In addition to the HFC balances documented above, sensitivity analysis was undertaken to assess the impacts of changes to the allocation of bulk gas to end use as well as changes to the assumptions about replenishment rates in equipment. These two elements of the HFC model are where critical assumptions are made about the areas of consumption of imported gas and the servicing/replenishment habits of the consumers of this gas.

The effect of end use allocation on total emissions was tested by altering the percentage of bulk gas allocated to domestic, commercial and transport refrigeration (which is the biggest user of imported bulk gas) by 1%, 5%, 10% and 20% in all years with the residual gas allocated equally among the other end-use categories. In addition to this change in allocation, all gas imports are ceased after 2009.

Table 4.39: Halocarbons: results of sensitivity testing of allocation assumptions (Mt CO₂-e)

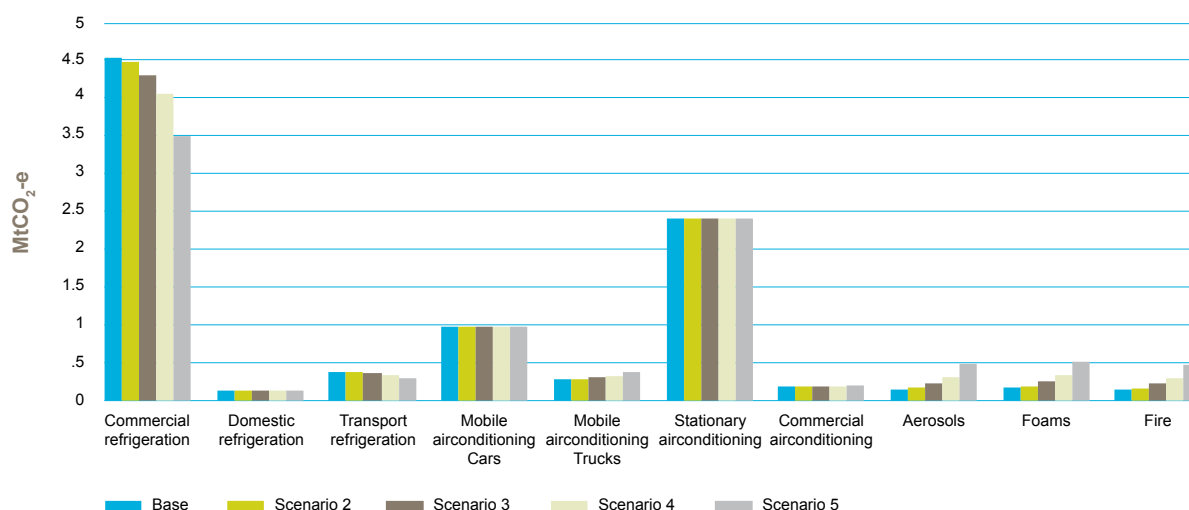
End use allocation	Allocation assumptions (% of total bulk imports)				
	Base	Case 1	Case 2	Case 3	Case 4
Aerosols/solvents	2%	2%	3%	4%	5%
Domestic/Commercial/ Transport refrigeration	60%	59%	55%	50%	40%
Fire	2%	2%	3%	4%	5%
Foam	2%	2%	3%	4%	6%
Mobile air conditioning	25%	25%	26%	27%	28%
Mobile OEM	1%	1%	2%	3%	5%
Stationary air conditioning	8%	8%	8%	9%	11%
Emissions in 2008 (Mt CO ₂ -e)					
Commercial refrigeration	3.23	3.19	3.06	2.89	2.50
Domestic refrigeration	0.04	0.04	0.04	0.04	0.04
Transport refrigeration	0.31	0.31	0.30	0.28	0.24
Mobile air conditioning cars	0.86	0.86	0.87	0.87	0.87
Mobile air conditioning trucks	0.19	0.19	0.20	0.22	0.25
Stationary air conditioning	0.62	0.62	0.62	0.62	0.62
Commercial air conditioning	0.06	0.06	0.06	0.06	0.07
Aerosols	0.13	0.15	0.20	0.27	0.43
Foams	0.13	0.14	0.18	0.24	0.37
Fire	0.05	0.06	0.08	0.11	0.18
Metered dose inhalers	0.14	0.14	0.14	0.14	0.14
Total	5.75	5.75	5.74	5.73	5.70
% change in total emissions compared with emissions in the base case		-0.04%	-0.19%	-0.40%	-0.86%

The results show that even with a 33% change in bulk gas allocation from domestic, transport and commercial refrigeration to other end use categories, total emissions in 2008 are changed by only 0.9%. This suggests that the estimate of emissions in any given year is relatively insensitive to changes in the allocation of bulk gas.

Figure 4.8 shows gas imports under the base end use assumption and each of the re-allocation assumptions. It can be seen that the gas diverted from domestic, commercial and transport refrigeration is re-allocated primarily to aerosols, foams, and fire protection. In total however, gas imports are unchanged as a result of the re-allocation.

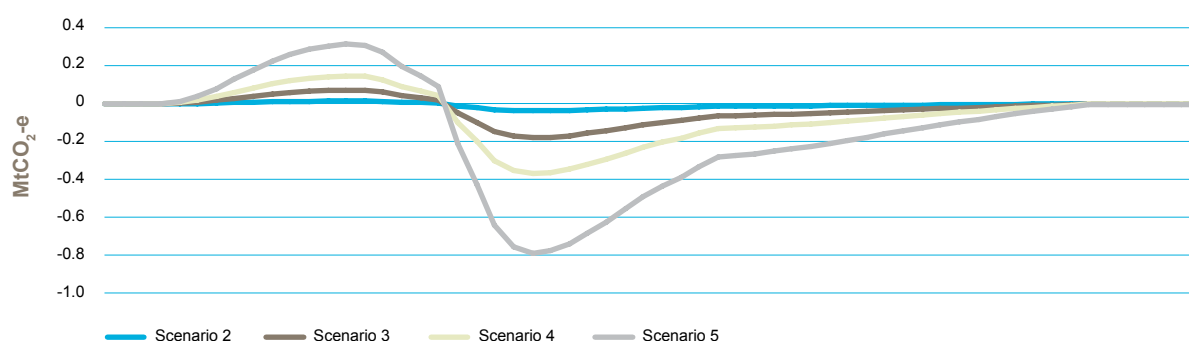
Under scenario 5 (a 33% re-allocation from domestic, commercial and transport refrigeration), approximately 1 million tonnes is re-directed in equal proportions towards aerosols, foam and fire protection. This results in a reduction in emissions of 0.79 million tonnes CO₂-e in domestic, commercial and transport refrigeration and a corresponding increase of 0.66 million tonnes in aerosols, foams and fire protection. The residual gas is accounted for as gas recovered and destroyed and stock change in the bank of gas in operating equipment.

Figure 4.8: Halocarbons: results of sensitivity testing of allocation assumptions: 2008 (Mt CO₂-e)



Total cumulative differences in emissions and destruction under each allocation scenario between 1990 and 2050 (where the last of the current stock of operating equipment is retired) are shown in Figure 4.9. The chart shows that while differences occur in emissions in individual years the total gas either emitted or destroyed is unchanged over the life of each equipment type. The gas end-use re-allocation results in an increase in emissions for years where imports are occurring (up to 2009 in the case of this test), followed by a decrease in emissions relative to the base assumption from 2009 onwards.

Figure 4.9: Halocarbons: results of sensitivity testing of allocation assumptions: 1990–2050 (Mt CO₂-e)



As information about servicing and replenishment practices is limited, the replenishment assumptions have been devised by DCCEE.

The effect of assumptions about gas replenishment was tested by reducing the replenishment rates for all sources where replenishment occurs by 10, 20, 30 and 50 per cent.

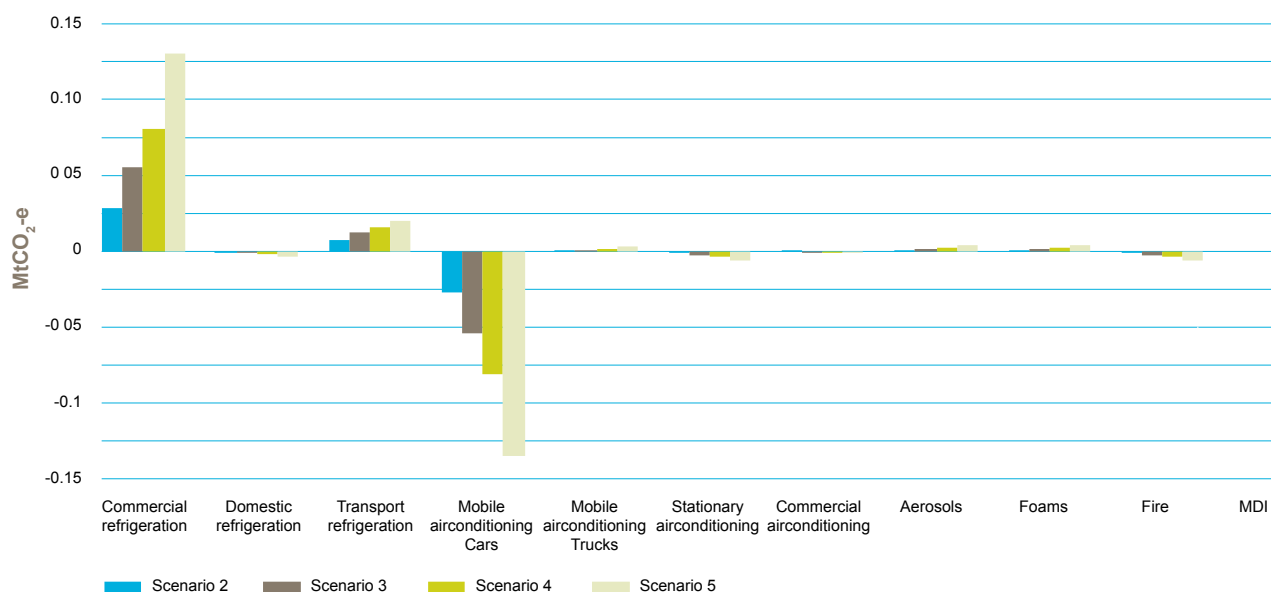
As with bulk gas allocation, the total emissions estimate was found to be insensitive to changes in assumed replenishment rates with a 50% reduction in replenishment resulting in only a 0.25% change in total emissions in 2008. The effects of changes to the replenishment assumptions on total emissions within the model, while minimal, are complex. The total gas allocated to equipment is unchanged under these scenarios such that when less gas is allocated to replenishment, more is available to be allocated to new equipment.

Figure 4.10 shows that emissions from commercial refrigeration increase as a result of a reduction in the general rates of replenishment as more gas is allocated to new equipment for this category. However, for domestic refrigeration, mobile air conditioning in cars and domestic stationary air conditioning the gas stocks are affected by the quantity of gas being replenished and thus, as a result of less gas being replenished, the gas bank and therefore emissions are lower for these categories.

Table 4.40: Halocarbons: results of sensitivity testing of replenishment assumptions (Mt CO₂-e)

	Replenishment assumptions				
	Base	Case 1	Case 2	Case 3	Case 4
Replenishment rate	100%	90%	80%	70%	50%
Emissions in 2008 (Mt CO ₂ -e)					
Commercial refrigeration	3.23	3.26	3.28	3.31	3.36
Domestic refrigeration	0.04	0.04	0.04	0.04	0.04
Transport refrigeration	0.31	0.32	0.32	0.33	0.33
Mobile air conditioning cars	0.86	0.84	0.81	0.78	0.73
Mobile air conditioning trucks	0.19	0.19	0.19	0.19	0.19
Stationary air conditioning	0.62	0.61	0.61	0.61	0.61
Commercial air conditioning	0.06	0.06	0.06	0.06	0.06
Aerosols	0.13	0.14	0.14	0.14	0.14
Foams	0.13	0.13	0.13	0.13	0.13
Fire	0.05	0.05	0.05	0.05	0.05
Metered dose inhalers	0.14	0.14	0.14	0.14	0.14
Total	5.75	5.76	5.77	5.77	5.77
% change on base case		0.17%	0.24%	0.25%	0.25%

Figure 4.10: Halocarbons: results of sensitivity testing of replenishment assumptions – change in emissions 2008 (Mt CO₂-e)

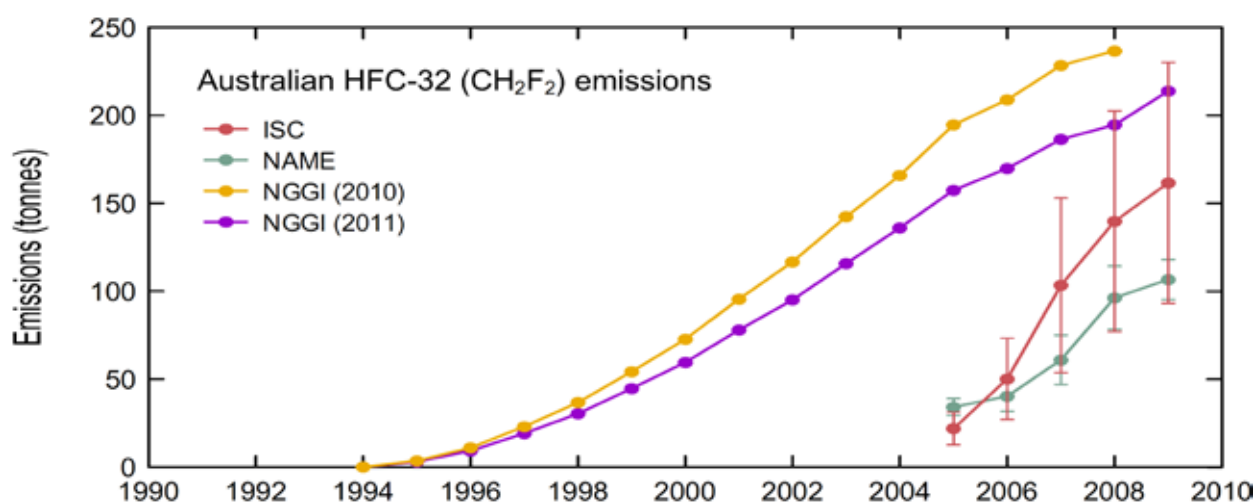


External verification through atmospheric testing

Monitoring of atmospheric HFC concentrations has been undertaken by the Commonwealth Scientific and Industrial Research Organisation (CSIRO) at the Cape Grim Baseline Air Pollution Station in Tasmania since the mid 1990's. The DCCEE has commissioned CSIRO to verify its annual estimates of HFC emissions in the Inventory.

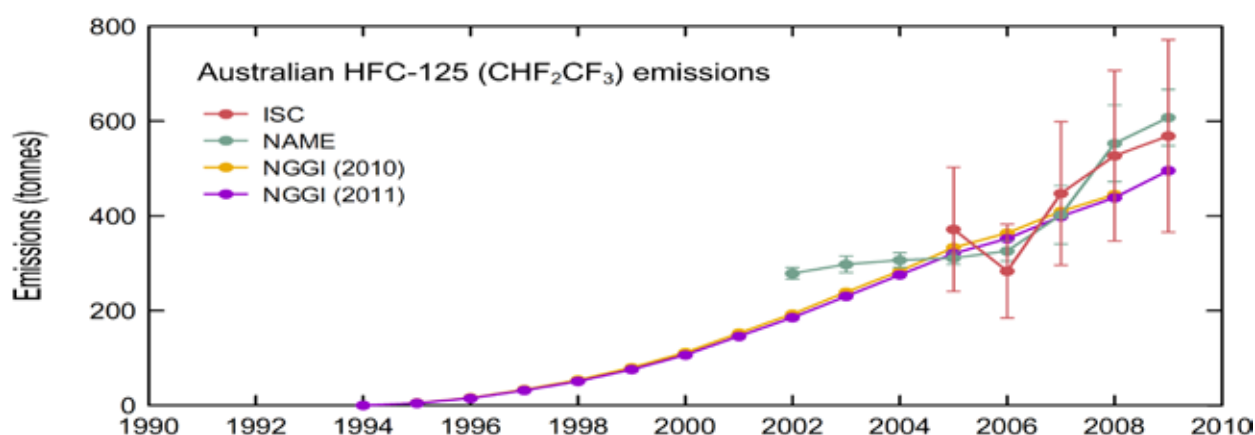
The verification process confirmed that total HFC, SF₆ and PFC emissions in the Inventory are in excellent agreement (within ~3%) with estimates based on Cape Grim data (CSIRO 2011). However, verification of individual HFC species produced varying results and suggests that further work is required to better determine the relative levels of gas species. Figures 4.11, 4.12, 4.13 and 4.14 show the comparisons of estimates based on Cape Grim measurements with inventory estimates by HFC species. Figure 4.15 shows the comparison of total emissions.

Figure 4.11: Comparison of HFC-32 Inventory estimates (NGGI 2011) with estimates derived from Cape Grim measurement data (ISC and NAME)



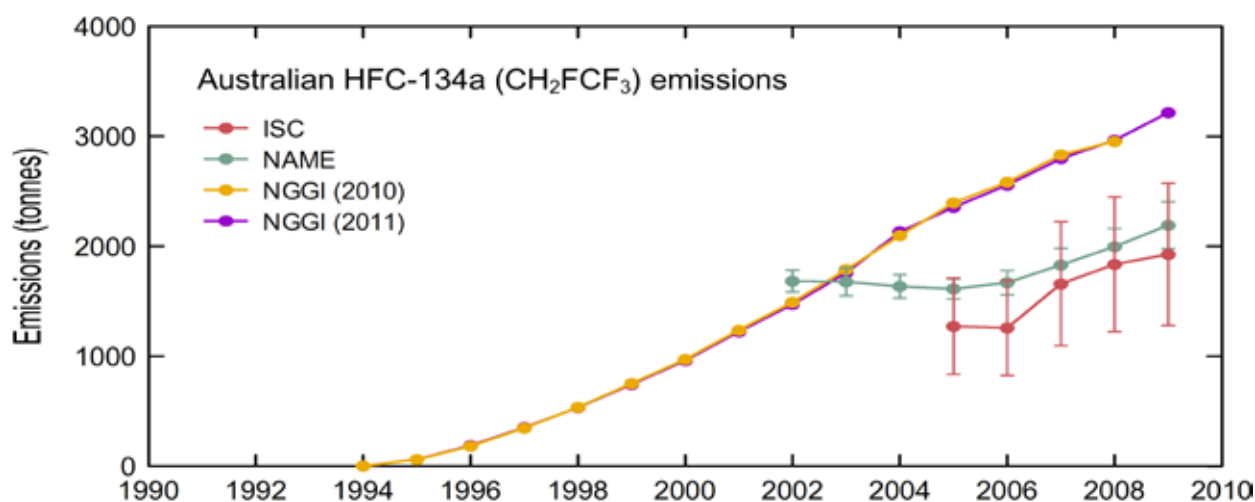
Source: CSIRO 2011

Figure 4.12: Comparison of HFC-125 Inventory estimates (NGGI 2011) with estimates derived from Cape Grim measurement data (ISC and NAME)



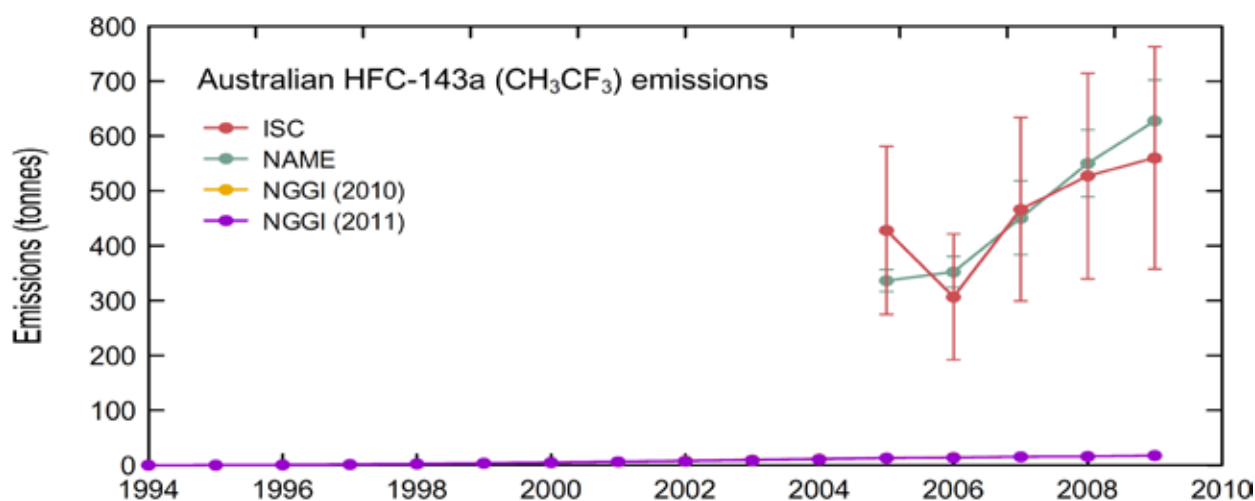
Source: CSIRO 2011

Figure 4.13: Comparison of HFC-134a Inventory estimates (NGGI 2011) with estimates derived from Cape Grim measurement data (ISC and NAME)



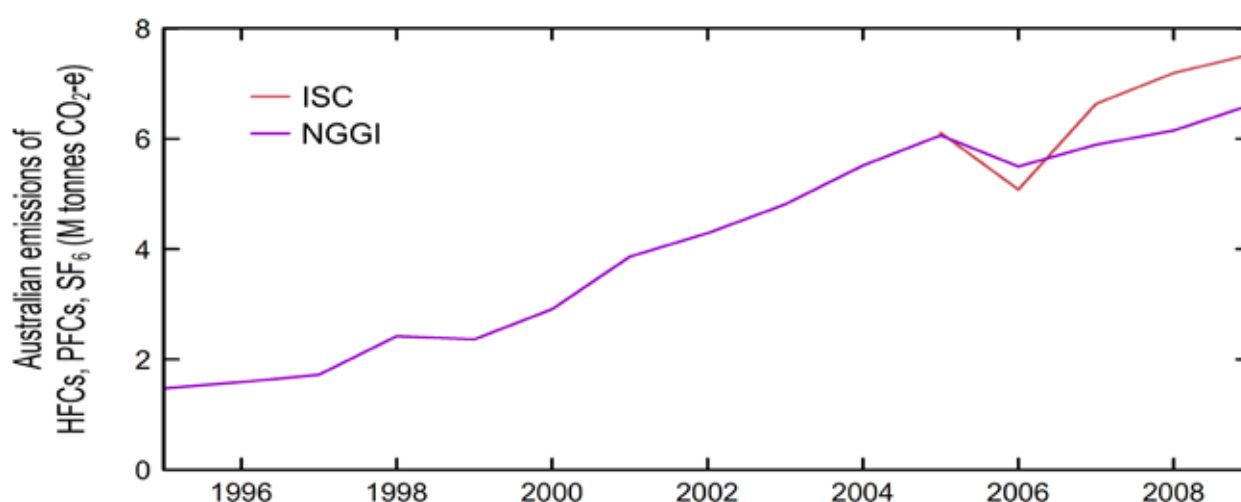
Source: CSIRO 2011

Figure 4.14: Comparison of HFC-143a Inventory estimates (NGGI 2011) with estimates derived from Cape Grim measurement data (ISC and NAME)



Source: CSIRO 2011

Figure 4.15: Comparison of total HFC, PFC and SF₆ emissions from Cape Grim Observations (ISC) and the National Accounts (NGGI)



Source: CSIRO 2011

4.8.5 Recalculations Since The 2008 Inventory

Recalculations of the estimates of emissions from the consumption of HFCs occurred for the following reasons:

- Revisions to motor vehicle stock and disposal estimates based on updated vehicle census data from the ABS.
- Use of NGERS data for the use of SF₆ in electricity supply and distribution
- Estimation of emissions from the use of SF₆ in other applications

Table 4.41: 2.F Consumption of halocarbons and SF₆: recalculation of total CO₂-e emissions (Gg), 1990-2008

	2010 Submission	2011 Submission	Change	Change
	Gg CO ₂ -e	Gg CO ₂ -e	Gg CO ₂ -e	%
2.F Consumption of halocarbons and SF ₆				
1990	521	56	- 465	-89.3%
1991	521	56	- 465	-89.2%
1992	521	56	- 465	-89.2%
1993	521	56	- 465	-89.2%
1994	522	57	- 464	-89.0%
1995	629	164	- 465	-73.9%
1996	842	380	- 462	-54.9%
1997	1,139	671	- 469	-41.1%
1998	1,481	998	- 482	-32.6%
1999	1,885	1,389	- 497	-26.3%
2000	2,315	1,808	- 506	-21.9%
2001	2,838	2,319	- 519	-18.3%
2002	3,341	2,808	- 532	-15.9%
2003	3,922	3,369	- 552	-14.1%
2004	4,517	4,046	- 470	-10.4%
2005	5,114	4,526	- 589	-11.5%
2006	5,476	4,909	- 567	-10.4%
2007	5,968	5,393	- 574	-9.6%
2008	6,273	5,769	- 504	-8.0%

4.8.6 Planned Improvements

Refinements to the *NGERS Measurement Determination 2008* have been made to allow reporting facilities to estimate their emissions on the basis of gas stock balance and ‘top-up’ models. It is anticipated that NGERS will provide country specific information on annual leakage rates from commercial and industrial refrigeration and air conditioning applications as well as gas insulated switch gear and circuit-breaker applications. Once this information becomes available, it will be assessed for potential use in future inventory submissions.

CSIRO verification of Inventory HFC estimates showed that total emissions are in excellent agreement with estimates based on independently measured atmospheric HFC concentrations. However further work is required in order to better estimate individual gas species. The DCCEE will continue its ongoing efforts to improve speciation of its HFC estimates in general and will also undertake further analysis in order to develop gas speciation data for emissions estimates from foam blowing, fire extinguishers, aerosols and solvents.

5. SOLVENT AND OTHER PRODUCT USE

5.1 OVERVIEW

Activities in the *solvent and other products use* source category results in emissions of N₂O and NMVOCs. Total net emissions estimated from *solvents and other products* were 153 Gg NMVOC in 2009 (Table 5.1). The main determinant of *solvent and other product use* emissions from year to year is the quantity of the relevant product that is produced or used.

Table 5.1: Solvent and other product use NMVOC emissions 2009

Greenhouse Gas Source and Sink Categories	N ₂ O Emissions (Gg)	NMVOC emissions (Gg)
3 SOLVENT AND OTHER PRODUCT USE	IE	153
A Paint application	NA	65
B Degreasing and dry cleaning	NA	38
C Chemical products manufacture and processing	NA	1
D Other	IE	50

Note: Confidential N₂O emissions are reported in the industrial processes sector – 2.B Chemical Industry

Paint application was the main source of NMVOCs contributing 42.2% (64.7 Gg NMVOC) of the sector's emissions. The next largest source was *other* (32.6%), followed by *degreasing and dry cleaning* (24.6%). *Chemical products manufacture and processing* contributed only 0.5%.

Emissions from *solvent and other products use* decreased by 0.7 Gg (0.5%) between 1990 and 2009. reductions in emissions from paint application have been offset to some degree by increases in emissions from degreasing and dry cleaning and other.

5.2 SOURCE CATEGORY DESCRIPTION

Surface coating operations involve the application of paint, varnish, lacquer or paint primer for decorative or protective purposes. Thinning solvents are normally used to dilute surface coating formulations or for cleaning purposes. Surface cleaning or degreasing operations involve the removal of materials such as oils, grease, waxes and moisture from surfaces. Chemical products manufacture and processing covers paint and ink manufacturing. General solvent use and consumer cleaning by the domestic and commercial sectors covers a large range of products including Domestic and Commercial Aerosol Products; Other Domestic and Commercial Products; and, Consumer Cleaning Products.

5.3 METHODOLOGY

5.3.1 Paint Application (3.A)

Surface coating operations involve the application of paint, varnish, lacquer or paint primer for decorative or protective purposes. Thinning solvents are normally used to dilute surface coating formulations or for cleaning purposes. Collectively, surface coatings and thinners represent the most significant solvent use category.

Surface coatings have differing solvent content. Average densities and emission factors are given in Table 5.2.

Table 5.2: Properties of surface coatings and thinners

Category	Average Density (kg/l)	EF kg NMVOC /kg Product
Architectural/Decorative	1.33	0.11
Industrial	1.24	0.44
Thinners	0.77	1.00

Source: Australian Paint Manufacturers Federation (pers.comm, 1995)

5.3.2 Degreasing And Dry Cleaning (3.B)

Surface cleaning or degreasing operations involve the removal of materials such as oils, grease, waxes and moisture from surfaces. In the commercial/industrial sectors, surface cleaning/degreasing is normally conducted in preparation for other treatment such as painting and electroplating.

There are three basic types of surface cleaning operations, which are used by commerce and industry. These operations are referred to as cold cleaning, vapour cleaning and in-line or conveyorised cleaning. Emission factors for surface cleaning/degreasing and dry cleaning are presented in Table 5.3. Population figures are mean populations for the relevant financial year.

Table 5.3: Surface cleaning/degreasing and dry cleaning emission factors

Category	Emission Factor kg NMVOC/capita/yr
Cold Cleaning, Automobile Repair	1.14 ^(a)
Cold Cleaning, Manufacturing	0.50 ^(a)
Dry Cleaning	0.17 ^(b)
Total	1.81

(a) USEPA (1991b). (b) ICI (pers.comm, 1994)

5.3.3 Chemical Products Manufacture And Processing (3.C)

Chemical products manufacture and processing covers paint and ink manufacturing. Emission factors for chemical products manufacture and processing are presented in Table 5.4.

Table 5.4: Emission factors for chemical products manufacture and processing

Category	EF _i kg NMVOC/tonne Product
Paint Manufacturing	3 ^a
Printing Ink Production	9 ^b

(a) Dulux (pers.comm, 1995). (b) Mid-range value reported for USA facilities, Buonicore and Davis (1992)

5.3.4 Other (3.D)

General solvent use and consumer cleaning by the domestic and commercial sectors covers a large range of products. These products include aerosol spray products such as insect sprays, hairsprays and household carpet cleaners, and non-aerosol spray products such as adhesives, waxes and general household solvents. Also included in this source category is the use of N₂O in aerosol products and for medical use.

N₂O emissions from Aerosol Products and Anaesthesia

Emissions of N₂O from aerosol products and anaesthesia are based on production data provided by the industrial gas manufacturers (BOC and Air Liquide). These data and the resultant emissions estimates are confidential and are included in the 2.B.5 *confidential chemical Industry emissions* in the Industrial Processes sector .

NM VOC Emissions from General Solvent use and Consumer Cleaning

Per-capita emission factors provided by the Aerosol Association of Australia (AAA, pers.comm, 1994) have been used for Domestic/Commercial Aerosol Products and US data for aerosol spray products and miscellaneous non-aerosol spray products (USEPA 1997) have been adopted for estimating NMVOC emissions from Other Domestic/Commercial products and Consumer Cleaning Products.

Emission factors for general solvent use and consumer cleaning products are presented in Table 5.5.

Table 5.5: Emission factors for general solvent use and consumer cleaning products

Product	Emission Factor kg NMVOC/capita/yr
Domestic/Commercial Aerosol Products ^a	
Insect/Garden Sprays	0.33
Hair Sprays	0.26
Personal Deodorants	0.17
Air Fresheners	0.14
Miscellaneous Automotive/Industrial	0.17
General Household	0.13
Other	0.10
Sub Total	1.30
Other Domestic/Commercial Products ^b	
Adhesives and Sealants	0.21
Coatings and Related Products	0.35
Sub Total	0.56
Consumer Cleaning Products ^b	
Automotive Products	0.51
Miscellaneous Products	0.03
Sub Total	0.54
Total	2.40

(a) Aerosol Association of Australia (pers.comm, 1994). (b) USEPA (1997)

NMVOC emissions from general solvent use and consumer cleaning products are reported in table 5.5. The mean population for the financial year is multiplied by the emission factor and the result is expressed in gigagrams (Gg). Emission factors are expressed in terms of per capita use per year.

5.3.5 Uncertainties And Time Series Consistency

The tier 1 uncertainty analysis in Annex 7 provides estimates of uncertainty according to IPCC source category and gas. Time series consistency is ensured by use of consistent models, model parameters and datasets for the calculations of emissions estimates.

5.3.6 Source Specific QA/QC

This source category is covered by the general QA/QC of the greenhouse gas inventory in Section 1.6

5.3.7 Recalculations Since The 2008 Inventory

Minor recalculations have occurred in the solvent and other product use sector due to minor revisions in population estimates throughout the time-series which are used for the estimation of emissions from degreasing and dry cleaning and general solvent use and consumer cleaning.

Table 5.6: Recalculation of NMVOC emissions (Gg), 1990-2008

	1990	2000	2001	2002	2003	2004	2005	2006	2007	2008
2010 submission	154	148	144	146	155	153	151	156	153	153
2011 submission	154	148	144	146	156	154	151	157	154	156
Change	0	0	0	0	1	1	1	1	1	2
Percent Change	0%	0%	0%	0%	0%	0%	0%	-1%	-1%	-2%

5.3.8 Planned Improvements

All activity data, methodologies and emission factors are kept under review.

6. AGRICULTURE

6.1 OVERVIEW

Agriculture produced an estimated 84.7 Mt CO₂-e emissions or 15.5% of net national emissions (excluding LULUCF) in 2009 (Table 6.1).

Enteric fermentation was the main source of *agriculture* emissions contributing 64.6% (54.7 Mt CO₂-e) of the sector's emissions. The next largest source was *agricultural soils* (16.7%), followed by *prescribed burning of savannas* (14.3%) and *manure management* (3.9%). *Rice cultivation* and *field burning of agricultural residues* contributed less than 1% of the sector's emissions.

The *agriculture* sector is the dominant national source of both methane and nitrous oxide – accounting for 57.9% and 75.9% respectively of the net national emissions (excluding LULUCF) for these two gases.

Table 6.1: Agriculture sector CO₂-e emissions, 2009

Greenhouse gas source and sink categories	CO ₂ -e emissions (Gg)			
	CO ₂	CH ₄	N ₂ O	Total
4 AGRICULTURE	NA	65278	19467	84746
A Enteric fermentation	NA	54736	NA	54736
B Manure management	NA	1752	1564	3316
C Rice cultivation	NA	46	NA	46
D Agricultural soils	NA	NA	14191	14191
E Prescribed burning of savannas	NA	8530	3617	12147
F Field burning of agricultural residues	NA	214	95	309

Trends

Emissions from *agriculture* decreased by 2.4% (2.1 Mt) between 1990 and 2008 and had declined by 3.6% (3.2 Mt) between 2008 and 2009 (Figure 6.1).

Enteric fermentation emissions declined by 14.4% (9.2 Mt) between 1990 and 2009 (Figure 6.2). The decline in emissions in the early 1990s was principally driven by a fall in sheep numbers, however by the late 1990s the emissions had begun to increase as the numbers of beef cattle began to rise, reflecting changing relative returns to each industry. Since 2002 many animal populations have been declining in response to the prolonged drought conditions which have occurred over extensive areas of southern and eastern Australia. Between 2008 and 2009, emissions from *enteric fermentation* declined by 2.5% (1.4 Mt).

Figure 6.1: CO₂-e emissions from agriculture, 1990–2009

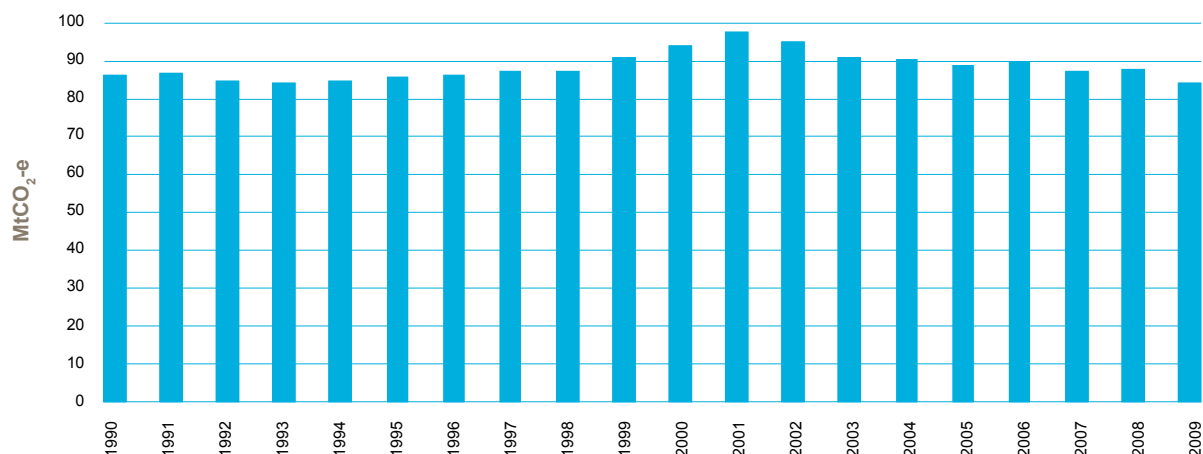
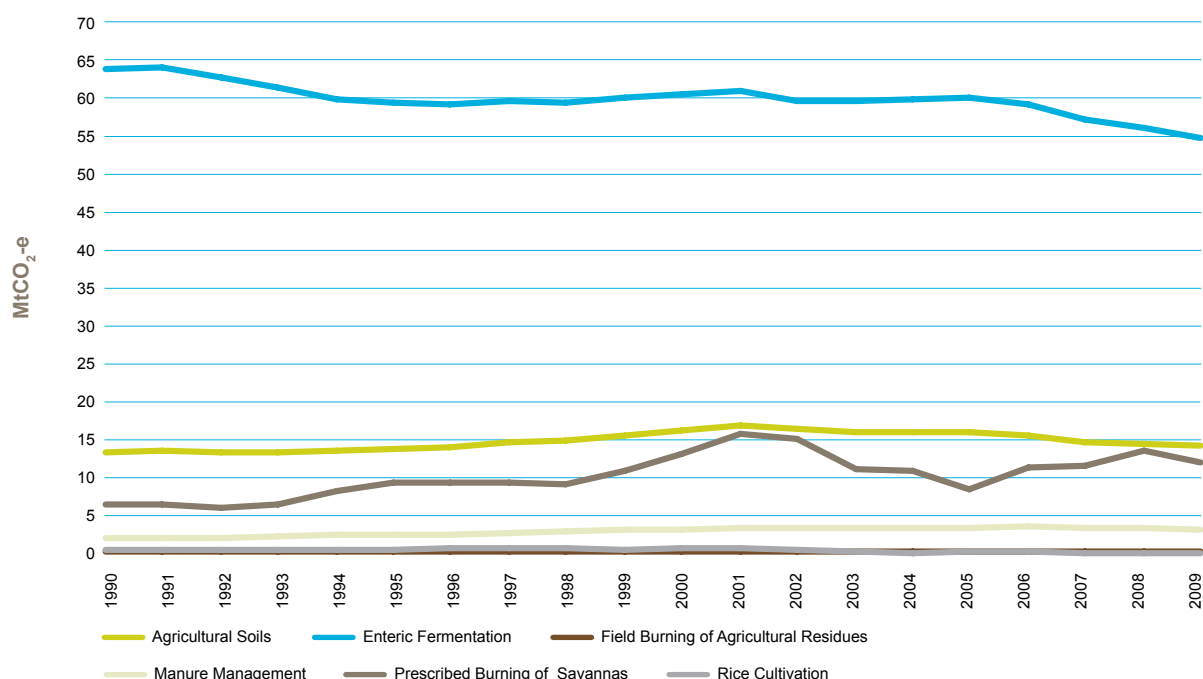


Figure 6.2: CO₂-e emissions from agriculture, by sub-sector, 1990–2009



Manure management emissions have actually increased by 60.6% (1.3 Mt) between 1990 and 2009 due to the strong growth in the intensive feedlot cattle industry over this period. Manure management emissions declined by 2.1% (0.1 Mt) between 2008 and 2009 reflecting the recent decline in populations of dairy and feedlot cattle and pigs.

As all rice cultivation in Australia is flood irrigated this industry has been severely affected by the ongoing drought. Emissions from *rice cultivation* in 2009 were 90.6% (0.4 Mt) lower than in 1990 and 7.4% (0.003 Mt) higher than in 2008.

Agricultural soils emissions have increased by 5.6% (0.8 Mt) between 1990 and 2009, but have declined by 1.9% (0.3 Mt) between 2008 and 2009. Since 1990 there has been a significant increase in the use of synthetic fertilisers, and increased retention of crop residues as the sugar cane industry has shifted towards green cane harvesting. The growth in the intensive livestock industries has also led to an increase in the amount of wastes applied to soils. As crop production, animal populations and fertiliser use has reduced during the ongoing drought, emissions have declined since 2008.

Emissions from *prescribed burning of savannas* peaked in 2001 and have declined significantly since then. This trend will not necessarily continue in the future, as emissions from this source are highly variable and largely reflect variations in climatic conditions (Meyer 2004).

The emissions from *field burning of agricultural residues* have increased by 5.9% (0.02 Mt) between 1990 and 2009 and increased by 12.6% (0.03 Mt) between 2008 and 2009.

6.2 OVERVIEW OF SOURCE CATEGORY DESCRIPTION AND METHODOLOGY – AGRICULTURE

CH₄ and N₂O emissions from livestock industries are reported under the enteric fermentation (4A) and manure management (4B). In Australia, the principal species comprise cattle and sheep, with breeds chosen to operate within pasture and paddock management systems and, in many cases, in semi-arid or tropical and sub-tropical climatic conditions. Typical animal performance, as a consequence, tends to vary significantly from those of other Annex I countries.

Other agricultural sources include methane emissions from rice cultivation (4C), nitrous oxide emissions from agricultural soils (4D), methane and nitrous oxide emissions from the burning of savannas (4E) and agricultural crop residues (4F). Emissions of CH₄, N₂O, NO_x, CO and NMVOCs from these sources are produced when living and dead biomass is eaten, consumed, decays or is burnt. These emissions are modified by human activities including cultivation, addition of fertilisers, deliberate burning and by the introduction of ruminant animals.

The Australian agriculture methodology contains both country specific and IPCC default methodologies and emission factors (Table 6.2).

The agriculture inventory is compiled on a State basis with State emission totals then aggregated to give national totals. The inventory is compiled in this way to reduce errors associated with averaging input data across areas with large physical and management differences. Australia has a land area of 769 million hectares which cover a wide range of climate zones and soil and vegetation types (see section 7.3.1 for more details). These large physical differences lead to significant differences between States in such things as fuel loads for fires and the quality and availability of feed and the performance of animals throughout the year. In addition as the States range from warm to temperate in climate the , methane conversion factors for manure management systems can vary significantly between the States.

In accordance with the IPCC (1997, vol. 1) guidelines, emissions are reported as three year averages where possible (i.e. the reported emissions for 1990 are the average of those for 1989, 1990 and 1991).

6.2.1 Data Sources

Data on livestock numbers and crop production are sourced from Australian Bureau of Statistic (ABS) census/survey data collected on 30 June in the relevant year. Relevant industry associations provide some rice and sugar cane data. Data on the number of feedlot cattle are sourced from the Australian Lot Feeders Association (ALFA) quarterly survey. Milk production statistics come from Dairy Australia.

The areas of savanna burnt are largely based on a time series of AVHRR satellite images processed by WA Department of Land Information. The assessments of the area burnt in some areas (particularly for earlier years up to the mid-1990s) are based on field surveys of State and Territory departments with responsibility for managing fires.

Other primary data used in the algorithms (liveweights, liveweight gains, pasture digestibility, savanna fuel loads, combustion emission factors, etc) are based on reviews of published data and expert assessments.

Process for eliciting expert assessments

Given the extensive nature of most of Australia's agricultural production there are few if any comprehensive State databases of information such as animal and pasture characteristics. As this data is required to estimate emissions, it has been necessary to use expert assessments to determine appropriate country specific information. The initial Australian methods and data were developed by expert working groups with representation from government, industry, academia and research institutions. The record of participants and outcomes are documented in the *Australian methodology for the Estimation of Greenhouse Gas Emissions and Sinks, Agriculture, Workbook for Non-Co₂ emissions from the biosphere 5.0* (NGGIC 1994a) and *Workbook for Livestock 6.0* (NGGIC 1994b). The livestock categories contain a large number of expert assessments and these values were reviewed/updated in 1995 (documented in *Workbook for Livestock 6.1* (NGGIC 1996) and again in 2000/01 (documented in Howden et al. (2002) and White (2002)). In each case consultants were used to coordinate the review. The consultant elicited expert assessments either through round table meetings with the experts or through surveys. These assessments were then compiled by the consultants and an agreed value recirculated to experts for final comment. The consultants also undertook a number of reality checks on the expert assessments to ensure that correlated values such as seasonal liveweights and daily liveweight gains, and pasture digestibility and crude protein contents were internally consistent (White 2002).

Comparison with international data

The ABS annually report agricultural data to the Food and Agriculture Organisation (FAO) of the United Nations. Some divergence occurs from year to year between the activity data in the inventory CRF tables and those published by the FAO. The reasons for these differences are as follows:

- Over the time frame of the inventory the ABS has changed the threshold of the estimated value of agricultural operations (EVAO) used to determine which agricultural operations are included in the census/survey. In the years 1989 to 1993 ABS used EVAO's of \$20-25,000 which is considerably higher than the \$5000 used since 1994. To ensure time series consistency in the data, a multiplier is applied to adjust the animal numbers to reflect the smaller farms that will have been left out of the 1989-1993 censuses;
- For the 2005-06 census the ABS introduced a new survey frame sourced from the Australian Taxation Office's Australian Business Register (ABR). Due to the progressive deterioration of the previous frame (based on a register of agricultural establishments maintained by ABS) the coverage of the two frames differed. To ensure time-series consistency, bridging estimates developed by ABS were used to revise animal numbers for dairy cattle, range kept cattle, sheep and pigs and production data for wheat, barley, oats, sorghum and other crops from 2002 to 2005.

Table 6.2a: Summary of methods and emission factors: Agriculture (CH₄ and N₂O)

Greenhouse Gas Source and Sink Categories		CH ₄		N ₂ O	
		Method Applied	Emission Factor	Method Applied	Emission Factor
A Enteric Fermentation					
1	Cattle				
	a. Dairy Cattle	CS, T2	CS		
	b. Free-range Beef Cattle	CS, T2	CS		
	c. Feedlot Cattle	CS, T2	CS		
2	Buffalo	T1	IPCC		
3	Sheep	CS, T2	CS		
4	Goats	T1	IPCC		
5	Camels and Llamas	T1	IPCC		
6	Horses	T1	IPCC		
7	Mules and Asses	T1	IPCC		
8	Swine	CS, T2	CS		
9	Poultry	NE	NE		
10	Other (alpacas, deer, emus and ostriches)	T1	CS		
B Manure Management					
1	Cattle				
	a. Dairy Cattle	T2	IPCC, CS	CS, T2	IPCC
	b. Free-range Beef Cattle	CS, T2	CS	CS, T2	IPCC
	c. Feedlot Cattle	T2	IPCC	CS, T2	IPCC
2	Buffalo	CS, T2	CS	CS, T1	IPCC
3	Sheep	CS, T2	CS	CS, T2	IPCC
4	Goats	CS, T2	CS	CS, T1	IPCC
5	Camels and Llamas	CS, T2	CS	CS, T1	IPCC
6	Horses	CS, T2	CS	CS, T1	IPCC
7	Mules and Asses	CS, T2	CS	CS, T1	IPCC
8	Swine	M	IPCC	M	IPCC
9	Poultry	T2	IPCC	CS, T1	IPCC
10	Other (alpacas, deer, emus and ostriches)	CS, T2	CS	CS, T1	IPCC
C Rice Cultivation		T1	CS		
D Agricultural Soils					
1	Direct Soil Emissions				
	a. Synthetic Fertilisers			T2	CS
	b. Animal Wastes Applied to Soils			T2	CS
	c. N-fixing crops			T1	IPCC
	d. Crop Residues			T1	IPCC
	e. Cultivation of Histosols			T1	IPCC
2	Animal Production			T2	CS
3	Indirect Emissions				
	a. Atmospheric Deposition			T1	IPCC
	b. Leaching and Run-off			CS	CS
E	Prescribed Burning of Savannas	CS	CS	CS	CS
F	Field Burning of Agricultural Residues	CS	CS	CS	CS

Table 6.2b: Summary of methods and emission factors: Agriculture (NO_x, CO and NMVOC)

Greenhouse Gas Source and Sink Categories		NO _x		CO		NMVOC	
		Method Applied	EF	Method Applied	EF	Method Applied	EF
E	Prescribed Burning of Savannas	CS	CS	CS	CS	CS	CS
F	Field Burning of Agricultural Residues	CS	CS	CS	CS	CS	CS

EF = emission factor, CS = country specific, IPCC = IPCC defaults, T1 = Tier 1, T2 = Tier 2, M = Model and NE= not estimated

6.3 SOURCE CATEGORY 4.A ENTERIC FERMENTATION

6.3.1 Source Category Description

Methane is produced in herbivores as a by-product of enteric fermentation, a digestive process by which plant material consumed by an animal is broken down by bacteria in the gut under anaerobic conditions. A portion of the plant material is fermented in the rumen to simple fatty acids, CO₂ and CH₄. The fatty acids are absorbed into the bloodstream, and the gases vented by eructation and exhalation by the animal. Unfermented feed and microbial cells pass to the intestines.

6.3.2 Methodology

Australia has identified Enteric Fermentation as a key source category using the *Tier 1* level and trend assessments as recommended in the IPCC Good Practice Report (IPCC 2000). In accordance with IPCC Good Practice requirements *Tier 2* methods are therefore used, to estimate enteric fermentation emissions from the major livestock sub-categories.

The proportion of intake that is converted into methane is dependent on both the characteristics of the animal, feed and the amount of feed eaten. The IPCC (2000) *Tier 2* approach has fixed methane conversion rates for each livestock category. Given the heterogeneity of feed types available within Australia it was considered important to use methodologies developed in Australia or methodologies that could reflect these differences, hence country specific methodologies have been used.

6.3.2.1 Cattle (4.A.1)

Pasture Fed

It was considered that the approach developed by Blaxter and Clapperton (1965) would be the most appropriate method for estimating methane emissions from pasture feed cattle in temperate Australia. This approach was selected as it has been widely used in methane inventories and is capable of reflecting the effect of feed quality on methane emissions. Although based on animals fed on diets representative of those in the UK, the digestibility of the feeds studied are within the ranges found in temperate Australia.

The equations derived by Blaxter and Clapperton (1965) do not reflect the breeds and feed types used in the tropics and sub-tropics of Australia. In a study of *Bos indicus* (Brahman) cattle on tropical feeds conducted in Australia, Kurihara et al. (1999) and Kurihara et al. (2006) found methane conversion rates that were significantly higher than the IPCC (2000) defaults (7%) or those calculated by Blaxter and Clapperton (1965). This approach has been used to estimate emissions from beef cattle in tropical areas.

The Blaxter and Clapperton (1965) approach requires the estimation of gross energy intake and then calculates the proportion of this energy that is converted into methane, based on the digestibility at maintenance of the feed energy⁶ and the level of feed intake relative to that required for maintenance. The figure for methane can then be expressed on an equivalent mass basis, using the conversion factor of 55.22 MJ/kg CH₄ (Brouwer 1965). The Kurihara et al. (1999) approach relates daily methane production with dry matter intake.

⁶ It should be noted that Blaxter and Clapperton (1965) use a figure for the digestibility of feed energy which is slightly smaller than the dry matter digestibility values used for pasture in Australia. However, there is no consolidated data that will allow conversion to be made between the two digestibility classes

Accurate estimation of intake is difficult as it depends on many factors. The IPCC (2000) *Tier 2* methods for estimating intake are based on work undertaken in the northern hemisphere (NRC 1996). This method was considered inappropriate for Australia and in line with good practice (IPCC 2000 pg 4.13) a country-specific method (Minson and McDonald 1987) based on research in Australia is used instead. Minson and McDonald (1987) have derived an equation that estimates feed intake relative to liveweight and liveweight gain of cattle. This relationship includes data for cattle on both tropical and temperate feeds. By setting the liveweight gain component of their equation to zero to calculate maintenance intake, it is possible to estimate the relative feed intake of cattle.

Lot Fed

Moe and Tyrrell (1979) devised equations for predicting methane emissions from dairy cattle fed diets consisting mostly of high digestibility grains and concentrates and high quality forages. The Moe and Tyrrell (1979) equation used relates methane production to the intake of three components of the dietary carbohydrate. These are: soluble residue, hemicellulose and cellulose. As feedlot cattle in Australia are fed diets consisting of high digestibility grains and concentrates, the Moe and Tyrrell (1979) equations were considered appropriate for estimating emissions.

Dairy Cattle

The large volumes of milk produced by dairy cattle under modern management regimes, nearly 5000 litres per annum (see Appendix 6.A.8), requires that they consume considerably more feed than an equivalent non-lactating cow. The approach used was to compile the average daily milk production per head of milking cows in each State, then calculate the increased energy requirements needed to produce this milk from relationships presented by the Standing Committee on Agriculture (SCA) (1990).

Table 6.3: Symbols used in algorithms for dairy cattle

State (i)	Dairy Cattle Classes (age) (j)	Season ^b (k)
i = 1 NSW/ACT	j = 1 Milking Cows ^(a)	k = 1 Spring
i = 2 Tasmania	j = 2 Heifers > 1 year	k = 2 Summer
i = 3 Western Australia	j = 3 Heifers < 1 year	k = 3 Autumn
i = 4 South Australia	j = 4 House Cows Milk and Dry	k = 4 Winter
i = 5 Victoria	j = 5 Bulls > 1 year	
i = 6 Queensland	j = 6 Bulls < 1 year	
i = 7 Northern Territory		

(a) Includes cows used for milk production but not currently lactating. (b) This category relates to Tasmania (i = 2) only. Data was not available to support disaggregation of the other States.

The equation presented in Minson and McDonald (1987) calculates feed intake of non-lactating cattle from liveweight and liveweight gain data. For lactating cattle the additional intake for milk production (MI_{ijk}) is included to give total intake (I_{ijk} kg dry matter/head/day):

$$I_{ijk} = (1.185 + 0.00454W_{ijk} - 0.0000026W_{ijk}^2 + 0.315LWG_{ijk})^2 \times MR_1 + M_{ijk} \dots \dots \dots (4A.1a_1)$$

Where: W_{ijk} = weight in kg (Appendix 6.A.1)

LWG_{ijk} = liveweight gain in kg/day (Appendix 6.A.2)

MR_1 = increase in metabolic rate when producing milk (SCA 1990) 1.1 for milking and house cows and 1 for all other classes

The additional intake required for milk production (MI_{ijk} kg DM/head/day) is calculated by:

$$MI_{ijk} = MP_{ijk} \times NE / k_l / q_{m,ijk} / 18.4 \dots\dots\dots (4A.1a_2)$$

Where: MP_{ijk} = milk production (kg/head/day) from Dairy Australia State statistics

$NE = 3.054$ MJ net energy/kg milk (SCA 1990)

$k_l = 0.60$ efficiency of use of metabolizable energy for milk production (SCA 1990)

$q_{m,ijk}$ = metabolizability of the diet. This is the ratio of metabolizable energy (ME) to gross energy (GE) in the diet (i.e. ME / GE). Metabolizable energy content is related to digestibility of dry matter (DMD_{ijk}). So using the equation of Minson and McDonald (1987), $q_{m,ijk} = 0.00795 DMD - 0.0014$; (where DMD is expressed as a %).

The gross energy intake (GEI_{ijk}) is the sum of the intake (I_{ijk}) converted into energy terms assuming a gross energy content of 18.4 MJ/kg (SCA 1990):

$$GEI_{ijk} = I_{ijk} \times 18.4 \dots\dots\dots (4A.1a_3)$$

The intake of the animals relative to that needed for maintenance (L_{ijk}) is calculated as actual intake divided by maintenance intake (ie. intake of non-lactating animal with liveweight gain is set to zero).

$$L_{ijk} = I_{ijk} / (1.185 + 0.00454W_{ijk} - 0.0000026W_{ijk}^2 + (0.315 \times 0)) \dots\dots\dots (4A.1a_4)$$

The percentage of the gross energy intake (GEI_{ijk}) that is yielded as methane (Y_{ijk}) is given by Blaxter and Clapperton (1965) as:

$$Y_{ijk} = 1.3 + 0.112DMD_{ijk} + L_{ijk}(2.37 - 0.050DMD_{ijk}) \dots\dots\dots (4A.1a_5)$$

Where: DMD_{ijk} = digestibility of feed (expressed as a %) (Appendix 6.A.3)

L_{ijk} = intake relative to that needed for maintenance

The total daily production of methane (M_{ijk} kg CH_4 /head/day) is thus:

$$M_{ijk} = Y_{ijk} / 100 \times GEI_{ijk} / F \dots\dots\dots (4A.1a_6)$$

Where: $F = 55.22$ MJ/kg CH_4 (Brouwer 1965)

Annual Australian methane production (Gg) for all classes of dairy cattle across all states can then be calculated as:

$$E = \sum_{i=1}^3 \sum_{j=3-7} (365 \times N_{ij} \times M_{ij}) \times 10^{-6} + \sum_{i=2}^3 \sum_j \sum_k (91.25 \times N_{ijk} \times M_{ijk}) \times 10^{-6} \dots\dots\dots (4A.1a_7)$$

Where: N_{ijk} = numbers of dairy cattle in each class for each State and season

M_{ijk} = methane production (kg/head/day)

Beef Cattle

Table 6.4: Symbols used in algorithms for beef cattle

State (i)	Regions (j)	Season (k)	Beef Cattle Classes (l)
i = 1 NSW/ACT	j = 1 NSW/ACT	k = 1 Spring	l = 1 Bulls > 1 year
i = 2 Tasmania	j = 2 Tas	k = 2 Summer	l = 2 Bulls < 1 year
i = 3 Western Australia	j = 3a South West	k = 3 Autumn	l = 3 Steers < 1 year
	j = 3b Pilbara	k = 4 Winter	l = 4 Cows 1 – 2 year
	j = 3c Kimberley		l = 5 Cows > 2 year
i = 4 South Australia	j = 4 SA		l = 6 Cows < 1 year
i = 5 Victoria	j = 5 Vic		l = 7 Steers > 1 year
i = 6 Queensland	j = 6 Qld		
i = 7 Northern Territory	j = 7 NT		

The equation presented by Minson and McDonald (1987) calculates feed intake (I_{ijkl} kg dry matter/head/day) from liveweight and liveweight gain:

$$I_{ijkl} = (1.185 + 0.00454W_{ijkl} - 0.0000026 W_{ijkl}^2 + 0.315 LWG_{ijkl})^2 \times MA_{ijkl=5} \dots (4A.1b_1)$$

Where: W_{ijkl} = liveweight in kg (Appendix 6.B.1)

LWG_{ijkl} = live weight gain in kg/head/day (Appendix 6.B.2)

Feed intakes can increase by up to 60% during lactation (ARC 1980). For this study, the intake of all breeding cattle was increased by 30% during the season in which calving occurs and by 10% in the following season based on relationships presented in SCA (1990).

The additional intake for milk production ($MA_{ijkl=5}$) is calculated by:

$$MA_{ijkl=5} = (LC_{ijkl=5} \times FA_{ijkl=5}) + ((1-LC_{ijkl=5}) \times 1) \dots (4A.1b_2)$$

Where: $LC_{ijkl=5}$ = proportion of Cows >2 lactating

$FA_{ijkl=5}$ = feed adjustment (Appendix 6.B.5)

As feed dry matter has, on average, a gross energy content of 18.4 MJ/kg (SCA 1990), the dry matter intake is converted into gross energy intake (GEI_{ijkl} MJ):

$$GEI_{ijkl} = I_{ijkl} \times 18.4 \dots (4A.1b_3)$$

The intake of the animals relative to that needed for maintenance (L_{ijkl}) is calculated as actual intake divided by maintenance intake (ie. intake of non-lactating animal with liveweight gain is set to zero).

$$L_{ijkl} = I_{ijkl} / (1.185 + 0.00454 W_{ijkl} - 0.0000026 W_{ijkl}^2 + (0.315 \times 0))^2 \dots (4A.1b_4)$$

The percentage of the gross energy intake that is yielded as methane (Y_{ijkl}) is given by Blaxter and Clapperton (1965) as:

$$Y_{ijkl} = 1.3 + 0.112DMD_{ijkl} + L_{ijkl}(2.37 - 0.050DMD_{ijkl}) \dots (4A.1b_5)$$

Where: DMD_{ijkl} = digestibility of feed (expressed as a %) (Appendix 6.B.3)

L_{ijkl} = feed intake relative to that needed for maintenance

The total daily production of methane (M_{ijkl} , kg CH₄/head/day) for animals on temperate pastures is thus:

$$M_{(i=1-5)(j=1,2,3a,3b,4,5)kl} = Y_{ijkl} / 100 \times GEI_{ijkl} / F \dots\dots\dots (4A.1b_6a)$$

For animals on tropical pastures total daily production of methane (M_{ijkl} , kg CH₄/head/day) is given by Kurihara et al. (1999) and Kurihara et al (2006) as:

$$M_{(i=3,6,7)(j=3c,6,7)kl} = (34.9 \times I_{ijkl} - 30.8) / 1000 \dots\dots\dots (4A.1b_6b)$$

Where: $M_{(i=1-5)(j=1,2,3a,3b,4,5)kl}$ = methane emissions from NSW/ACT, Tas, SA, Vic and the South West and Pilbara regions of WA
 $M_{(i=3,6,7)(j=3c,6,7)kl}$ = methane emissions from Qld, NT and the Kimberley region of WA
 $F = 55.22$ MJ/kg CH₄ from Brouwer (1965)

Feedlot cattle are assumed to originate entirely from the steers > 1 year old beef cattle class. To ensure these cattle are not double counted, it is necessary to subtract feedlot cattle numbers from beef cattle numbers in the steers > 1 year class in that State for the same year. Because feedlot cattle, on average, spend between 75 and 250 days on feedlots prior to slaughtering, an annual equivalent number is derived to represent the figure for feedlot cattle which is subtracted from steers > 1 year old—beef cattle numbers—to calculate emissions. The emissions from the feedlot cattle are calculated in section 2.1.3.

The approach is represented in the following equation:

$$N_{ijkl} = N_{ijk(l=1, l=2, l=3 \dots l=6, [(l=7) - \text{total feedlot numbers}])} \dots\dots\dots (4A.1b_7)$$

Where: N_{ijkl} = numbers of non-feedlot beef cattle in each State, region, season and class.
 $N_{ijk l=1, l=2, l=3 \dots l=6}$ = number of cattle in State i, region j, season k and class l.
 $(l=7) - \text{total feedlot numbers}$ = from Table 6.4, $l=7$ corresponds with steers >1 year old. In order to calculate total beef cattle numbers in this class, total feedlot numbers, adjusted to annual equivalence must be subtracted from $l=7$. For WA 99 per cent of feedlot cattle are assumed to be sourced from the South-West region and the balance from the Pilbara and Kimberley.

Annual Australian methane production (Gg) for all classes of beef cattle across all seasons can then be calculated as:

$$E = \sum_i \sum_j \sum_k \sum_l (91.25 \times N_{ijkl} \times M_{ijkl}) \times 10^{-6} \dots\dots\dots (4A.1b_8)$$

Where: N_{ijkl} = numbers of beef cattle in each State, region, season and class
 M_{ijkl} = methane production (kg/head/day)
 91.25 = number of days in each season

Beef Cattle in Feedlots

Table 6.5: Symbols used in algorithms for feedlot cattle

State (i)	Feedlot Cattle Classes (duration of stay) (j)	Average Length of Stay(days)
i = 1 NSW/ACT	j = 1 Domestic/(70–80 days)	75
i = 2 Tasmania	j = 2 Export/(80–200 days)	140
i = 3 Western Australia	j = 3 Japan ox/(200+ days)	250
i = 4 South Australia		
i = 5 Victoria		
i = 6 Queensland		
i = 7 Northern Territory		

The equation developed by Moe and Tyrrell (1979) to predict daily methane yields (Y_{ij} MJ CH₄/head/day) is:

$$Y_{ij} = 3.406 + 0.510SR_{ij} + 1.736H_{ij} + 2.648C_{ij} \dots\dots\dots (4A.1c_1)$$

Where: SR_{ij} = intake of soluble residue (kg/day)
 H_{ij} = intake of hemicellulose (kg/day)
 C_{ij} = intake of cellulose (kg/day)

Each of SR_{ij} , H_{ij} and C_{ij} is calculated from the total intake of the animal (Appendix 6.C.2), the proportion of the diet of each class of animal that is grass, legume, grain (including molasses) and other concentrates (Appendix 6.C.3) and the soluble residue, hemicellulose and cellulose fractions of each of these components (Appendix 6.C.4).

The total daily production of methane (M_{ij} kg CH_4 /head/day) is thus:

$$M_{ij} = Y_{ij} / F \dots\dots\dots (4A.1c_2)$$

Where: $F = 55.22$ MJ/kg CH_4 (Brouwer 1965)

Methane production (Gg) for all classes of feedlot cattle across all States can then be calculated as:

$$E = \sum_i \sum_j (365 \times N_{ij} \times M_{ij}) \times 10^{-6} \dots\dots\dots (4A.1c_3)$$

Where: N_{ij} = numbers of feedlot cattle as an annual equivalent in each class in each State
 M_{ij} = methane production (kg/head/day)

6.3.2.2 Sheep (4.A.3)

Estimated emissions are based on Howden et al. (1994) who report a close relationship between dry matter intake and methane production based on an analysis of Australian respiration chamber experiments on sheep (inter alia Margan et al. 1985, 1987, 1988 and Graham 1964a,b, 1967, 1969). Howden et al. (1994) found that feed intake alone explained 87% of the variation in methane production.

The Agriculture and Food Research Council (AFRC) (1990) equation for intake is used here, as it corresponded well with intakes reported by State experts for seasonal digestibilities common in their State.

The methodology for estimating emissions from sheep has been independently verified. Leuning et al. (1999) found close agreement between the methane emissions estimated by the inventory methods and direct field measurements made using micrometeorological mass-balance and SF_6 tracer techniques. Using the inventory methods and default livestock characterisation Leuning et al. (1999) estimated CH_4 emissions to be 12.6 g/head/day compared with 11.9 (± 1.5) and 11.7 (± 0.4) g/head/day measured by the mass-balance and SF_6 tracer techniques respectively. When the actual experimental livestock characterisation was used with the inventory methods, CH_4 emissions were estimated to be 11.1 g/head/day.

In addition, an analysis of recent Australian respiration chamber experiments by Williams and Wright (2005) showed a very similar relationship between methane output and dry matter intake ($CH_4 = 0.0187 \times DMI - 0.0003$) to that reported in Howden et al. (1994) ($CH_4 = 0.0188 \times DMI + 0.00158$).

Table 6.6: Symbols used in algorithms for sheep

State (i)	Season (j)	Sheep Classes (k)
		<i>Sheep > 1 year</i>
i = 1 NSW/ACT	j = 1 Spring	k = 1 Rams
i = 2 Tasmania	j = 2 Summer	k = 2 Wethers
i = 3 Western Australia	j = 3 Autumn	k = 3 Maiden Ewes (intended for breeding)
i = 4 South Australia	j = 4 Winter	k = 4 Breeding Ewes
i = 5 Victoria		k = 5 Other Ewes
i = 6 Queensland		<i>Sheep < 1 year</i>
i = 7 Northern Territory		k = 6 Lambs and Hoggets

The potential, or maximum, intake of feed by sheep occurs when feed is abundant and of high quality. Potential intake is determined largely by body size and the proportion of the diet that is able to be metabolised by the animal. Potential intake (PI_{ijk} kg DM/head/day) is given by AFRC (1990) as:

$$PI_{ijk} = (104.7 q_{m,ijk} + 0.307 W_{ijk} - 15.0) W_{ijk}^{0.75} / 1000 \dots\dots\dots (4A.3_1)$$

Where: W_{ijk} = liveweight (kg) (Appendix 6.D.1)

$q_{m,ijk}$ = metabolizability of the diet. This is the ratio of metabolizable energy (ME) to gross energy (GE) in the diet (i.e. ME / GE). Metabolizable energy content is related to digestibility of dry matter (DMD_{ijk}) so, using the equation of Minson and McDonald (1987), $q_{m,ijk} = 0.00795 DMD - 0.0014$ (DMD is expressed as a %)

However, the actual feed intake of animals is often less than the potential intake. This can be caused by many factors, especially by low feed availability. Relative intake is defined as the proportion of potential intake that the animal will consume. The relative intake due to feed availability is given by White *et al.* (1983) as:

$$RI_{ijk} = 1 - \exp(-2(DMA_{ijk})^2) \dots\dots\dots (4A.3_2)$$

Where: DMA_{ijk} = dry matter availability tonnes/hectare (Appendix 6.D.3)

Note: Actual feed intake will be less than potential intake only when feed availability is less than 1.63 tonnes/hectare.

The actual intake (I_{ijk} kg DM/head/day) of a sheep is thus:

$$I_{ijk} = PI_{ijk} \times RI_{ijk} \times MA_{ijk=4} \dots\dots\dots (4A.3_3)$$

Where: $MA_{ijk=4}$ = additional intake for milk production

Feed intakes can increase by up to 60% during lactation (ARC 1980). For emissions estimates, the intake of all breeding ewes was assumed to increase by 30% during the season in which lambing occurs, based on relationships presented in SCA (1990).

The additional intake for milk production ($MA_{ijk=4}$) is calculated by:

$$MA_{ijk=4} = (LE_{ijk=4} \times FA_{ijk=4}) + ((1-LE_{ijk=4}) \times 1) \dots\dots\dots (4A.3_4)$$

Where: $LE_{ijk=4}$ = proportion of breeding ewes lactating, calculated as the annual lambing rates x proportion of lambs receiving milk in each season (Appendix 6.D.6)

$FA_{ijk=4}$ = feed adjustment (assumed to be 1.3)

Methane production (M_{ijk} kg/head/day) is calculated using daily intake figures (I_{ijk}) via the relationship of Howden et al. (1994):

$$M_{ijk} = I_{ijk} \times 0.0188 + 0.00158 \dots\dots\dots(4A.3_5)$$

The annual methane production (in Gg) of Australian sheep is calculated as:

$$E = \sum_i \sum_j \sum_k (91.25 \times N_{ijk} \times M_{ijk}) \times 10^{-6} \dots\dots\dots(4A.3_6)$$

Where: N_{ijk} = numbers of sheep in each class for each season and state
 M_{ijk} = methane production (kg/head/day)

6.3.2.3 Pigs (4.A.8)

Pigs are non-ruminant animals, and convert a smaller proportion of feed energy intake into methane than do ruminants. The OECD (1991) estimated that, depending on diet, pigs typically convert around 1% of gross energy intake into methane, compared with around 6–7% for cattle and sheep.

Whittemore (1993) suggests that the output of methane by a 60 kg pig is about 0.2 MJ/day. Assuming that, on average, a 60 kg pig consumes 1.95 kg DM/day of a diet containing 18.6 MJ GE/kg, the GE intake was 36.3 MJ GE. Thus pigs would convert around 0.6% of gross energy into methane.

Other values in the literature suggest methane conversions of 1.2% of gross energy (Christensen and Thorbek 1987), 0.6 to 0.8% of gross energy (Moss 1993) and 0.4% gross energy (Kirchgessner et al. 1991).

A methane conversion of 0.7% of gross energy intake is used for Australia.

Table 6.7: Symbols used in algorithms for pigs

State (i)	Pig Classes (j)
i = 1 NSW/ACT	j = 1 Boars
i = 2 Tasmania	j = 2 Sows
i = 3 Western Australia	j = 3 Gilts
i = 4 South Australia	j = 4 Others
i = 5 Victoria	
i = 6 Queensland	
i = 7 Northern Territory	

The relationship for enteric fermentation in pigs gives the total daily production of methane (M_{ij} kg CH₄/head/day) as:

$$M_{ij} = I_{ij} \times 18.6 \times 0.007 / F \dots\dots\dots(4A.8_1)$$

Where: I_{ij} = intake (kg DM/day) (Appendix 6.E.1)
 F = 55.22 MJ/kg CH₄ (Brouwer 1965)

18.6 = MJ GE/kg feed DM

The annual production of methane (Gg) for all classes of pigs is calculated as:

$$E = \sum_i \sum_j (N_{ij} \times M_{ij} \times 365) \times 10^{-6} \dots\dots\dots(4A.8_2)$$

Where: N_{ij} = the number of pigs in each class for each State
 M_{ij} = methane production (kg/head/day)

6.3.2.4 Other Livestock (Also Including 4A.2, 4, 5, 7, 9, and 10)

Livestock production in Australia is not confined to sheep, pigs and various types of cattle. A range of ‘other livestock’ are commercially farmed throughout all States, with a trend towards increasing diversity since 1990. Included in calculations of enteric methane emissions are both ruminants (goats, deer and buffalo), quasi-ruminants (camels and alpacas) and non-ruminants (horses, donkeys, mules, emus, ostriches and a range of poultry).

The contribution of other livestock to total methane production is comparatively small. Given the relative absence of data for enteric methane production for other livestock under Australian conditions, a complex methodology, incorporating relationships between intake and methane generated, is inappropriate. In its place, a simplified methodology, based on the IPCC (1997, vol. 3) *Tier I* method, using aggregated numbers of the various livestock types and an annual methane emissions factor is used.

The IPCC (1997, vol. 3) identifies average emissions factors for a number of minor livestock types, based on the work of Crutzen et al. (1986). Findings were presented for developed and developing countries. Values pertaining to developed countries were used for the Australian inventory.

- Goats—5 kg/head/year
- Buffalo—55 kg/head/year
- Camels—46 kg/head/year
- Horses—18 kg/head/year
- Donkeys and Mules—10 kg/head/year
- Poultry—not estimated by IPCC (1997)

Emission factors for deer, alpacas, emus and ostriches are not provided by IPCC (1997, 2000), so alternative sources were used. The emission factor for deer was adopted from Howden (1991). For alpacas, emus and ostriches, estimates were made based on animal size and anatomy, to be equivalent with emission factors for donkeys and goats respectively.

- Deer—10.7 kg/head/year
- Alpacas—10 kg/head/year
- Emus and Ostriches—5 kg/head/year

Table 6.8: Symbols used in algorithms for other livestock

State (i)	Other Livestock Types (j)	Digestive Type
i = 1 NSW/ACT	j = 1 Goats	ruminant
i = 2 Tasmania	j = 2 Horses	non-ruminant (equine)
i = 3 Western Australia	j = 3 Deer	ruminant
i = 4 South Australia	j = 4 Buffalo	ruminant
i = 5 Victoria	j = 5 Donkeys/Mules	non-ruminant (equine)
i = 6 Queensland	j = 6 Emus/Ostriches	non-ruminant
i = 7 Northern Territory	j = 7 Alpacas	quasi-ruminant
	j = 8 Camels	quasi-ruminant
	j = 9 Poultry	non-ruminant

The calculation of methane emissions from enteric fermentation in minor livestock types uses a simplified approach based on the default emissions factors listed previously.

By applying the emission factor to the number of each species in each State, total methane production (Gg) from the enteric fermentation of minor livestock types can be calculated as follows:

$$E = \sum_i \sum_j (N_{ij} \times M_{ij} \times 10^{-6}) \dots\dots\dots (4A.10_1)$$

Where: N_{ij} = numbers of 'other livestock' types in each State
 M_{ij} = methane emission factor (kg/head/year) (Appendix 6.F.1)

6.3.7 Uncertainties and Time Series Consistency

A quantitative assessment of uncertainty was undertaken and uncertainties for enteric fermentation were estimated to be in the order of 5%. Further details on the analysis are provided in Annex 7. Time series consistency is ensured by the use of consistent methods and full recalculations in the event of any refinement to methodology.

6.3.8 Source Specific QA/QC

6.3.8.1 Activity Data

The Australian Bureau of Statistics (ABS) is the national statistical agency of Australia and is the key provider of activity data for this source category. ABS has in place a range of quality assurance-quality control procedures associated with survey design, data input and consistency checks on the survey results and the aggregated values. Sampling errors are also evaluated.

Data quality used in the inventory is also kept under review by the DCCEE. This source category is also covered by the general QA/QC procedures detailed in Section 1.6.

As recommended by the IPCC GPG (2000), inverse modelling of cattle and sheep populations have been undertaken to ensure consistency with reported populations. These studies show no apparent bias in the sheep numbers (Howden 2001) but possible differences in cattle numbers in the order of 3-4% (Howden and Barrett 2003). It is important to note that, with the limited datasets available for this study, the parameter solutions were non-unique and it is possible that there were no systemic differences in the numbers. Given the size of the possible differences and the inherent uncertainty in animal numbers it was agreed with ABS to incorporate this information into the uncertainty estimates rather than adjust activity data.

6.3.8.2 Implied Emission Factors

As country specific tier 2 methods are used to estimate emissions from cattle, sheep and pigs, the implied emission factors (IEFs) have been compared with the IPCC defaults (Table 6.9).

The dairy cattle IEF is significantly higher than the IPCC default primarily because the default is based on milk production levels of 1,700 kg/head/yr when Australian milk production generally ranges from about 4000-5500 kg/head/year. Australia's IEF is more consistent with the default values for the regions with milk production at these levels (100-118 kg/head/year). The non-dairy cattle IEFs differ from the defaults as the methane conversion rates differ to the IPCC defaults (see section 6.3.8.4).

The lower IEFs for sheep and pigs primarily reflect the inclusion of an age structure in the Australian method and a difference in methane conversion rates for sheep (see section 6.3.8.4).

Table 6.9: Implied emission factors – enteric fermentation (kg/head/year)

Livestock Type	Australia	IPCC Default (Oceania)
Dairy Cattle	113	68
Non-dairy cattle		
Range-kept	72	53
Feedlot	77	53
Sheep	6.8	8
Swine	1.45	1.5

Trends in IEFs

As part of the standard Tier 1 QC activities, IEFs are checked for possible anomalies in the time-series. As Australia's tier 2 methods disaggregates animals by age class and regions (in which feed and animal characteristics differ) some inter-annual variability can be expected in the IEFs for cattle, sheep and pigs as the relative numbers in each age class and region change year to year. Only those inter-annual changes outside the expected range are described below:

- The IEF for feedlot cattle shows a period of increase (from 1990 to 1995), then an abrupt decrease (between 1995 and 1996), after which there is little change to the IEFs. There are two factors influencing the IEF. The most significant is that there is a step change in the input (average intakes, feed components and liveweight gain) in 1996 (see Appendix 6.C). There is an increase in the amount of concentrates fed post 1996 which leads to lower IEFs. The reason for the different trends between the two periods is the rapid shift from animals from domestic consumption towards the Japan Ox market in the early 1990s while from the mid 1990s there are only minor shifts between markets ($\pm 10\%$).

6.3.8.3 Feed Intake

As Australia uses country specific Tier 2 methods for estimating feed intake, intakes have been checked for reasonableness and have been compared with average intakes reported by other Parties,

Cattle

For dairy cattle average herd intakes are within the range reported by other Parties. The mean is lower (Table 6.10) because the herd average for Australia includes all animals in the dairy cattle herd not just milking cows and this would drive herd averages down compared with other Parties which have only reported milking cows.

The intakes of Australian dairy cattle are in the order of 1-3% of live weight (range from 1.4% to 3.1%) as recommended by the IPCC (2000).

Comparison of the non-dairy cattle between Parties is complicated by the fact that animals kept under feedlot conditions are not reported separately from range-kept animals as is done in the Australian inventory. The average herd intakes reported in Australia are within the range reported by other Parties although the means are significantly different with range-kept animals lower and feedlot higher than the mean of other Parties (Table 6.10).

Table 6.10: Average herd intake (MJ/head/day)

Livestock Type	Australia		Other Parties	
	Range	Mean	Range	Mean
Dairy Cattle	214-235	228	164-339	258
Non-dairy cattle			95-205	143
Range-kept	129-130	129		
Feedlot	190-200	198		
Sheep	17	17	19-46	25

There is a high level of confidence in the estimates of intake for feedlot cattle as these have been based on standard industry practices and range from 1.9 to 2.7% of live weights.

Intakes for other cattle are calculated based on the relationship of Minson and MacDonald (1987).

Howden et al. (2002) tested the Australian approach and the IPCC Tier 2 approach against animal growth and feed intake data collated from four research papers, and while neither method showed itself superior across all data sets, the Minson and McDonald approach provided better overall performance.

Howden et al. (2002) also compared the two approaches against feed intakes drawn from the Australian Feeding Standards (SCA 1990) for different animal sizes, feed digestibility and pregnancy status in Queensland. They found that the IPCC approach consistently over-predicted intake (mean over prediction of 43%). In contrast, the Australian method provided dry matter intake values that were not significantly different from those reported in the Feeding Standards.

On the basis of the above investigations, Howden et al. (2002) concluded that the Australian approach to estimating intakes was more robust than the IPCC approach for livestock in Australia.

Sheep

The country specific method used to estimate intake from Australian sheep produces lower average intakes than that reported by other Parties (Table 6.10). However, an analysis of intake as percentage of liveweight shows that intakes are in the order of 1-3% (range from 1.0% to 2.7%) as recommended by IPCC (2000).

The IPCC Tier 2 method for estimating intake predicts DMI based on how much feed must be consumed to meet estimated requirements and does not consider the biological capacity of the animal to, in fact, consume the predicted quantity of feed. For example, in Australia the actual feed intake is often less than the estimated intake due to low feed availability. The Australian method calculates the proportion of the potential intake that the animal will actually consume (potential intake is restricted when feed availability is less than 1.63 tonnes/hectare). Restricted feed conditions occur in one or more seasons in all States with animals experiencing weight loss over the season. When intakes are not limited, estimated intakes (average 21 MJ/day) are similar to levels reported by other Parties.

6.3.8.4 Methane Conversion Rates

As Australia uses country specific methods for estimating methane emissions, methane conversion rates (Y_m) have been compared against IPCC defaults.

Dairy and Beef Cattle

The conversion rate for dairy cows (6.9-7.2%) is higher than the IPCC *Good Practice Guidance* (2000) default of 6% but is within the range of the 2006 IPCC Guidelines default (6.5% \pm 1%). For other dairy cattle the methane conversion rates are slightly higher with overall herd average rates of about 7.3%.

The average conversion rate for temperate beef cattle (7.4%) is higher than the IPCC *Good Practice Guidance* (2000) default of 6% but is within the range of the 2006 IPCC Guidelines default (6.5% \pm 1%) although at the upper end. The IPCC (2000) indicate that where poorer feed is available the higher bound of emissions are appropriate, and in Australia cattle in temperate regions frequently experience diets with less than 60% digestibility. The IPCC (2000) acknowledge that animals on tropical feeds may have higher methane conversion rates, however the Australian method predicts considerably higher methane yields (10.9%) on tropical feeds than the IPCC defaults (7%). The IPCC (2000, 2006) acknowledges that Y_m values outside the default ranges may occur on tropical pastures and that further research is needed in this area.

Feedlot Cattle

The IPCC (2000) indicate that animals fed diets containing 90 percent concentrates should use Y_m values ranging from 3.5-4.5%. The Australian methodology for estimating emissions produces values of about 6%. Australian feedlot cattle diets consist of approximately 70-80% concentrates (predominately grain with small quantities of molasses and other concentrates) rather than the 90% indicated for the default value. The Australian method was tested using intake and feed quality data from Kurihara et al. (1999, 2006) who measured methane conversion rates of 5.6% using calorimetry chambers for cattle on a high grain (75%) plus lucerne diets compared with the 5.9% estimated by the Australian method.

Sheep

The herd average Ym for Australian sheep is 6.2%. This is lower than the IPCC *Good Practice Guidance* (2000) default of 7% but is within the range of the *2006 IPCC Guidelines* default (6.5% ±1%).

As described in section 6.3.2.2, independent verification has been undertaken on the emission estimates for sheep.

6.3.8.5 External Review

A comprehensive expert peer review of the methodologies, activity data and livestock characterisation data was conducted in 2000-01 involving agricultural experts from industry, government and academia.

6.3.9 Recalculations Since The 2008 Inventory

Emissions for the most recent year in the *agriculture* sector are recalculated each year due to the recalculation of the three-year average of emissions once the third year of data becomes available. In addition there have been updates to preliminary milk production estimates for 2008 and replacement of extrapolated animal numbers for some “other livestock” with interpolated estimates. The allocation of feedlot cattle into different export classes was revised for 2007-2008 based on information reported by the industry.

The net effect of these changes was an increase of 576 Gg CO₂-e in the 2008 estimate and no change in the 1990 estimate.

Table 6.11: 4.A Enteric fermentation: recalculation of total CO₂-e emissions, 1990-2008

Year	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1990	63,919	63,919	0	0.0
2000	60,442	60,442	0	0.0
2001	60,849	60,849	0	0.0
2002	59,523	59,523	0	0.0
2003	59,555	59,555	0	0.0
2004	59,834	59,834	0	0.0
2005	60,124	60,124	0	0.0
2006	59,082	59,082	-0	0.0
2007	57,185	57,201	17	0.0
2008	55,552	56,128	576	1.0

6.3.10 Source Specific Planned Improvements

The Australian Government is funding research into two key areas of methane emission generation from enteric fermentation that aim to provide data on emission levels under a variety of conditions as well as the estimation of the impacts of changes in management practices on emissions.

The Reducing Emissions from Livestock Research Program is managed by the Department of Agriculture, Forestry and Fisheries and will deliver a range of outputs by 2012.

There are a number of measurement techniques being analysed and compared with each other to estimate emissions from livestock: these include open path lasers and FTIR systems, respiration chambers, SF₆ and a new enteric methane infra red wireless sensor. These projects will generate data on emissions from livestock under a range of climate and management conditions. In particular, currently, the tropical cattle emission factor is based on an analysis of a single pasture type. CSIRO is conducting research into the effects of different tropical grass pasture diets on methane emissions from cattle. This project will also consider the effects of southern Australian temperate pastures on methane emissions from cattle.

A number of additional projects are investigating the potential for genetics and rumen modification (protozoa, microbial population manipulation, alternative hydrogen acceptors, feed quality, and feed additives) to deliver management strategies for the farming community to reduce methane emissions in the agriculture sector.

Information obtained from the research program will be used to review the key parameters currently adopted for emissions estimation, to consider the development of a more disaggregated specification of emission development to better reflect regional variations in conditions, where appropriate, and to develop models and monitoring mechanisms to reflect mitigation actions taken in the industry in the national inventory.

6.4 SOURCE CATEGORY 4.B MANURE MANAGEMENT

6.4.1 Source Category Description

Methane is produced from the decomposition of the organic matter remaining in the manure under anaerobic conditions. These conditions occur when large numbers of animals are managed in a confined area where manure is typically stored in large piles or lagoons.

Nitrous oxide emissions from manure management systems can occur via combined nitrification-denitrification of ammoniacal nitrogen contained in the wastes. The amount released depends on the system and duration of waste management.

6.4.2 Methodology

Methane

Methane production from the manure of dairy cattle, feedlot cattle, pigs and poultry are calculated based on the approach of the IPCC (1997) using a combination of default IPCC and country specific input values. Subscripts for the algorithms are the same as used for calculating enteric fermentation.

Australian experts considered that methane production was likely to be negligible in the manure of range-kept livestock (e.g. free-range beef cattle, sheep, goats etc). There was agreement that the generally high temperatures, high solar radiation and low humidity environments of Australia would dry manure rapidly. In combination with scarab (or dung) beetles that rapidly infest manure in most Australian environments, there was considered to be little likelihood of anaerobic conditions and hence little methanogenesis in the manure of range-kept animals. This is supported by González-Avalos and Ruiz-Suárez (2001) who recorded negligible amount of methane from the manure of cattle kept under conditions similar to those experienced in Australia. The González-Avalos and Ruiz-Suárez (2001) methodology is, therefore, used to estimate methane emissions from range-kept livestock manure.

Gonzalez-Avalos and Ruiz-Suarez (2001) calculate their methane emissions factor (M_{ijk} kg/head/year) as follows:

$$M_{ijk} = (\text{kg CH}_4 / \text{kg DM fresh manure}) \times (\text{kg fresh manure/day}) \times (\% \text{ DM}) \times 365$$

The second and third term on the right hand side of the equation quantify the amount of dry matter (DM) in manure. This can be calculated for beef cattle and sheep using estimates of Intake and DMD to estimate DM manure (kg/head/day):

$$\text{DMM}_{ijk} = I_{ijk} \times (1 - \text{DMD}_{ijk})$$

Gonzalez-Avalos and Ruiz-Suarez (2001) do not document emission factors in kg CH₄/kg DM fresh manure. However, this can be calculated from the data in their paper using their Extensive Dual Purpose (temperate and warm) categories. The manure emission factors (MEF) are:

Temperate (21°C) — 1.4×10^{-5} kg CH₄/kg DM manure

Warm — 5.4×10^{-5} kg CH₄/kg DM manure

Nitrous Oxide

The methodology for estimating nitrogen excretion is based on equations from SCA (1990) and Freer et al. (1997) which have been developed in Australia, rather than the default IPCC (1997) values. This methodology takes a mass balance approach where N output = N input – N storage. The total N output is then split into urinary and faecal components.

The methodology for estimating emissions is based on the IPCC (1997, vol 3) guidelines incorporating manure management systems that reflect Australian conditions. Subscripts for algorithms are the same used for calculating methane emissions with an additional manure management system subscript (Table 6.12). The nitrogen voided in the waste of domestic livestock estimated in this section provides the basis of calculating nitrous oxide emissions from soils in source category 4D.

In the UNFCCC common reporting format tables nitrous oxide emissions are reported for the different manure management systems rather than on the basis of livestock type. The management system to which the livestock types contribute are shown in Table 6.13. The ‘other’ category includes pig manure treated in a digester and poultry manure with and without bedding.

Table 6.12: Symbols and emissions factors used in algorithms for nitrous oxide emissions

Manure Management Systems (MMS)	Emission Factor ^(a) (kg N ₂ O-N/kg N excreted)
MMS = 1 Anaerobic lagoon	0.001
MMS = 2 Liquid systems	0.001
MMS = 3 Daily spread	0 ^(b)
MMS = 4 Solid storage and drylot	0.02
MMS = 5 Digester	0.001
MMS = 6 Poultry manure with bedding	0.02
MMS = 7 Poultry manure without bedding	0.005
MMS = 8 Pasture range and paddock	0 ^(b)

(a) IPCC (1997, 2000). (b) There are no direct emissions from these sources. Emissions from the soil are reported under Agricultural Soils (4D).

Table 6.13: Manure management system to which livestock contribute

Greenhouse gas source and sink categories	Anaerobic lagoon	Liquid system	Daily spread	Solid storage and dry lot	Pasture range and paddock	Other
Non-Dairy Cattle – Free Range	NO	NO	NO	NO	Source	NO
Dairy Cattle	Source	Source	Source	NO	Source	NO
Sheep	NO	NO	NO	NO	Source	NO
Swine	Source	Source	Source	Source	NO	Source
Poultry	NO	NO	NO	NO	Source	Source
Non-Dairy Cattle – Feedlot	NO	NO	NO	Source	NO	NO
Goats, horses, deer, buffalo, donkeys, mules, emus, ostriches, alpacas, camels	NO	NO	NO	NO	Source	NO

Table based on CRF Table 4.B(b)

6.4.2.1 Cattle (4.B.1)

Dairy Cattle

Methane

Dairy cattle are generally kept in higher rainfall areas than other Australian livestock. This, and the disposal of excreta washed from milking sheds, gives opportunities for the generation of methane. However, only a small fraction of the potential methane emissions appears to be released. Williams (1993) measured methane production from dairy cattle manure under field conditions in Australia and found that only about 1% of the methane production potential was achieved. On this basis, the temperate methane conversion factor values for manure voided in the field were reduced to 1% from the IPCC (1997, vol. 3) default value of 1.5%.

Methane from manure is formed from the organic fraction of the manure (volatile solids). Volatile solid production for dairy cattle (VS_{ijk} kg/head/day) was estimated using the data developed to calculate enteric methane production as this included information on intakes and dry matter digestibility. For dairy cattle, volatile solids were calculated:

$$VS_{ijk} = I_{ijk} \times (1 - DMD_{ijk}) \times (1 - A) \dots\dots\dots(4B.1a_1)$$

Where: I_{ijk} = dry matter intake calculated in Section 6.3.2.1

DMD_{ijk} = dry matter digestibility expressed as a fraction (Appendix 6.A.3)

A = ash content expressed as a fraction (assumed to be 8% of faecal DM)

Methane production from manure (M_{ijk} kg/head/day) is then calculated as:

$$M_{ijk} = VS_{ijk} \times B_o \times MCF \times \rho \dots\dots\dots(4B.1a_2)$$

Where: B_o = emissions potential – $0.24m^3 CH_4/kg VS$ (IPCC 1997)

MCF = integrated methane conversion factor—based on the proportion of different manure management regimes (Appendix 6.A.4) and MCF values for ‘warm’ regions for Qld and the NT and MCF values for ‘temperate’ regions for all other States (Appendix 6.A.5)

ρ = density of methane ($0.662 kg/m^3$)

The annual methane production (Gg) from the manure of dairy cattle is calculated as:

$$Total = \sum_{i=1,3-7} \sum_j (365 \times N_{ij} \times M_{ij}) \times 10^{-6} + \sum_{i=2} \sum_j \sum_k (91.25 \times N_{ijk} \times M_{ijk}) \times 10^{-6} \dots\dots\dots(4B.1a_3)$$

Where: N_{ijk} = numbers of dairy cattle in each State, class and season

M_{ijk} = methane production (kg/head/day)

Nitrous Oxide

The methodology for calculating the excretion of nitrogen from dairy cattle makes use of the following algorithms to calculate crude protein input (CPI_{ijk}) and storage (NR_{ijk}) and from these the output of nitrogen in the faeces and urine.

The crude protein intake CPI_{ijk} (kg/head/day) of dairy cattle is calculated thus:

$$CPI_{ijk} = I_{ijk} \times CP_{ijk} \dots\dots\dots(4B.1a_4)$$

Where: I_{ijk} = dry matter intake (kg/day) as calculated in Section 6.3.2.1

CP_{ijk} = crude protein content of feed intake expressed as a fraction (Appendix 6.A.6)

Dairy calves are usually removed from their mothers very early and are either sold for slaughter or raised as replacements. The calves receive a supplement (milk replacer and feed pellets) and are on average weaned after six weeks. The protein from the supplements is not included in the equation which assumes that intake pre-weaning is from pasture.

Nitrogen excreted in faeces (F_{ijk} kg/head/day) is calculated, using functions developed by the SCA (1990) and Freer et al. (1997), as the indigestible fraction of the undegraded protein from solid feed and the microbial crude protein plus the endogenous faecal protein, such that:

$$F_{ijk} = \{0.3(CPI_{ijk} \times (1 - [(DMD_{ijk} + 10)/100])) + 0.105(ME_{ijk} \times I_{ijk} \times 0.008) + (0.0152 \times I_{ijk})\}/6.25 \dots (4B.1a_5)$$

Where: DMD_{ijk} = dry matter digestibility expressed as a % (Appendix 6.A.3)

ME_{ijk} = metabolizable energy (MJ/kg DM) calculated as: $0.1604 DMD_{ijk} - 1.037$ (Minson and McDonald 1987)

I_{ijk} = dry matter intake (kg/day)

1/6.25 = factor for converting crude protein into nitrogen

The amount of nitrogen that is retained by the body (NR_{ijk} kg/head/day) is calculated as the amount of nitrogen retained in milk and body tissue such that:

$$NR_{ijk} = \{(0.032 \times MP_{ijk}) + \{0.212 - 0.008(L_{ijk} - 2) - [(0.140 - 0.008(L_{ijk} - 2)) / (1 + \exp(-6(Z_{ijk} - 0.4)))]\} \times (LWG_{ijk} \times 0.92)\}/6.25 \dots (4B.1a_6)$$

Where: MP_{ijk} = milk production in kg/head/day

L_{ijk} = relative intake as calculated in Section 6.32.1

Z_{ijk} = relative size (liveweight / standard reference weight (Appendix 6.A.7))

LWG_{ijk} = liveweight gain (kg/day) (Appendix 6.A.2)

Nitrogen excreted in urine (U_{ijk} kg/head/day) is calculated by subtracting NR_{ijk} , F_{ijk} and dermal protein loss from the nitrogen intake such that:

$$U_{ijk} = (CPI_{ijk}/6.25) - NR_{ijk} - F_{ijk} - [(1.1 \times 10^{-4} \times W_{ijk}^{0.75})/6.25] \dots (4B.1a_7)$$

Where: W_{ijk} = liveweight (Appendix 6.A.1)

The total annual faecal (AF_{jk} Gg) and urinary (AU_{jk} Gg) nitrogen excreted is calculated as:

$$AF_{ijk} = \sum_{i=1,3-7} \sum_j (365 \times N_{ij} \times F_{ij}) \times 10^{-6} + \sum_{i=2} \sum_j \sum_k (91.25 \times N_{ijk} \times F_{ijk}) \times 10^{-6} \dots (4B.1a_8a)$$

$$AU_{ijk} = \sum_{i=1,3-7} \sum_j (365 \times N_{ij} \times U_{ij}) \times 10^{-6} + \sum_{i=2} \sum_j \sum_k (91.25 \times N_{ijk} \times U_{ijk}) \times 10^{-6} \dots (4B.1a_8b)$$

Where: N_{ijk} = the number of dairy cows in each State, season and class

The total emissions of nitrous oxide from the different manure management systems can then be calculated as follows:

$$Faecal_{MMS} = (AF_{ijk} \times MMS \times EF_{(MMS)} \times C_g) \dots (4B.1a_9a)$$

$$Urine_{MMS} = (AU_{ijk} \times MMS \times EF_{(MMS)} \times C_g) \dots (4B.1a_9b)$$

$$Total_{MMS} = \sum_i \sum_j \sum_k (Faecal_{MMS} + Urine_{MMS}) \dots (4B.1a_9c)$$

Where: MMS = the fraction of nitrogen that is managed in the different manure management systems (Appendix 6.A.4).

$EF_{(MMS)}$ = emission factor (N_2O -N kg/ N excreted) for the different manure management systems (Table 6.11)

C_g = 44/28 factor to convert elemental mass of N_2O to molecular mass

Beef Cattle

Methane

Methane production from the manure (M_{ijkl} kg/head/day) of free-range beef cattle is calculated as:

$$M_{ijkl} = I_{ijkl} \times (1 - \text{DMD}_{ijkl}) \times \text{MEF} \dots\dots\dots(4B.1b_1)$$

Where: I_{ijkl} = dry matter intake calculated in Section 6.3.2.1

MEF = emission factor (kg CH_4 / kg DM Manure) (Gonzalez-Avalos and Ruiz-Suarez 2001). The warm factor is used for Qld, NT and the Kimberly region of WA and the temperate factor is used for all other States and regions.

The annual methane production (Gg) from the manure of free-range beef cattle is calculated as:

$$\text{Total} = \sum_i \sum_j \sum_k \sum_l (N_{ijkl} \times M_{ijkl} \times 91.25) \times 10^{-6} \dots\dots\dots(4B.1b_2)$$

Where: N_{ijkl} = numbers of beef cattle in each State, class and season

M_{ijkl} = methane production (kg/head/day)

Nitrous Oxide

The methodology for calculating the excretion of nitrogen from beef cattle makes use of the following algorithms to calculate crude protein input (CPI_{ijkl}) and storage (NR_{ijkl}) and from these the output of nitrogen in the faeces and urine.

The crude protein intake CPI_{ijkl} (kg/head/day) of beef cattle is calculated thus:

$$\text{CPI}_{ijkl} = I_{ijkl} \times \text{CP}_{ijkl} + (0.032 \times \text{MC}_{ijkl}) \dots\dots\dots(4B.1b_3)$$

Where: I_{ijkl} = dry matter intake (kg/head/day) as calculated in Section 6.3.2.1

CP_{ijkl} = crude protein content of feed dry matter expressed as a fraction (Appendix 6.B.4)

MC_{ijkl} = milk intake (kg/day) – for areas where Brahman cross breeds are dominant (NT, Qld and Kimberly WA) considered to be 4 kg/day for animals in the first season after birth and 3 kg/day in the second season. In other areas where Hereford or Shorthorn breeds are dominant (all other States) considered to be 6 and 4 kg/day (Appendix 6.B.5)

Nitrogen excreted in faeces (F_{ijkl} kg/head/day) is calculated, using equations developed by the SCA (1990) and Freer et al. (1997), as the indigestible fraction of the undegraded protein from solid feed, microbial crude protein and milk protein plus the endogenous faecal protein, such that:

$$F_{ijkl} = \{0.3(\text{CPI}_{ijkl} \times (1 - [(\text{DMD}_{ijkl} + 10)/100])) + 0.105(\text{ME}_{ijkl} \times I_{ijkl} \times 0.008) + 0.08(0.032 \times \text{MC}_{ijkl}) + (0.0152 \times I_{ijkl})\} / 6.25 \dots\dots\dots(4B.1b_4)$$

Where: DMD_{ijkl} = dry matter digestibility (expressed as a %) (Appendix 6.B.3)

ME_{ijkl} = metabolizable energy (MJ/kg DM) calculated by Minson and McDonald (1987) as: $\text{ME} = 0.1604 \text{DMD}_{ijkl} - 1.037$; (DMD expressed as %)

I_{ijkl} = feed intake (kg DM/head/day)

MC_{ijkl} = milk intake (kg/head/day). In areas where Brahman cross breeds are dominant (NT, Qld and Kimberly WA) milk intake is 4 kg/day for animals in the first season after birth and 3 kg/day in the second season. In other areas where Hereford or Shorthorn breeds are dominant (all other States) intake is 6 and 4 kg/day (Appendix 6.B.5)

1/6.25 = factor for converting crude protein into nitrogen

The amount of nitrogen that is retained by the body (NR_{ijkl} kg/head/day) is calculated as the amount of nitrogen retained as milk and body tissue such that:

$$NR_{ijkl} = \{ (0.032 \times MP_{ijkl}) + \{ 0.212 - 0.008(L_{ijkl} - 2) - [(0.140 - 0.008(L_{ijkl} - 2)) / (1 + \exp(-6(Z_{ijkl} - 0.4)))] \} \times (LWG_{ijkl} \times 0.92) \} / 6.25 \dots\dots\dots (4B.1b_5)$$

Where: MP_{ijkl} = milk production (kg/head/day) calculated as: proportion of cows lactating (LC_{ijkl}) x milk production. In areas where Brahman cross breeds are dominant (NT, Qld and Kimberly WA) milk production is 4 kg/day for cows >2 years old in the first season after calving and 3 kg/day in the second season. In other areas where Hereford or Shorthorn breeds are dominant (all other States) considered to be 6 and 4 kg/day (Appendix 6.B.5)

L_{ijkl} = relative intake as calculated in Section 6.3.2.1

Z_{ijkl} = relative size (liveweight / standard reference weight (Appendix 6.B.6))

LWG_{ijkl} = liveweight gain (kg/day) (Appendix 6.B.2)

Nitrogen excreted in urine (U_{ijkl} kg/head/day) is calculated by subtracting NR_{ijkl} , F_{ijkl} and dermal protein loss from the nitrogen intake such that:

$$U_{ijkl} = (CPI_{ijkl} / 6.25) - NR_{ijkl} - F_{ijkl} - [(1.1 \times 10^{-4} \times W_{ijkl}^{0.75}) / 6.25] \dots\dots\dots (4B.1b_6)$$

Where: W_{ijkl} = liveweight (Appendix 6.B.1)

The total annual faecal (AF_{ijkl} Gg) and urinary (AU_{ijkl} Gg) nitrogen excreted is calculated as:

$$AF_{ijkl} = (N_{ijkl} \times F_{ijkl} \times 91.25) \times 10^{-6} \dots\dots\dots (4B.1b_7a)$$

$$AU_{ijkl} = (N_{ijkl} \times U_{ijkl} \times 91.25) \times 10^{-6} \dots\dots\dots (4B.1b_7b)$$

Where: N_{ijkl} = the number of beef cattle adjusted for feedlot cattle in each State, region, season and class

The total emissions of nitrous oxide from the different manure management systems can then be calculated as follows:

$$\text{Faecal}_{MMS} = (AF_{ijkl} \times MMS \times EF_{(MMS)} \times C_g) \dots\dots\dots (4B.1b_8a)$$

$$\text{Urine}_{MMS} = (AU_{ijkl} \times MMS \times EF_{(MMS)} \times C_g) \dots\dots\dots (4B.1b_8b)$$

$$\text{Total}_{MMS} = \sum_i \sum_j \sum_k \sum_l (\text{Faecal}_{MMS} + \text{Urine}_{MMS}) \dots\dots\dots (4B.1b_8c)$$

Where: MMS = the fraction of nitrogen that is managed in the different manure management systems. In the case of beef cattle all manure is voided at pasture ($MMS = 8$).

$EF_{(MMS)}$ = emissions factor (N_2O -N kg/ N excreted) for the different manure management systems (Table 6.11)

C_g = 44/28 factor to convert elemental mass of N_2O to molecular mass

Beef Cattle In Feedlots

Methane

The high density of animals in feedlots results in high concentrations of manure from which methane can be produced when the dung pack becomes moistened and anaerobic microsites occur. This, and the disposal of manure washed out from the facility, gives opportunities for the generation of methane. However, the dry packing arrangement used in most feedlots in Australia is likely to result in only a small fraction of the potential methane emissions being generated (IPCC 1997). The IPCC (1997, vol. 3) default values for drylot Methane Conversion Factors are 1.5% and 2% for temperate and warm regions, respectively.

Methane from manure is formed from the organic fraction of the manure (volatile solids). Volatile solid production for beef cattle in feedlots (VS_{ij} kg/head/day) was estimated using the data developed to calculate enteric methane production as this included information on intakes and dry matter digestibility:

$$VS_{ij} = I_{ij} \times (1 - DMD_{ij}) \times (1 - A) \dots\dots\dots (4B.1c_1)$$

Where: I_{ij} = dry matter intake (Appendix 6.C.2)
 DMD_{ij} = digestibility expressed as a fraction (assumed to be 80%)
 A = ash content expressed as a fraction (assumed to be 8% of faecal DM)

Methane production from the manure (M_{ij} kg/head/day) is then calculated as:

$$M_{ij} = VS_{ij} \times B_o \times MCF \times \rho \dots\dots\dots (4B.1c_2)$$

Where: B_o = emissions potential ($0.17m^3 CH_4/kg VS$ (IPCC 1997))
 MCF = methane conversion factor. IPCC drylot MCF value for 'warm' regions used for Queensland and the Northern Territory (5%) and MCF value for 'temperate' regions used for all other States (1.5%).
 ρ = density of methane ($0.662 kg/m^3$)

The annual methane production (Gg) from the manure of beef cattle in feedlots is calculated as:

$$E = \sum_{ij} (365 \times N_{ij} \times M_{ij} \times 10^{-6}) \dots\dots\dots (4B.1c_3)$$

Where: N_{ij} = Annual equivalent numbers of beef cattle in feedlots
 M_{ij} = methane production (kg/head/day)

Nitrous Oxide

The methodology for calculating the excretion of nitrogen from feedlot cattle makes use of the following algorithms to calculate crude protein input (CPI_{ij}) and storage (NR_{ij}) and from these the output of nitrogen in the faeces and urine.

The crude protein intake CPI_{ij} (kg/head/day) of feedlot cattle is calculated by:

$$CPI_{ij} = NI_{ij} \times 6.25 \dots\dots\dots (4B.1c_4)$$

Where: NI_{ij} = nitrogen intake (kg/day)
 6.25 = factor for converting nitrogen into crude protein

NI_{ij} is calculated from the total intake (Appendix 6.C.2) of the animals as the proportion of the diet that is grass, legume, grain (including molasses) and other concentrates (Appendix 6.C.3) and the nitrogen fraction of each of these components (Appendix 6.C.4).

Nitrogen excreted in faeces (F_{ij} kg/head/day) is calculated, using functions developed by the SCA (1990) and Freer et al. (1997), as the indigestible fraction of the undegraded protein from solid feed and the microbial crude protein plus the endogenous faecal protein, such that:

$$F_{ij} = \{0.3(CPI_{ij} \times (1 - [(DMD_{ij} + 10)/100])) + 0.105(ME_{ij} \times I_{ij} \times 0.008) + 0.0152 \times I_{ij}\} / 6.25 \dots\dots\dots (4B.1c_5)$$

Where: DMD_{ij} = digestibility expressed as a percentage (assumed to be 80%)
 ME_{ij} = metabolizable energy (MJ/kg DM) calculated as: $0.1604 DMD_{ij} - 1.037$ (Minson and McDonald 1987)
 I_{ij} = feed intake (kg/day) (Appendix 6.C.2)

The amount of nitrogen that is retained by the body (NR_{ij} kg/head/day) is calculated as the amount of nitrogen retained as body tissue such that:

$$NR_{ij} = \{[0.212 - 0.008(L_{ij} - 2) - \{(0.140 - 0.008(L_{ij} - 2)) / (1 + \exp(-6(Z_{ij} - 0.4)))\}]\} \times (LWG_{ij} \times 0.92) / 6.25 \quad (4B.1c_6)$$

Where: L_{ij} = relative intake. Determined by dividing feed intake by the intake require for maintenance (as per Minson and McDonald 1987)

Z_{ij} = relative size (liveweight (Appendix 6.C.1) / standard reference weight Steers>1 (Appendix 6.B.6))

LWG_{ij} = liveweight gain (Appendix 6.C.5)

Nitrogen excreted in urine (U_{ij} kg/head/day) is calculated by subtracting NR_{ij} , F_{ij} and dermal protein loss from the nitrogen intake such that:

$$U_{ij} = (CPI_{ij} / 6.25) - NR_{ij} - F_{ij} - [(1.1 \times 10^{-4} \times W_{ij}^{0.75}) / 6.25] \quad (4B.1c_7)$$

Where: W_{ij} = liveweight (Appendix 6.C.1)

The total annual faecal (AF_{ij} Gg) and urinary (AU_{ij} Gg) nitrogen excreted is calculated as:

$$AF_{ij} = (N_{ij} \times F_{ij} \times 365) \times 10^{-6} \quad (4B.1c_8a)$$

$$AU_{ij} = (N_{ij} \times U_{ij} \times 365) \times 10^{-6} \quad (4B.1c_8b)$$

Where: N_{ij} = the annual equivalent number of feedlot cattle in each State and class

The total emissions of nitrous oxide from the different manure management systems can then be calculated as follows:

$$Faecal_{MMS} = (AF_{ij} \times MMS \times EF_{(MMS)} \times Cg) \quad (4B.1c_9a)$$

$$Urine_{MMS} = (AU_{ij} \times MMS \times EF_{(MMS)} \times Cg) \quad (4B.1c_9b)$$

$$E_{MMS} = \sum_i \sum_j (Faecal_{MMS} + Urine_{MMS}) \quad (4B.1c_9c)$$

Where: MMS = the fraction of nitrogen that is managed in the different manure management systems. In the case of feedlot cattle all manure is dry packed (MMS = 4).

$EF_{(MMS)}$ = emission factor (N_2O -N kg/ N excreted) for the different manure management systems (Table 6.12)

$Cg = 44/28$ factor to convert elemental mass of N_2O to molecular mass

6.4.2.2 Sheep (4.B.3)

Methane

Methane production from the manure (M_{ijk} kg/head/day) of sheep is calculated as:

$$M_{ijk} = I_{ijk} \times (1 - DMD_{ijk}) \times MEF \quad (4B.3_1)$$

Where: I_{ijk} = dry matter intake calculated in Section 6.3.2.2

MEF = manure emission factor ($kg CH_4 / kg DM Manure$) (Gonzalez-Avalos and Ruiz-Suarez 2001). The warm factor is used for Qld and NT and the temperate factor is used for all other States.

The annual methane production (Gg) from the manure of sheep is calculated as:

$$\text{Total} = \sum_i \sum_j \sum_k (N_{ijk} \times M_{ijkl} \times 91.25) \times 10^{-6} \dots\dots\dots (4B.3_2)$$

Where: N_{ijk} = numbers of sheep in each State, class and season
 M_{ijkl} = methane production (kg/head/day)

Nitrous Oxide

The methodology for calculating the excretion of nitrogen from sheep makes use of the following algorithms to calculate crude protein input (CPI_{ijk}) and storage (NR_{ijk}) and from these the output of nitrogen in the faeces and urine.

The crude protein intake CPI_{ijk} (kg/head/day) of sheep is calculated thus:

$$CPI_{ijk} = I_{ijk} \times CP_{ijk} + (0.045 \times MC_{ijk}) \dots\dots\dots (4B.3_3)$$

Where: I_{ijk} = feed intake (kg DM/head/day) as calculated in Section 6.3.2.2
 CP_{ijk} = crude protein content of feed intake expressed as a fraction (Appendix 6.D.4)
 MC_{ijk} = milk intake (kg/head/day) calculated as: proportion of lambs receiving milk in each season x milk intake (Appendix 6.D.6). Milk intake assumed to be 1.6 kg/day for the first three months after the birth of lambs

Nitrogen excreted in faeces (F_{ijk} kg/head/day) is calculated, using functions developed by the SCA (1990) and Freer et al. (1997), as the indigestible fraction of the undegraded protein from solid feed, the microbial crude protein and milk protein plus the endogenous faecal protein, such that:

$$F_{ijk} = \{0.3(CPI_{ijk} \times (1 - [(DMD_{ijk} + 10)/100])) + 0.105(ME_{ijk} \times I_{ijk} \times 0.008) + 0.08(0.045 \times MC_{ijk}) + 0.0152 \times I_{ijk}\} / 6.25 \dots\dots\dots (4B.3_4)$$

Where: DMD_{ijk} = digestibility expressed as a percentage (Appendix 6.D.2)
 ME_{ijk} = metabolizable energy (MJ/kg DM) calculated as $0.1604 DMD_{ijk} - 1.037$ (Minson and McDonald 1987)
 MC_{ijk} = milk intake (kg/day) calculated as: proportion of lambs receiving milk in each season x milk intake (Appendix 6.D.6). Milk intake assumed to be 1.6 kg/day for the first three months after the birth of lambs
 $1/6.25$ = factor for converting crude protein into nitrogen

The amount of nitrogen retained by the body (NR_{ijk} kg/head/day) is calculated as the nitrogen retained in milk, wool and body tissue such that:

$$NR_{ijk} = \{(0.045 \times MP_{ijk}) + (WP_{ijk} \times 0.84) + \{[(212 - 4\{[(EBG_{ijk} \times 1000)/(4 \times SRW_{ijk}^{0.75})] - 1\}) - (140 - 4\{[(EBG_{ijk} \times 1000)/(4 \times SRW_{ijk}^{0.75})] - 1\}) / \{1 + \exp(-6(Z_{ijk} - 0.4))\}] \times EBG_{ijk} / 1000\} / 6.25 (4B.3_5)$$

Where: MP_{ijk} = milk production (kg/day) calculated as: proportion of ewes lactating (LE_{ijk}) x milk production. Milk production is considered to be 1.6 kg/day for breeding ewes in the first three months after the birth of lambs.
 WP_{ijk} = clean wool production (kg/day) based on ABS average greasy wool production per head multiplied by State average clean yield percentage. Wool production may be reduced by 50% for lactating ewes (SCA 1990). Accordingly, wool production of ewes was apportioned pro rata to give recorded annual average wool production. It is assumed that clean wool consists of 16% water and 84% protein.
 EBG_{ijk} = empty body gain which is equivalent to $LWG_{ijk} \times 0.92$
 SRW_{ijk} = standard reference weight (SCA 1990) in Appendix 6.D.7
 Z_{ijk} = relative size (liveweight / standard reference weight)

Nitrogen excreted in urine (U_{ijk} kg/head/day) is calculated by subtracting the nitrogen retained (NR_{ijk}) and the nitrogen excreted in the faeces (F_{ijk}) from the nitrogen intake such that:

$$U_{ijk} = (CPI_{ijk} / 6.25) - NR_{ijk} - F_{ijk} \dots\dots\dots(4B.3_6)$$

The annual faecal (AF_{ijk} Gg) and urinary (AU_{ijk} Gg) nitrogen excreted is calculated as:

$$AF_{ijk} = (N_{ijk} \times F_{ijk} \times 91.25) \times 10^{-6} \dots\dots\dots(4B.3_7a)$$

$$AU_{ijk} = (N_{ijk} \times U_{ijk} \times 91.25) \times 10^{-6} \dots\dots\dots(4B.3_7b)$$

Where: N_{ijk} = the number sheep in each State, season and class

The total emissions of nitrous oxide from the different manure management systems can then be calculated as follows:

$$Faecal_{MMS} = (AF_{ijk} \times MMS \times EF_{(MMS)} \times C_g) \dots\dots\dots(4B.3_8a)$$

$$Urine_{MMS} = (AU_{ijk} \times MMS \times EF_{(MMS)} \times C_g) \dots\dots\dots(4B.3_9b)$$

$$E_{MMS} = \sum_i \sum_j \sum_k (Faecal_{MMS} + Urine_{MMS}) \dots\dots\dots(4B.3_10c)$$

Where: MMS = the fraction of nitrogen that is managed in the different manure management systems. In the case of sheep all manure is voided at pasture (MMS = 8).

$EF_{(MMS)}$ = emission factor (N_2O -N kg/ N excreted) for the different manure management systems (Table 6.12)

C_g = 44/28 factor to convert elemental mass of N_2O to molecular mass

6.4.2.3 Pigs (4.B.8)

Methane

In Australia, pigs are generally housed and the liquid waste slurry produced during cleaning is often channelled into lagoons. These lagoons tend to create anaerobic conditions, resulting in a high proportion of the volatile solids being fermented with the formation of methane.

A significant proportion of feed given to pigs can be wasted (ranging from 5-20%). This waste feed also contributes volatile solids to the manure management system and will result in methane emissions. For completeness emissions are estimated from all waste entering the manure management system.

Casey et al. (1996) have developed a nutrient balance model for intensive piggeries in Australia-PIGBAL. By entering typical animal characteristic, intakes, diet compositions and wastage rates the model calculates the volatile solids in the animal manure and waste feed (Appendix I).

Pigs are fed high quality diets with substantial nitrogen contents. The rapid growth rates of most pigs result in a relatively high proportion of this nitrogen being retained in the body. Pigs may excrete between 45 and 65% of nitrogen consumed in feed (King and Brown 1993, King et al. 1993).

Wasted feed also contributes nitrogen to the manure management systems and is included in the estimation of emissions for completeness. The nutrient balance model PIGBAL (Casey et al. 1996) is used to estimate total nitrogen in wastes based on typical animal characteristics, intakes, feed types and wastage rates (Appendix G).

Methane is formed from the organic fraction (volatile solids) of the faeces and waste feed. Volatile solid production for Australian pigs (VS_{ij} kg/head/day) was estimated using the PIGBAL model (Casey et al. 1996).

Using this information methane production from the wastes (M_{ij} kg/head/day) can thus be calculated as:

$$M_{ij} = VS_{ij} \times B_o \times MCF \times \rho \dots\dots\dots (4B.8_1)$$

Where: VS_{ij} = volatile solids production (kg/head/day) (Appendix 6.E.2)
 B_o = emissions potential ($0.45\text{m}^3 \text{CH}_4/\text{kg VS}$ (IPCC 1997))
 MCF = integrated methane conversion factor based on the proportion of different manure management regimes (Appendix 6.E.3) and MCF values for 'warm' regions for Qld and the NT and MCF values for 'temperate' regions for all other States (Appendix 6.E.4).
 ρ = density of methane (0.662 kg/m_3)

The annual methane production (Gg) from the wastes of Australian pigs is calculated as:

$$E = \sum_i \sum_j (365 \times N_{ij} \times M_{ij} \times 10^{-6}) \dots\dots\dots (4B.8_2)$$

Where: N_{ij} = numbers of pigs in each class for each State
 M_{ij} = methane production (kg/head/day)

Nitrous Oxide

The annual nitrogen (AE_{ij} Gg/year) from pig manure and waste feed is calculated as:

$$AE_{ij} = N_{ij} \times E_{ij} \times 10^{-6} \dots\dots\dots (4B.8_3)$$

Where: N_{ij} = numbers of pigs in each class in each State
 E_{ij} = nitrogen in waste (kg/head/year) as calculated by PIGBAL (Appendix 6.E.5)

The total emissions of nitrous oxide from the different manure management systems (E_{MMS} Gg) can then be calculated as follows:

$$E_{MMS} = \sum_i \sum_j (AE_{ij} \times MMS \times EF_{(MMS)} \times C_g) \dots\dots\dots (4B.8_4)$$

Where: MMS = the fraction of AE that is managed in the different manure management systems (Appendix 6.E.3).
 $EF_{(MMS)}$ = emission factor ($\text{N}_2\text{O-N kg/ N excreted}$) for the different manure management systems (Table 6.12)
 C_g = 44/28 factor to convert elemental mass of N_2O to molecular mass

6.4.2.4 Other Livestock (Including 4.B.2, 4, 5, 6, 7, 9, and 10)

Methane

Poultry

The majority of Australia's poultry population are housed indoors which promotes conditions for the concentration and concentrated treatment of faecal wastes. Methane from manure is formed from the organic fraction of the manure (volatile solids). Volatile solid production for poultry (VS_{ij} kg/head/day) was estimated using information on intakes and dry matter digestibility:

$$VS_{ij} = I_{ij} \times (1 - DMD_{ij}) \times (1 - A) \dots\dots\dots (4B.10_1)$$

Where: I_{ij} = dry matter intake (assumed to be 0.11 kg/day)
 DMD_{ij} = digestibility expressed as a fraction (assumed to be 80%)
 A = ash content expressed as a fraction (assumed to be 8% of faecal DM)

Methane production from the manure (M_{ij} kg/head/day) can then be calculated as:

$$M_{ij} = VS_{ij} \times B_o \times MCF \times \rho \dots\dots\dots(4B.10_2)$$

Where: B_o = emission potential ($0.32m^3$ CH₄/kg VS (IPCC 1997))

MCF = methane conversion factor. IPCC (1997) MCF value for 'warm' regions used for Queensland and the Northern Territory (2%) and MCF values for 'temperate' region used for all other States (1.5%)

ρ = density of methane (0.662 kg/m³)

The annual methane production (Gg) from the manure of Australian poultry is calculated as:

$$E = \sum_i \sum_j (365 \times N_{ij} \times M_{ij} \times 10^{-6}) \dots\dots\dots(4B.10_3)$$

Where: N_{ij} = numbers of all types of poultry

M_{ij} = methane production (kg/head/day)

Other Livestock

Goats, deer, buffalo, camels, alpaca, horses, donkeys, mules, emus and ostriches are range-kept livestock and hence, manure deposition typically occurs in a dispersed fashion. Little is known about the amount of manure produced by the livestock types in this group. In the absence of adequate information, it was assumed that the rates of manure production (DMM_{ij} kg DM/head/year) can be scaled to those calculated for either sheep or beef cattle based on the comparative size of the animals. For example, the default weight for horses (377 kg) and buffalo (380 kg) are consistent with the average weight of beef cattle (380 kg) while the default weight of donkeys (130 kg) and goats (38.5 kg) are consistent with one third of beef cattle (127 kg) and sheep (45 kg) weights respectively. Accordingly, the following assumptions were made:

- goats, alpacas, emus and ostriches are equivalent to those from a sheep (national average 114 kg/head/year)
- deer, donkeys and mules are one-third those of beef cattle (national average 319 kg/head/year);
- horses, buffalo and camels are equivalent to those of beef cattle (national average 957 kg/head/year).

Methane production from the manure (M_{ij} kg/head/day) is calculated as:

$$M_{ijk} = DMM_{ij} \times MEF \dots\dots\dots(4B.3_1)$$

Where: DMM_{ij} = dry matter in manure Appendix 6.F.3

MEF = manure emission factor (kg CH₄ / kg DM Manure) (Gonzalez-Avalos and Ruiz-Suarez 2001). The warm factor is used for Qld and NT and the temperate factor is used for all other States.

The annual methane production (Gg) from the manure of other livestock is calculated as:

$$Total = \sum_i \sum_j (N_{ij} \times M_{ij}) \times 10^{-6} \dots\dots\dots(4B.3_2)$$

Where: N_{ij} = numbers of animals in each State

M_{ij} = methane production (kg/head/day)

Nitrous Oxide

The nitrogen excretion from poultry was calculated based on an average intake for both layers and meat chickens of 0.11 kg DM/head/day, with approximately 19% crude protein for meat chickens and 16.5% for layers. If poultry are assumed to retain 43% of the nitrogen intake (NRC 1994), this gives rates of nitrogen excretion of 0.7 kg N/bird/year for meat chicken and 0.6 kg N/bird/year for layers.

The nitrogen excretion rates and manure management systems of other poultry types (eg. turkeys and ducks) are assumed to be the same as for meat chickens.

- Little is known about the nitrogen intake and nitrogen excretion rates of the remaining livestock types in this group. The IPCC Guidelines (IPCC 1997) provide a single N excretion rate for “other livestock” of 25 kg/head/year. In the absence of adequate species specific information, it was assumed that the rates of nitrogen excretion (E_{ij} kg/head/year) can be scaled to those calculated for either sheep or beef cattle based on the comparative size of the animals. Accordingly, the following assumptions were made:
- goats, alpacas, emus and ostriches are equivalent to those from a sheep (national average 7.0 kg/head/year)
- deer, donkeys and mules are one-third those of beef cattle (national average 13.2 kg/head/year)
- horses, buffalo and camels are equivalent to those of beef cattle (national average 39.5 kg/head/year).

The annual nitrogen (AE_{ij} Gg/year) excreted from these types of livestock is calculated as:

$$AE_{ij} = (N_{ij} \times E_{ij}) \times 10^{-6} \dots\dots\dots (4B.10_4)$$

Where: N_{ij} = numbers in each State

E_{ij} = nitrogen excreted (kg/head/year) (Appendix 6.F.2)

The total emissions of nitrous oxide from the different manure management systems (E_{MMS} Gg) can then be calculated as follows:

$$E_{MMS} = \sum_i \sum_j (AE_{ij} \times MMS \times EF_{(MMS)} \times C_g) \dots\dots\dots (4B.10_5)$$

Where: MMS = the fraction of AE that is managed in the different manure management systems. The manure of all ‘other livestock’ classes with the exception of poultry is voided at pasture (MMS = 8). For poultry 3% is assumed to be voided at pasture (MMS = 8) and 97% into MMS = 6 for meat birds and MMS = 7 for layers.

$EF_{(MMS)}$ = emission factor (N_2O -N kg/ N excreted) for the different manure management systems (Table 6.12)

C_g = 44/28 factor to convert elemental mass of N_2O to molecular mass

6.4.3 Uncertainties and Time Series Consistency

A quantitative assessment of uncertainty was undertaken and uncertainties for manure management were estimated to be in the order of 10%. Further details on the analysis are provided in Annex 7. Time series consistency is ensured by the use of consistent methods and full recalculations in the event of any refinement to methodology.

6.4.4 Source Specific QA/QC

6.4.4.1 Activity data

The Australian Bureau of Statistics (ABS) is the national statistical agency of Australia and is the key provider of activity data for this source category. ABS has in place a range of quality assurance-quality control procedures associated with survey design, data input and consistency checks on the survey results and the aggregated values. Sampling errors are also evaluated.

Data quality used in the inventory is also kept under review by the DCCEE. This source category is also covered by the general QA/QC procedures detailed in Section 1.6. The QC procedure ‘ensuring consistency in data between categories’ is of specific importance for this category. The AGEIS ensures that activity and livestock characterisation data used across multiple categories is entered only once and that intakes or emissions calculated in one category form the input for other categories.

6.4.4.2 Implied Emission Factors

Comparison with IPCC defaults

As country specific tier 2 methods are used to estimate emissions from cattle, sheep, pigs and poultry, the implied emission factors (IEFs) have been compared with the IPCC defaults (Table 6.14).

Table 6.14: Implied emission factors – CH₄ manure management (kg/head/year)

Livestock Type	Australia	IPCC Default (Oceania)	Reason for difference
Dairy Cattle	8.8	32-33	Different waste allocation, CS EF and estimated VS
Non-dairy cattle			
Range-kept	0.04	6-7	CS EF
Feedlot	2.86	6.7	Different waste allocation
Sheep	0.002	0.28-0.37	CS EF
Swine	23	20	Inclusion of waste feed
Poultry	0.02	0.117-0.157	Difference in estimated VS

CS = country specific; EF = emission factor; VS = volatile solids

The IEFs for dairy cattle differ from the IPCC defaults for three reasons: 1) allocation of waste to difference MMS. Australia assumes that at least 90% of waste is voided at pasture (based on average milking times) compared with 76% in the IPCC default; 2) for temperate regions Australia uses a slightly lower MCF for pasture/range than the IPCC default (1% vs 1.5%); and 3) average volatile solid production is lower than the IPCC default (see section 6.4.4.3).

The IEFs for range-kept non-dairy cattle, sheep and all other range kept animals are significantly lower than the IPCC defaults as a country specific emission factor is used (see section 6.4.2).

Trends in IEFs

As part of the standard Tier 1 QC activities, IEFs are checked for possible anomalies in the time-series. As Australia uses tier 2 methods, which disaggregate animals by age class and regions (in which climate ranges from temperate to warm) some inter-annual variability can be expected in the IEFs for cattle, sheep, swine and poultry as the relative numbers in each age class and region change year to year.

6.4.4.3 Volatile solids

The major sources of methane emissions from manure management are from pigs, dairy cattle and feedlot cattle. As the intake calculation for cattle and the volatile solid calculations for pigs differ from the IPCC Tier 2 methodology the estimated volatile solids were compared against the IPCC defaults. These were found to be comparable for pigs and feedlot cattle (Table 6.15). The volatile solid production of dairy cows was lower than the IPCC (1997) defaults for north American and western European dairy cows. The main difference from the defaults and the Australian methodology appear to be in the feed digestibility which is 60% in the IPCC Guidelines. When the average Australian digestibility (75%) is used with the other IPCC default data, volatile solid levels are similar to those in Australia.

Volatile solids for poultry are considerably lower than the IPCC default. The average intake for both layers and meat chicken is considered to be about 110 g DM/head/day with a digestibility of 80%. This gives volatile solids of 0.0202 kg/head/year. The IPCC default of 0.1kg VS/head/day is not feasible as this is almost the complete daily intake of poultry in Australia.

Table 6.15: Volatile Solids (kg/head/day)

	Australia	IPCC defaults
Pigs		
Sows and Gilts (Default – 80kg)	0.4-0.54	0.5
Other Pigs (Default 28-50 kg)	0.22-0.28	0.3
Feedlot Cattle	1.98	2.02
Dairy Cows	3.5	
North American		5.23 (3.8)
Western Europe		5.08 (3.1)
Poultry	0.0202	0.1

Note Volatile solid values in parenthesis are recalculated defaults using 75% digestibility

6.4.4.4 Nitrogen excretion

The country specific estimates of nitrogen excretion were compared against the IPCC defaults (Table 6.16) and those reported by other Parties. Pig and poultry excretion rates are consistent with the IPCC (1997) defaults. Sheep excretion rates are significantly lower than the IPCC default but are comparable with those of other Parties using Tier 2 methods (mean 9 (range 4-21)).

Excretion rates of range kept cattle are at the low range of the IPCC defaults. On reviewing other Parties' excretion rates (mean 51 (range 34-73)) many are estimating excretion rates lower than the IPCC defaults for their region. Australia would expect to be at the low end of the range of excretion rates due to the quality of pasture fed to range-kept cattle. As expected, excretion rates for feedlot cattle were higher than for range-kept animals.

Dairy cattle excretion rates are significantly higher than the IPCC defaults and the rates reported by other Parties. Unfortunately, unlike volatile solids, the IPCC Guidelines do not provide the data on which the default excretion/retention rates are based, so it is impossible to determine whether it is the assumption regarding feed quality causing the difference in excretion rates. This source had thus been identified as an area for future review.

Table 6.16: Nitrogen Excretion Rates (kg/head/year)

	Australia	IPCC defaults
Non-dairy cattle		40-70
Range-kept	40	
Feedlot	85	
Dairy cattle	117	60-100
Sheep	7	12-20
Pigs (sows)	20.7	16-20
Poultry	0.6-0.7	0.6

The current method for calculating nitrogen excretion from dairy cattle assumes that the intake of calves is all from pasture although on average calves are fed milk replacer and feed pellets for six weeks until weaned onto pasture. In response to a recommendation from the 2009 UNFCCC expert review, an analysis was undertaken to determine what effect implementing the pre-weaning feeding regime would have on emissions. This analysis showed that while there would be minor increases in nitrogen excretion and N₂O from the manure management, overall emissions would be slightly lower as enteric fermentation emissions would be reduced. Implementation of the pre-weaning feeding regimes will be undertaken for the 2012 submission.

In response to a recommendation from the 2008 UNFCCC expert review the N excretion rates for horse and donkeys (mules/asses) were reviewed. The tier 1 default N excretion rate for “other livestock” from the IPCC Guidelines (IPCC 1997) is 25 kg/head/year. The excretion rates for horses and donkeys used in the Australian inventory are assumed to be equivalent to rates calculated for beef cattle (39.5 kg) and one third of beef cattle (13.2 kg) respectively based on the comparative sizes of the animals.

The 2006 IPCC Guidelines (IPCC 2006) provide a different tier 1 method for estimating N excretion based on the weight of the animals. The IPCC default weight for horses (377 kg) is consistent with that for Australian beef cattle (380 kg) while the default for donkeys (130 kg) is consistent with one third of beef cattle (127 kg). Application of the default weights to the tier 1 2006 IPCC guidelines method gives N excretion rates of 41.3 and 14.2 for horses and donkeys respectively. These values are consistent with the values used in the Australian inventory. Based on these results it was concluded that no change in the N excretion rates was required as the rates based on Australia feed quality data are likely to produce more accurate estimates than the IPCC defaults.

6.4.4.5 External review

A comprehensive expert peer review of the methodologies, activity data and livestock characterisation data was conducted in 2000-01 involving agricultural experts from industry, government and academia.

6.4.5 Recalculations Since The 2008 Inventory

Emissions for the most recent year in the *agriculture* sector are recalculated each year due to the recalculation of the three-year average of emissions once the third year of data becomes available. In addition there have been updates to preliminary milk production estimates for 2008 and replacement of extrapolated animal numbers for some “other livestock” with interpolated estimates. The allocation of feedlot cattle into different export classes was revised for 2007-2008 based on information reported by the industry.

The net effect of these changes was no change to the 1990 estimate and a 42 Gg CO₂-e increase in the 2008 estimate.

Table 6.17: 4.B Manure Management: recalculation of total CO₂-e emissions: 1990-2008

Year	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1990	2,064	2,064	0	0.0
2000	3,297	3,297	0	0.0
2001	3,358	3,358	0	0.0
2002	3,386	3,386	0	0.0
2003	3,384	3,384	0	0.0
2004	3,438	3,438	0	0.0
2005	3,549	3,549	0	0.0
2006	3,594	3,596	2	0.0
2007	3,488	3,492	4	0.1
2008	3,346	3,388	42	1.2

6.4.6 Source Specific Planned Improvements

The Australian Government is funding research into emission generation from manure management that aim to provide data on emission levels under a variety of conditions. The Reducing Emissions from Livestock Research Program is managed by the Department of Agriculture, Forestry and Fisheries and will deliver a range of outputs by 2012. This project will provide better estimates for indirect emissions from the deposition of ammonium from composted manure and urine deposition as well as nitrous oxide emissions from the deposition of urine.

In response to UNFCCC expert review recommendations:

- intakes and nitrogen excretion rates for dairy calves will be revised in the 2012 submission to reflect pre-weaning feeding regimes.

6.5 SOURCE CATEGORY 4.C RICE CULTIVATION

6.5.1 Source Category Description

Methane is generated during rice growing from the decomposition of plant residues and other organic carbon material in the soil. This generation occurs through microbial action under anaerobic conditions following flooding of the rice crop.

Methane emission rates vary widely, both diurnally in response to immediate environmental factors such as temperature, and also throughout the season in response to crop development and accompanying changes in soil condition. Emission rates are also dependent on more stable factors including soil type and cultivation method (e.g. irrigation regimes, fertiliser application).

All Australian rice is grown under flooded cultivation. Australian rice culture does not have the large inputs of organic matter as rice stubble is usually burnt and mineral fertilisers are used rather than manures.

6.5.2 Methodology

For the purpose of estimating the emission rate of CH₄ from rice grown in Australia, the extensive data compilation of field measurements of CH₄ emissions from rice in Texas, USA made by Sass (1994) and Sass and Fisher (1994) is used. The climate and agronomic practice of growing rice in Australia is similar to that in the USA and it is therefore assumed that the rate of emission of methane from rice fields is the same (J.R. Freney and P.E. Bacon, personal communication).

The range of 0.5 to 5.5 kg CH₄/ha/day given by Sass (1994) and Sass and Fisher (1994) as the summary of extensive measurements in the USA is used in the estimation of CH₄ emissions from rice in Australia. A seasonally weighted average of this data of 1.5 kg CH₄/ha/day or 225 kg CH₄/ha for the 150 day growing season has been selected as the appropriate average emission rate for Australia.

Table 6.18: Symbols used in algorithms for rice cultivation

State (i)	Activity (j)
1 = NSW	1 = rice cultivation
2 = Tas	
3 = WA	
4 = SA	
5 = Vic	
6 = Qld	
7 = NT	
8 = ACT	

Annual production of methane from rice cultivation is calculated as:

$$E_{ij} = A_{ij} \times EF_j \times 10^{-6} \dots \dots \dots (4C_1)$$

Where: E_{ij} = annual emission (Gg)

A_{ij} = area under rice cultivation (ha).

EF_j = emission factor integrated over the whole season (225 kg CH₄/ha).

6.5.3 Uncertainties and Time Series Consistency

A quantitative assessment of uncertainty was undertaken and uncertainties for rice cultivation were estimated to be in the order of 20%. Further details on the analysis are provided in Annex 7. Time series consistency is ensured by the use of consistent methods and full time series recalculations for all refinements to methodology.

6.5.4 Source Specific QA/QC

The Australian Bureau of Statistics (ABS) is the national statistical agency of Australia and is one of the providers of activity data for this source category. ABS has in place a range of quality assurance-quality control procedures associated with survey design, data input and consistency checks on the survey results and the aggregated values. Sampling errors are also evaluated.

Data quality used in the inventory is also kept under review by the DCCEE. This source category is also covered by the general QA/QC procedures detailed in Section 1.6.

6.5.5 Recalculations Since The 2008 Inventory

There have been no recalculations since the 2008 inventory.

Table 6.19: 4.C Rice Cultivation: recalculation of total CO₂-e emissions (Gg), 1990-2008

Year	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1990	490	490	0	0.0
2000	741	741	0	0.0
2001	738	738	0	0.0
2002	589	589	0	0.0
2003	399	399	0	0.0
2004	237	237	0	0.0
2005	341	341	0	0.0
2006	265	265	0	0.0
2007	196	196	0	0.0
2008	43	43	0	0.0

6.5.6 Planned Improvements

All data and methodologies are kept under review.

6.6 SOURCE CATEGORY 4.D AGRICULTURAL SOILS

6.6.1 Source Category Description

The emissions of nitrous oxide from soils arise from microbial and chemical transformations that produce and consume nitrous oxide in the soil. These transformations involve inorganic nitrogen compounds in the soil, namely ammonium, nitrite and nitrate. These nitrogen compounds can be added to the soil through the following processes:

1. the application of inorganic fertilisers
2. the application of animal wastes to pastures
3. biological nitrogen fixation
4. the application of crop residues

5. mineralisation due to cultivation of organic soils
6. atmospheric nitrogen deposition
7. leaching of inorganic nitrogen and subsequent denitrification in rivers and estuaries.

The IPCC (2000) also indicates that application of sewage sludge to agricultural soils can be included if sufficient information is available. In Australia's inventory the nitrous oxide emissions from the land application of sewage sludge are reported in the *waste* sector under *domestic and commercial wastewater handling*.

6.6.2 Methodology

6.6.2.1 Synthetic Fertiliser

In Australia, synthetic nitrogen fertiliser is applied to a wide range of crops and pastures. The major proportion is applied to rainfed (and relatively low yielding [1-3 t/ha]) cereal grains, and to a lesser degree to sown pastures used for intensive grazing systems. The recommended application rates for these systems, at maximum, 80 kg N/ha for cereal crops and approximately 40 kg N/ha for sown pastures is low relative to international practice. More intensive cropping systems, cotton, sugar cane, irrigated summer crops and horticultural crops require higher rates of synthetic nitrogen application ranging up to 300 kg N/ha.

Recent experimental work on the application of fertilisers to different crop types in Australia and internationally has shown large variations from the IPCC default emission factor of 1.25% across different classes of crop and pasture systems. Variation in emission factors with region and cropping system is to be expected. For example, more than 75% of Australia's cereal production is in low-rainfall regions (<550 mm annual rainfall) with porous soils where both nitrification and the denitrification potential is low. Also, in many of these cropping systems the substantial use of nitrogen fertiliser is a recent practice, with nitrogen use increasing extremely rapidly from a low base in the early 1980s. Consequently, in contrast to European and North American agriculture, there has been little accumulation of soil nitrogen from previous cropping which might predispose these soils to substantial background N₂O emission rates. Finally, it is now becoming more apparent that emission factors increase with both the nitrogen application rate and where either irrigation is applied or rainfall exceeds 550mm annually. Consequently, high emission factors are obtained when application rates, high soil moisture content and timing of fertiliser application produce soil nitrate levels substantially in excess of plant demand.

Recent research conducted in Australia by the Universities, State government Departments of Agriculture and the CSIRO through the Australian governments' research programs (Climate Change Research program) has delivered substantial datasets from a range of crops in different climatic regions. In each study, soil emissions of nitrous oxide have been obtained using automated chambers that collect multiple daily samples from each site over a complete 12 month cycle. Research outputs show much lower emission factors for crops grown in low rainfall (<550mm) environments (Table 6.20) than high rainfall (>550mm) environments, which are closer to the 1.25% IPCC default emission factor.

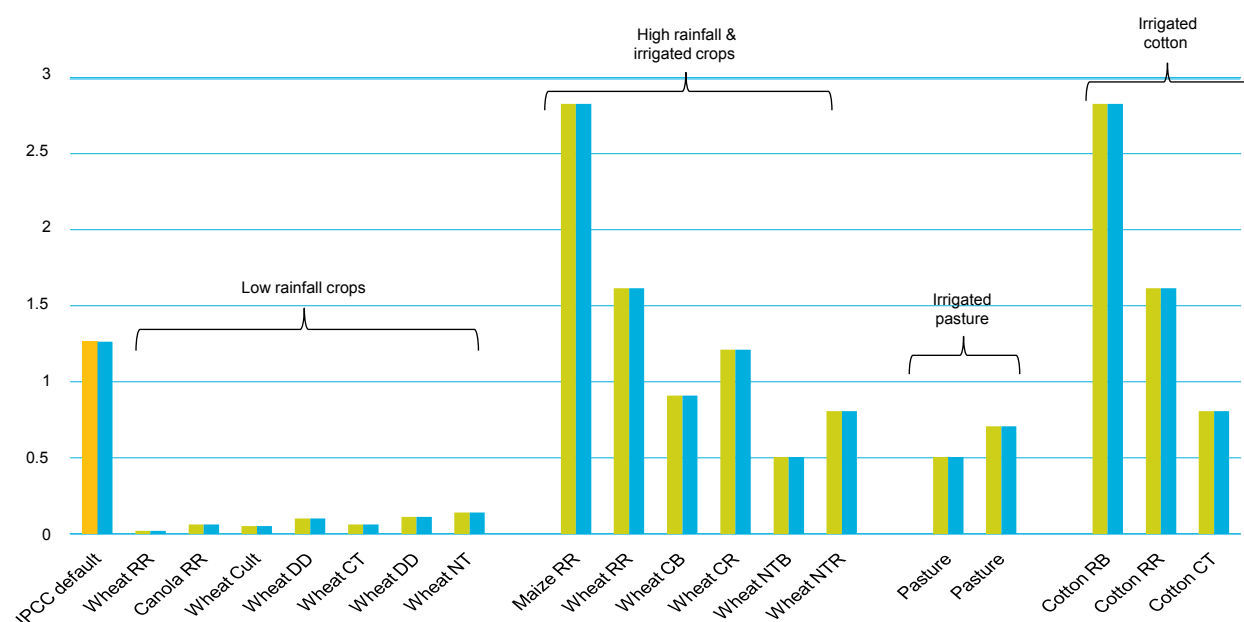
Data presented in Table 6.20 is relatively new research based on several years of intensive experimentation to identify nitrous oxide emissions from cropping systems in Australia. These data can be aggregated into two distinct regions; non-irrigated low rainfall crops and high rainfall and irrigated crops and pastures (Figure 6.3). In this figure, data from wheat (Barker-Reid et al. 2005, Galbally et al. 2005, Barton et al. 2008, Officer et al. 2010) and canola crops (Barton et al. 2010) from low rainfall environments have substantially lower emission factors than the IPCC default of 1.25%. This difference is largely due to low rates of nitrogen fertiliser inputs, low decomposition rates and low levels of microbial activity (Barton et al. 2008). These are summarised in Table 6.20.

Table 6.20: N₂O emission factors from four agricultural systems in Australia

Reference	Site	Crop	Treatment ^(a)	EF(%)
Galbally et al. (2005)	Rutherglen, VIC	Rainfed wheat	Conventional cultivation 83N/ Direct drilling 83N	0.05-0.1
Barker-Reid et al. (2005)	Rutherglen, VIC	Rainfed wheat	Cultivation	0.06
			Direct Drill	0.11
Barton et al. (2008)	Cundedin, WA	Rainfed wheat	Residue retained	0.02
Barton et al. (2010)	Cundedin, WA	Rainfed Canola	Residue retained	0.06
Officer et al. (2010)	Horsham, VIC	Rainfed wheat	Minimum Till	0.14
Irrigated and high rainfall crops				
Galbally et al. (2005)	Griffith, NSW	Irrigated maize	Stubble burning 300N	2.8
			Stubble retention 300N	1.6
Wang et al. (in press)	Warwick, QLD	Rainfed wheat	Cultivated + residue burnt	0.9
			Cultivated + residue retained	1.2
			No till + residue burnt	0.5
			No-till + Residue retained	0.8
Irrigated pastures				
Galbally et al. (2005)	Kyabram, VIC	Irrigated dairy pasture	Urine 1000N	0.4-0.5
			Urea 150N	0.4-0.5
Phillips et al. (2007)	Tatura, VIC	Irrigated pasture	Grazing + fertiliser	0.7
Irrigated cotton				
Galbally et al. (2005)	Narrabri, NSW	Irrigated Cotton (C) in rotation with vetch (V) and wheat (W)	(Rotation sequence) CC 100N	0.03
			CC 200N	0.24
			WVC 100N	0.39
			WVC 200N	0.51
			WVC 300N	2.47
			WC 100N	0.09
			WC 200N	0.26
Grace et al. (2010)	Dalby, QLD	Irrigated cotton	Conventional tillage	0.8

Notes: (a) Annual application rates of fertiliser nitrogen in kg N/ha are listed with the treatments.

Figure 6.3: Emission factors for a range of crops and pastures for rainfed and irrigated systems



Irrigated crops, particularly summer crops (eg. maize), typically receive high rates of N-fertiliser of up to 300 kg N/ha in conditions which favour denitrification. In soils amended by stubble retention higher soil porosity tends to mitigate these rates to some degree. Stubble retention is applied to approximately 50% of irrigated summer row crops, therefore an emission factor of 2.1%, (the average of the incorporated and standard stubble treatments) is used to estimate emissions from irrigated crops and horticultural crops and vegetables. Other high rainfall and irrigated crops and pastures also have typically higher rates of fertiliser N inputs above that of low rainfall crops. For those crops receiving high amounts of water either as rainfall or irrigation the emission factors are also high, sometimes higher than the IPCC default EF (Figure 6.3).

Several studies of both urine and urea application to grazed pastures have been undertaken in Australia (Gabally et al. 1994; 2005). Since the nitrogen content of urine is principally urea there is typically little difference between losses of N_2O from synthetic or animal-source urea (Table 6.20). These local studies indicate that an EF of 0.4% is appropriate for nitrogen fertilisers applied to both irrigated and sown pastures in Australia. Similar emission factors have been reported from well drained silty loam soils in New Zealand (de Klein et al. 2004) where the agricultural practices are similar to Australia.

Average nitrogen application rates for cotton are 200 kg N/ha, therefore an emission factor of 0.5 is considered to be a conservative estimate, based on the ranged presented in Table 6.20, and is used to estimate emissions.

Weier (1999) reported some ^{15}N tracer measurements of N_2O emissions from sugar cane production. He estimated that between 0.06% and 0.28% of synthetic nitrogen applied at a rate of 160 kg/ha was emitted as N_2O , or an average of 0.15%. However this loss was additional to a substantial background emission of $^{14}N_2O$. Taking the worst case in which the total background N_2O emission to be the result of residual fertiliser N from previous crops, the overall emission rate was 0.7% of applied N. Some further Australian studies are in progress, and therefore, in the interim, the IPCC default EF of 1.25% to estimate emissions from sugar cane is retained.

The majority of Australian grain production is from rain-fed cultivation in relatively low rainfall areas where denitrification potential is low. The current studies of rain-fed wheat in dryland regions of Australia including Rutherglen in Victoria, Cundedin in WA, and Horsham in Victoria (Table 6.20), indicate extremely low emission factors ranging from 0.02 to 0.14%. From the disaggregation of crop production areas into leaching, non-leaching and irrigated, as defined in the next section, 81% of rain-fed crop production occurs in dryland areas where this low EF is applicable. Recent publications (Phillips et al. 2010, Grace et al. 2010, Wang et al. *in press*) in the more temperate regions of Australia show higher EFs than the dryland regions and in some cases values that are higher than the IPCC default of 1.25%.

Choice of emission factor

The emission factors used in the inventory are summarised in Table 6.22

The fraction of fertiliser applied to each production system (FN_{ijk}) was determined for each State by first estimating the mass of N-fertiliser applied to irrigated crops, irrigated pasture, cotton, sugar cane and horticulture from the production areas reported by ABS (eg. ABS 2009b) and the average fertiliser application rates for each of these crops, and then distributing the balance of fertiliser use to rain-fed crops and modified pastures (derived from Stewart et al. 2001) in proportion to their respective areas. The fertiliser application rates assigned to irrigated crops, irrigated pastures, cotton, sugar cane and horticultural crops and vegetables are respectively 80 kg N/ha, 40 kg N/ha, 200 kg N/ha, 200 kg N/ha and 125 kg N/ha.

Table 6.21: Symbols used in algorithms for synthetic fertiliser

State (i)	Activity (j)	Subset (k)
1 = NSW	1 = Synthetic fertilisers	1 = Irrigated pasture
2 = Tas		2 = Irrigated crop
3 = WA		3 = Non-irrigated pasture
4 = SA		4 = Non-irrigated crop
5 = Vic		5 = Sugar cane
6 = Qld		6 = Cotton
7 = NT		7 = Horticulture/vegetables
8 = ACT		

Table 6.22: Nitrous oxide emissions factors for synthetic fertiliser

Production System	Emission Factor (Gg N ₂ O-N/ Gg N)
Irrigated pasture	0.004
Irrigated crop	0.021
Non-irrigated pasture	0.004
Non-irrigated crop	0.003
Sugar cane	0.0125
Cotton	0.005
Horticulture/vegetables	0.021

Notes: These EF are based on total fertiliser N applied, not the fraction post volatilisation. Hence equation 4D1_1 differs from the calculation of F_{sn} in equation 4 22 of the IPCC Good Practice Guidance (2000).

The mass of fertiliser applied to soils is calculated as

$$M_{ijk} = T_{Mij} \times FN_{ijk} \dots\dots\dots(4D1_1)$$

Where: M_{ijk} = Mass of fertiliser applied to production system k (Gg N)
 T_{Mij} = Total mass of fertiliser (Gg N)
 FN_{ijk} = Fraction of N applied to production system k (Appendix 6.H)

Annual nitrous oxide production from the addition of synthetic fertilisers is calculated as:

$$E_{ijk} = \sum_i \sum_j \sum_k (M_{ijk} \times EF_{ijk} \times C_g) \dots\dots\dots(4D1_2)$$

Where: E_{ijk} = annual emissions from fertiliser (Gg N₂O)
 M_{ijk} = mass of fertiliser in production system k applied (Gg N)
 EF_{ijk} = emission factor (Gg N₂O-N/Gg N applied) (Table 6.22)
 C_g = 44/28 factor to convert elemental mass of N₂O to molecular mass

6.6.2.2 Animal Wastes Applied to Soils

Nitrous oxide is emitted from soil through the metabolism of animal manure derived principally from dairies, feedlots, piggeries and poultry houses and applied to crops and pastures as organic fertiliser. The current IPCC default emission factor for the N₂O emissions from animal nitrogen deposition to pasture is 2% based on emission studies in which synthetic fertiliser, animal urine and faeces, animal manure or combinations of these were applied to pastures. In 2002, Bouwman et al. published a summary of experimental studies of this type. In reviewing the experiments in which animal manure alone was applied to crops and pastures (Table 6.23) it suggests that while the potential range of emissions is large (0.04%

to 3.3% of applied N), the average is lower than the IPCC default EF. It appears that it is primarily the organic manures which have the higher EFs while wastes from waste management systems and animal houses are less active sources.

The average of the generic class 'organic', sewage sludge and wastes from animal waste management systems (0.4%), gives an emission factor of 1% (0.01 Gg N₂O-N/Gg N deposited) for manure spread on pastures and crops. Inputs to this subsector are calculated in section 6.4.

Table 6.23: N₂O emissions factors (% applied N) for manure applied to crops and pastures

Fertilizer type	Mean	Range
Organic	1.56	0.21 – 3.31
Sewage sludge	0.90	0.80 – 1.00
AWMS effluent	0.40	
<i>Cattle slurry</i>	<i>0.25</i>	<i>0.04 – 0.57</i>
<i>Pig slurry</i>	<i>0.45</i>	<i>0.17 – 0.95</i>
<i>Poultry litter</i>	<i>0.50</i>	<i>0.50 – 0.50</i>

Source: Bouwman et al. 2002

Table 6.24: Symbols used in algorithms for animal wastes

State (i)	Activity (j)	Subset (k)	Manure Management Systems (MMS)
1 = NSW	1 = manure applied to soils	1 = dairy cattle	1 = Anaerobic lagoon
2 = Tas		2 = feedlot cattle	2 = Liquid systems
3 = WA		3 = pigs	3 = Daily spread
4 = SA		4 = other livestock	4 = Solid storage and drylot
5 = Vic			5 = Digester
6 = Qld			6 = Poultry manure with bedding
7 = NT			7 = Poultry manure without bedding
8 = ACT			8 = Pasture range and paddock

The amount of nitrogen applied to soils is the nitrogen excreted adjusted for the nitrogen that has already been lost as N₂O, NH₃ and NO_x during storage in the different manure management systems.

Thus the nitrogen content of animal manure applied to agricultural soils (4D1b) is calculated as:

$$MN_{Soil} = \sum_{MMS=1-7} ((AE_{ij} \times MMS_{=1-7}) - (E_{MMS=1-7} / C_g) - MN_{atmos_{iMMS=1-7}}) \dots \dots \dots (4D1_3)$$

Where: AE = mass of nitrogen excreted as calculated in section 6.4. For dairy and feedlot cattle AE is the sum of faecal (AF) and urinary (AU) nitrogen.

MMS = the fraction of AE that is managed in the different manure management systems.

C_g = 44/28 factor to convert elemental mass of N₂O to molecular mass

E_{MMS=1-7} = total emissions of nitrous oxide from the different manure management systems as estimated in section 6.4

MN_{atmos_{mms}} = mass of animal wastes N volatilised as calculated in section 6.6.2.7 Atmospheric deposition

Annual nitrous oxide production is calculated as:

$$E_{ijk} = \sum_i \sum_j \sum_k (MN \text{ Soil}_{ijk} \times EF_j \times C_g) \dots\dots\dots(4D1_4)$$

Where: E_{ijk} = annual emission from animal waste (Gg N_2O)
 $MN \text{ Soil}_{ijk}$ = mass of nitrogen in manure applied to agricultural soils (Gg N) (as calculated above)
 EF_j = emission factor (Gg N_2O -N/Gg N deposited)
 C_g = 44/28 factor to convert elemental mass of N_2O to molecular mass

6.6.2.3 Nitrogen Fixing Crops

The default IPCC *Tier 1* methodology is used to estimate emissions from N fixing crops and pastures.

Table 6.25: Symbols used in algorithms for N fixing crops

State (i)	Activity (j)	Subset (k)
1 = NSW	1 = N fixing crops	1 = Pulse
2 = Tas		2 = Legume Pasture
3 = WA		3 = Peanuts
4 = SA		
5 = Vic		
6 = Qld		
7 = NT		
8 = ACT		

The mass of N fixed by crops and legume pastures is calculated as:

$$M_{ijk} = P_{ik} \times R_k \times DM_k \times CC_k \times NC_k \dots\dots\dots(4D1_5)$$

Where: M_{ijk} = mass of N fixed by crops and pastures (Gg N)
 P_{ik} = annual production of crop (Gg)
 R_k = residue to crop ratio (kg crop residue/kg crop) (Appendix 6.I)
 DM_k = dry matter content (kg dry weight/kg crop residue) (Appendix 6.I))
 CC_k = mass fraction of carbon in crop residue (Appendix 6.I))
 NC_k = nitrogen to carbon ratio in crop residue (Appendix 6.I))

Annual nitrous oxide production is calculated as:

$$E_j = \sum_i \sum_k (M_{ijk} \times EF_{ijk} \times C_g) \dots\dots\dots(4D1_6)$$

Where: E_j = annual emissions from N fixing crops (Gg N_2O)
 M_{ijk} = mass of N fixed by crops and pastures (Gg N)
 EF_j = 0.0125 (Gg N_2O -N/Gg N) IPCC default emission factor
 C_g = 44/28 factor to convert elemental mass of N_2O to molecular mass

6.6.2.4 Crop Residues

The default IPCC *Tier 1* methodology is used to estimate emissions from crop residues returned to the soil.

Table 6.26: Symbols used in algorithms for crop residues

State (i)	Activity (j)	Subset (k)
1 = NSW	1 = crop residues	1=Wheat
2 = Tas		2=Barley
3 = WA		3=Maize
4 = SA		4=Oats
5 = Vic		5=Rye
6 = Qld		6=Rice
7 = NT		7=Millet
8 = ACT		8=Sorghum
		9=Triticale
		10=Pulse
		11=Tuber and Roots
		12=Sugar Cane
		13=Peanuts
		14=Other Crops

The mass of N in crop residues returned to soils is calculated as:

$$M_{ijk} = P_{ik} \times R_k \times DM_k \times CC_k \times NC_k \times (1 - F_{ik} - FFOD_{ik}) \dots (4D1_7)$$

Where: M_{ijk} = mass of N in crop residues (Gg N)
 P_{ik} = annual production of crop (Gg)
 R_k = residue to crop ratio (kg crop residue/kg crop) (Appendix 6.I)
 DM_k = dry matter content (kg dry weight/kg crop residue) (Appendix 6.I)
 CC_k = mass fraction of carbon in crop residue (Appendix 6.I)
 NC_k = nitrogen to carbon ratio in crop residue (Appendix 6.I)
 F_{ik} = fraction of the crop that is burnt (Appendix 6.I)
 $FFOD_{ik}$ = fraction of the crop that is removed (Appendix 6.I)

Annual nitrous oxide production is calculated as:

$$E_j = \sum_i \sum_k (M_{ijk} \times EF_{ijk} \times C_g) \dots (4D1_8)$$

Where: E_j = annual emissions from crop residues (Gg N_2O)
 M_{ijk} = mass of N in crop residues (Gg N)
 EF_{ijk} = 0.0125 (Gg N_2O -N/Gg N) IPCC default emission factor
 C_g = 44/28 factor to convert from elemental mass of N_2O to molecular mass

6.6.2.5 Cultivation Of Histosols

The default IPCC *Tier 1* methodology is used to estimate emissions from the cultivation of histosols.

The areas of cultivated histosols are very small in Australia occurring only in Queensland where they are mostly used for sugar cane production and small locations of Gippsland and Western Victoria where peatlands were cleared and subsequently grazed or cropped. The individual patches are typically very small, which leads to significant uncertainty when estimating the national area.

Table 6.27: Symbols used in algorithms for cultivation of histosols

State (i)	Activity (j)
1 = NSW	1 = cultivation of histosols
2 = Tas	
3 = WA	
4 = SA	
5 = Vic	
6 = Qld	
7 = NT	
8 = ACT	

Annual nitrous oxide production is calculated as:

$$E_j = \sum_i (A_{ij} \times EF_{ij} \times C_g \times 10^{-6}) \dots\dots\dots(4D1_9)$$

Where: E_j = annual emissions from cultivation of histosols (Gg N_2O)

A_{ij} = area of cultivated histosols (ha)

EF_{ij} = 8 kg N_2O -N/ha. IPCC (2000) default emissions factor for mid-latitude organic soils

C_g = 44/28 factor to convert elemental mass of N_2O to molecular mass

6.6.2.6 Animal Production

Nitrous oxide is emitted from soil through the metabolism of urine and faeces deposited directly on pastures. The current IPCC default emission factor for the N_2O emissions from animal nitrogen deposition to pasture is 2% based on emission studies in which synthetic fertiliser, animal urine and faeces, animal manure or combinations of these were applied to pastures.

There are still relatively few measurements of EFs from animal faeces deposited directly to soil in the absence of urine. Nitrogen in faecal matter is primarily organic and therefore must first be mineralised before it becomes a source of N_2O . In regions of higher rainfall this occurs at significant rates, however in dryland regions, decay of deposited faeces is much slower, with the faeces remaining largely intact for months to years. Approximately 55% of sheep and 65% of beef cattle are grazed in dryland areas. Table 6.28 reports three studies where N_2O emissions from soil to which animal faeces had been added. The average EF, 0.5%, is used to estimate N_2O emissions from animal faecal deposition to soil.

The EF for urine deposition to soil is assumed to be 0.4% of applied N based on experiments conducted on rainfed legume and annual pastures at Book Book, central NSW (Galbally et al. 1994), and irrigated pastures in Kyabram, Victoria (Galbally et al. 2005), as discussed above.

Table 6.28: Animal waste-derived N_2O emissions: (% of N deposited emitted as N_2O -N)

Application	Per cent nitrogen emitted as N_2O	Reference
Faeces		
Cattle faeces	0.5	Flessa et al. (1996)
Dung	0.3	Yamulki and Jarvis (1997)
Animal faeces	0.7	Oenema et al. (1997)

In summary, emission factors of 0.005 and 0.004 Gg N_2O -N/Gg N deposited are considered to be appropriate for nitrous oxide emissions from faeces, and urine voided in the field by grazing livestock respectively. Inputs to this subsector are calculated in section 6.4.

Table 6.29: Symbols used in algorithms for animal wastes

State (i)	Activity (j)	Subset (k)	Manure Management Systems (MMS)
1 = NSW	1 = urine excreted on pastures	1 = dairy cattle	1 = Anaerobic lagoon
2 = Tas	2 = faeces excreted on pastures	2 = beef cattle	2 = Liquid systems
3 = WA		3 = sheep	3 = Daily spread
4 = SA		4 = other livestock	4 = Solid storage and drylot
5 = Vic			5 = Digester
6 = Qld			6 = Poultry manure with bedding
7 = NT			7 = Poultry manure without bedding
8 = ACT			8 = Pasture range and paddock

The amount of nitrogen applied to soils is the nitrogen excreted adjusted for the nitrogen that has already been lost as N_2O , NH_3 and NO_x during storage in the different manure management systems.

Thus the faecal and urinary nitrogen excreted on pasture range and paddock (4D2) is calculated as:

$$FN\ Soil = \sum_{MMS=8} (AF_{ij} \times MMS_{=8}) \dots \dots \dots (4D2_1)$$

$$UN\ Soil = \sum_{MMS=8} (AU_{ij} \times MMS_{=8}) \dots \dots \dots (4D2_2)$$

Where: AF and AU = mass of faecal and urinary nitrogen excreted as calculated in section 6.4. For the 'other livestock' class category AF is calculated as AE x PMF and AU is calculated as AE x PMU.

MMS = the fraction of nitrogen that is managed in the different manure management systems.

AE = total mass of nitrogen excreted

PMF = the proportion of waste that is faeces. Assumed to be 0.29 (based on average of cattle and sheep) for all 'other livestock' classes except poultry, For poultry PMF is assumed to be 1.

PMU = the proportion of waste that is urine. Assumed to be 0.71 (based on average of cattle and sheep) for all 'other livestock' classes except poultry, For poultry PMU is assumed to be 0.

Annual nitrous oxide production is calculated as:

$$E_{ijk} = \sum_i \sum_j \sum_k ((FN\ Soil_{ijk} \times EF_j \times C_g) + (UN\ Soil_{ijk} \times EF_j \times C_g)) \dots \dots \dots (4D2_3)$$

Where: E_{ijk} = annual emission from animal waste (Gg N_2O)

EF_j = emission factor (Gg N_2O -N/Gg N deposited)

C_g = 44/28 factor to convert elemental mass of N_2O to molecular mass

6.6.2.7 Atmospheric Deposition

The default IPCC *Tier 1* methodology is used to estimate emissions from the atmospheric deposition.

Table 6.30: Symbols used in algorithms for atmospheric deposition

State (i)	Activity (j)	Subset (k)
1 = NSW	1 = Atmospheric deposition	1=Synthetic fertiliser
2 = Tas		2=Manure
3 = WA		3=Savanna burning
4 = SA		4=Field burning of Agricultural residues
5 = Vic		
6 = Qld		
7 = NT		
8 = ACT		

The mass of N fertiliser volatilised is calculated as:

$$M_{ijk=1} = TM_{ijk=1} \times \text{FracGASF}_j \dots\dots\dots(4D3_1)$$

Where: $M_{ijk=1}$ = Mass of synthetic fertiliser volatilised (Gg N)
 $TM_{ijk=1}$ = Total mass of fertiliser (Gg N)
 $\text{FracGASF}_j = 0.1$ (Gg N/Gg applied) IPCC default fraction of synthetic fertiliser N that volatilised as $\text{NH}_3\text{-N}$ and $\text{NO}_x\text{-N}$.

The mass of animal waste volatilised is calculated as:

$$M_{ijk=2} = \sum_{\text{MMS}} (\text{AE} \times \text{MMS} \times \text{FracGASM}_{\text{MMS}}) \dots\dots\dots(4D3_2)$$

Where: AE = mass of nitrogen excreted as calculated in section 6.4. For cattle and sheep AE is the sum of faecal (AF) and urinary (AU) nitrogen.
MMS = the fraction of AE that is managed in the different manure management systems.
 $\text{FracGASM}_{\text{MMS}}$ = the fraction of N volatilised in each manure management systems (Table 6.31).

Table 6.31: Fraction of N volatilised in each manure management system (FracGASM)

	Dairy Cattle	Feedlot Cattle	Swine	Other
Anaerobic lagoon	0.35		0.4	
Liquid slurry	0.4		0.48	
Solid storage	0.3		0.45	
Drylot	0.2	0.3	0.45	
Daily spread	0.07		0.07	
Pasture range and paddock	0.2			0.2
Digester			0	
Poultry manure with bedding				0.4
Poultry manure without bedding				0.55

Source. IPCC 2006

The mass of savanna burning and field burning of agricultural residue $\text{NO}_x\text{-N}$ emissions that volatilise is calculated as:

$$M_{ijk=3,4} = (E_{ik=3,4} / (46/14)) \dots\dots\dots(4D3_3)$$

Where: M_{ijk} = mass of N volatilised from savanna burning and field burning of agricultural residues NO_x emissions (Gg N)
 $E_{ik=3,4}$ = NO_x emissions from savanna burning and field burning of agricultural residues (Gg NO_x) as estimated in sections 6.7 and 6.8
 $46/14$ = factor to convert elemental mass of NO_x to molecular mass

Annual nitrous oxide production from atmospheric deposition is calculated as:

$$E = \sum_i \sum_k (M_{ijk} \times \text{EF}_{ijk} \times C_g) \dots\dots\dots(4D3_4)$$

Where: E = annual emissions from atmospheric deposition (Gg N_2O)
 M_{ijk} = mass of N volatilised from subset k (Gg N)
 $\text{EF}_{ijk} = 0.01$ (Gg $\text{N}_2\text{O-N/Gg N}$) IPCC default emission factor
 $C_g = 44/28$ factor to convert elemental mass of N_2O to molecular mass

6.6.2.8 Leaching And Runoff

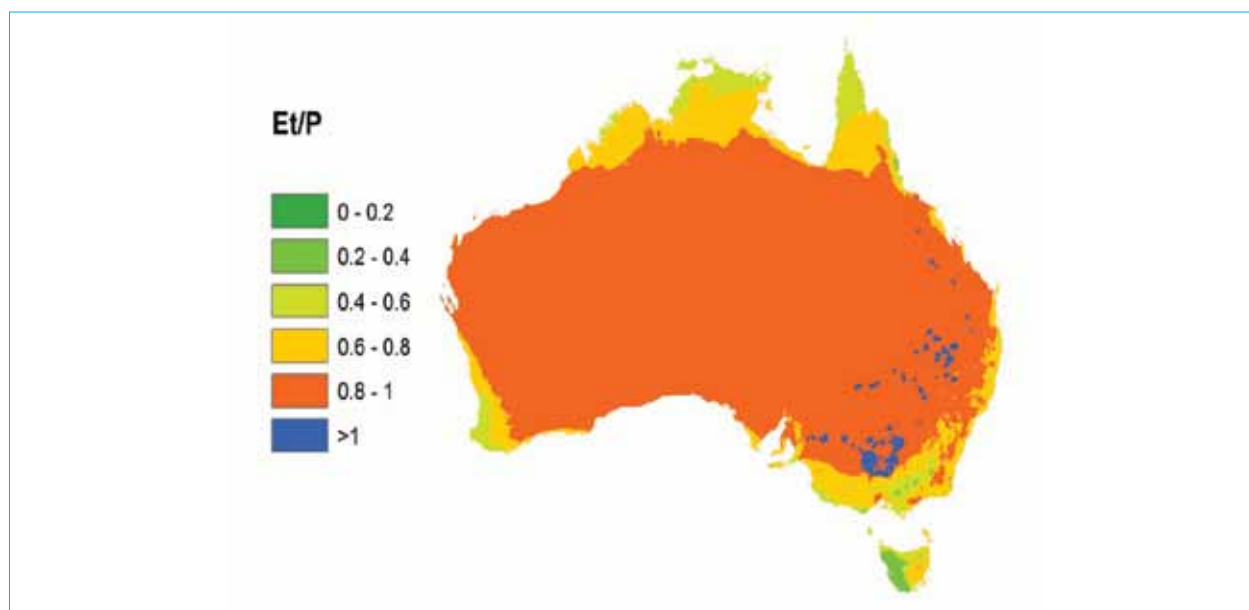
Australia is the driest continent, with substantially less runoff than all other continents. In Australia, much of the cropping takes place in semi-arid regions, or regions of marginal rainfall. The revised 1996 IPCC methodology makes no allowance for reduced rates of leaching in regions of low rainfall and/or high evaporative demand. However, in the 2006 IPCC guidelines it is accepted that leaching of applied nitrogen into waterways and estuaries is unlikely where evaporation exceeds precipitation.

The areas of Australia which are unlikely to be susceptible to significant leaching can be identified using the ratio of evapotranspiration to annual precipitation (Et/P). Evapotranspiration is a better measure than evaporation as it takes into accounts climatic factors (rainfall, humidity, temperature, wind speed) as well as the effect of differences in vegetation classes (forest, shrubland, grassland) on the demand for soil water.

Evapotranspiration has been estimated using the biogeochemical model BIOS (Raupach et al. 2000) for the National Land and Water Audit. Et/P ranges up to 1 where all rainfall is returned to the atmosphere. In areas such as wetlands and irrigation areas in inland regions, where water supply additional to precipitation is available Et/P can exceed 1.

In this methodology, we consider leaching to occur where $Et/P < 0.8$ or $Et/P > 1$ (Figure 6.4). Regions outside these areas are considered to be 'dryland' and not subject to leaching. The fraction of each crop and animal class occurring outside the dryland areas (Frac WET) were determined by overlaying the dryland area mask onto the spatial map of crops, pastures and animal density from the 1997 Agricultural census.

Figure 6.4: The ratio of mean annual evapotranspiration to annual precipitation (Et/P)



The IPCC EF for leaching and runoff incorporates three components which are the factors for groundwater and surface drainage (0.015), rivers (0.0075), and estuaries (0.0025). Recent studies indicated that the previously used factor for groundwater and surface drainage was too high and should be reduced to 0.0025 kg N₂O-N/kg N (Reay et al. 2004, Sawamoto et al. 2005). Therefore, the emission factor has been changed from the IPCC default of 0.025 to 0.0125 kg N₂O-N/kg N leached/runoff as recommended in the 2006 IPCC Guidelines (IPCC 2006).

Table 6.32: Symbols used in algorithms for leaching and runoff

State (i)	Activity (j)	Subset (k)
1 = NSW	1 = Leaching and runoff	1 = Synthetic fertiliser
2 = Tas		2 = Animal waste
3 = WA		
4 = SA		
5 = Vic		
6 = Qld		
7 = NT		
8 = ACT		

The mass of fertiliser N applied to soils that is lost through leaching and runoff is calculated as:

$$M_{ijk=1} = M_{ik} \times \text{FracWET}_{ik} \times \text{FracLEACH}_j \dots\dots\dots(4D3_5)$$

Where: $M_{ijk=1}$ = Mass of synthetic fertiliser lost through leaching and runoff (Gg N)

M_{ik} = mass of fertiliser in each production system (Gg N)

FracWET_{ik} = fraction of N available for leaching and runoff (Appendix 6.J.1)

$\text{FracLEACH}_j = 0.3$ (Gg N/Gg applied) IPCC default fraction of N lost through leaching and runoff. .

The mass of animal waste N applied to soils that is lost through leaching and runoff is calculated as:

$$M_{ijk=2} = (\text{MNsoil}_{ik} + \text{UNsoil}_{ik} + \text{FNsoil}_{ik}) \times \text{FracWET}_{ik} \times \text{FracLEACH}_j \dots\dots\dots(4D3_6)$$

Where: $M_{ijk=2}$ = Mass of animal waste N lost through leaching and runoff (Gg N)

MNsoil_{ik} = mass of manure N applied to soils (Gg N) as calculated in the section 6.6.2.2 Animal Wastes Applied to Soils

UNsoil_{ik} = mass of urinary N applied to soils (Gg N) as calculated in the section 6.6.2.6 Animal Production

FNsoil_{ik} = mass of faecal N applied to soils (Gg N) as calculated in the section 6.6.2.6 Animal Production

FracWET_{ik} = fraction of N available for leaching and runoff (Appendix 6.J.2)

$\text{FracLEACH}_j = 0.3$ (Gg N/Gg applied) IPCC default fraction of N lost through leaching and runoff. .

Annual nitrous oxide production from leaching and runoff is calculated as:

$$E = \sum_i \sum_k (M_{ijk} \times \text{EF}_{ijk} \times C_g) \dots\dots\dots(4D3_7)$$

Where: E = annual emissions from leaching and runoff (Gg N_2O)

M_{ijk} = mass of N lost through leaching and runoff (Gg N)

$\text{EF}_{ijk} = 0.0125$ (Gg $\text{N}_2\text{O-N}$ /Gg N)

$C_g = 44/28$ factor to convert elemental mass of N_2O to molecular mass

6.6.3 Uncertainties and Time Series Consistency

A quantitative assessment of uncertainty was undertaken and uncertainties for agricultural soils were estimated to be in the order of –30 to +50%. Further details on the analysis are provided in Annex 7. Time series consistency is ensured by the use of consistent methods and full time series recalculations for all refinements to methodology.

6.6.4 Source Specific QA/QC

The Australian Bureau of Statistics (ABS) is the national statistical agency of Australia and is the key provider of activity data for this source category. ABS has in place a range of quality assurance-quality control procedures associated with survey design, data input and consistency checks on the survey results and the aggregated values. Sampling errors are also evaluated. Data quality used in the inventory is also kept under review by the DCCEE.

This source category is also covered by the general QA/QC procedures detailed in Section 1.6. The QC procedure 'ensuring consistency in data between categories' is of specific importance for this category. The AGEIS ensures that data used across multiple categories is entered only once and that intakes or emissions calculated in one category form the input for other categories.

6.6.5 Recalculations Since The 2008 Inventory

Emissions for the most recent year in the *agriculture* sector are recalculated each year due to the recalculation of the three-year average of emissions once the third year of data becomes available. In addition there have been a number of improvements and corrections to the activity and input data.

- Animal input and activity data have been revised as described in section 6.4.5.

The net effect of these changes was no change in the 1990 estimate and a decrease of 86 Gg CO₂-e in the 2008 estimate.

Table 6.33: 4.D Agricultural soils: recalculation of total CO₂-e emissions, 1990-2008

Year	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1990	13,438	13,438	0	0.0
2000	16,384	16,384	0	0.0
2001	16,903	16,903	0	0.0
2002	16,511	16,511	0	0.0
2003	16,131	16,131	0	0.0
2004	16,064	16,064	0	0.0
2005	16,007	16,007	0	0.0
2006	15,517	15,517	0	0.0
2007	14,823	14,829	6	0.0
2008	14,557	14,471	-86	-0.6

6.6.6 Planned Improvements

Australian Government research programs such as the Climate Change Research Program managed by the Department of Agriculture, Forestry and Fisheries are delivering key strategies for mitigating nitrous oxide emissions through management practice change. These programs enlist the support of State government agencies, universities, CSIRO and rural industry bodies to conduct robust scientific research across a wide range of farming systems and climatic regions. Currently there are six automatic chamber systems operating across Australia providing fundamental science data on nitrous oxide emissions from crops and pastures. This information provides a solid foundation for the on-going improvement of emissions estimates for national and international reporting as well as providing landholders with a capacity to reduce emissions by changing management practice.

6.7 SOURCE CATEGORY 4.E PRESCRIBED BURNING OF SAVANNAS

6.7.1 Source category description

Savanna burning is a key source of CH₄ and N₂O and is an important source of CO, NO_x and NMVOCs. Carbon dioxide emissions from savanna burning are estimated and reported as stock changes in the grassland remaining grassland category in the LULUCF sector.

The definition of savanna used by IPCC is ‘tropical and sub-tropical formations with continuous grass cover occasionally interrupted by trees and shrubs’ (IPCC, 1997). This includes monsoonal through to semi arid grasslands. In the Australian inventory, this reporting category also includes burning of temperate grasslands in southern Australia and Tasmanian moorland.

Savanna and temperate grassland ecosystems in Australia are burnt either anthropogenically or as a result of wildfires. The anthropogenic burning occurs for a variety of reasons including pasture management, fuel reduction, prevention of uncontrollable wildfires, and traditional Aboriginal burning. These anthropogenic fires replace wildfires that would occur naturally otherwise, albeit later in the dry season, or at other times of the year. Anthropogenic fires probably have occurred in Australian savannas and grasslands since at least 38000 BP (Singh et al. 1981).

It should not be assumed that stopping anthropogenic fires would lead to a reduction in greenhouse gas emissions, because they would be replaced by natural wildfires. In many cases it is impossible to determine whether a fire has been deliberately lit or is the result of a natural ignition process such as a lightning strike. As recommended by the IPCC (1997) all fires in Australian savannas and temperate grasslands are included in the inventory.

6.7.2 Methodology

Stratification of Savanna Types

The inventory methodology stratifies the savanna regions of the Northern Territory (NT) and Western Australia (WA) into two broad vegetation classes of woodlands and grasslands. These are defined in terms of the agroecological zones (AEZs) of the Interim Biogeographic Regionalisation for Australia (IBRA)⁷ version 4.1.

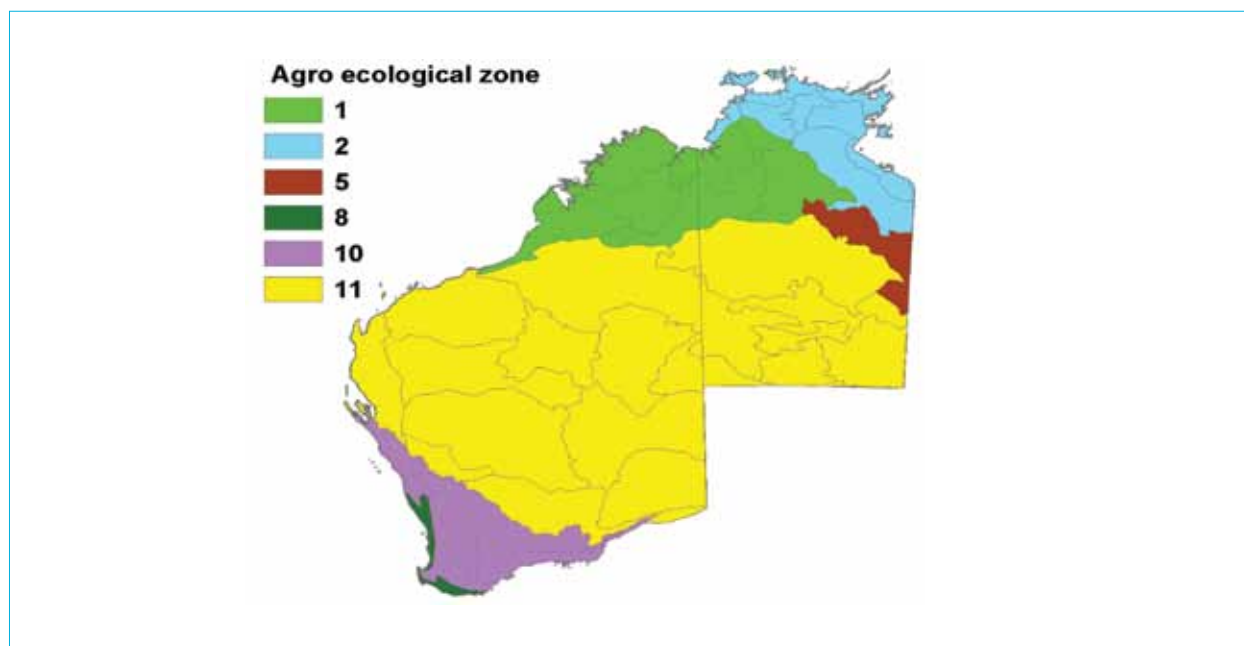
The 11 AEZs are an aggregation of the IBRA regions based mainly on rainfall, soils and climatic conditions, and are used to help identify key land and vegetation management issues at a regional level. The inventory methodology uses groupings of the AEZs as a proxy for classifying regions as grassland or woodland.

The northern part of the NT and WA is characterised by savanna woodland and higher rainfall than the arid centre, and is known as the ‘Top End’. The Top End corresponds approximately to AEZ 1 and AEZ 2, which are predominantly woodland with smaller areas of open forest and grassland (these are under similar management to the surrounding woodland and are not distinguished in this analysis). The inventory methodology uses these two zones as its definition of savanna woodlands. The savanna grasslands of the arid interior of central Australia (‘the Centre’) comprise AEZ 5 and AEZ 11 of the NT and WA, and these zones are used as the inventory definition of grassland. The higher rainfall south-western corner of WA is not classified as savanna, and makes up the remainder of the state (AEZ 8 and AEZ 10). These zones are shown in Figure 6.5.

Savannas in Queensland have not been stratified at this time. All savanna fires are currently reported under the savanna grassland category.

⁷ IBRA is a framework used for sustainable resource management and conservation planning. The 80 IBRA regions represent a landscape-based approach to classifying the land surface from a range of continental data on environmental attributes such as vegetation, geology, soils and climate. Background information and a map of the IBRA regions is available at www.environment.gov.au/parks/nrs/science/bioregion-framework/ibra/index.html.

Figure 6.5: Agroecological zones of NT and WA from IBRA 4.1



Area Burnt

Prior to 1998 the areas of savanna burning for the NT and WA were obtained from a mix of regional fire authority data based on local records or the expert judgment of district officers and high resolution AVHRR satellite imagery. Since 1998 the Western Australian Department of Land Administration (DOLA) has mapped fires for Queensland, Northern Territory, Western Australia and South Australia using the AVHRR imagery

Together the NT and WA jurisdictions make up approximately 80% of the total area of savannas burnt nationally. To improve the time-series for these States, Meyer (2004) supplemented the fire areas measured by regional fire authorities with statistically predicted areas based on rainfall. Regressions on rainfall over the preceding two years were used to predict fire areas for the Northern Territory for 1989 and 1993–1996, and for Western Australia for 1989 and 1993.

With a spatial resolution of 100 hectares, AVHRR imagery provides a major improvement to the reliability of fire data in northern Australia. Firescars are mapped at a resolution of 400 hectares, i.e. 4 pixels. In the southern states (New South Wales, Tasmania and Victoria), however, the areas of individual grass fires are frequently less than 100 hectares and, therefore, statistics recorded by the State fire authorities continue to form the basis of estimates.

Table 6.34: Source of data on area of savanna and temperate grassland burnt

State	ACT, NSW, TAS, VIC	Qld	SA	NT	WA
1989	State agencies	State agencies	State agencies	Meyer 2004	Meyer 2004
1990	State agencies	State agencies	State agencies	AVHRR	AVHRR
1991	State agencies	State agencies	State agencies	AVHRR	AVHRR
1992	State agencies	State agencies	State agencies	AVHRR	AVHRR
1993	State agencies	State agencies	State agencies	Meyer 2004	Meyer 2004
1994–1996	State agencies	State agencies	State agencies	Meyer 2004	AVHRR
1997	State agencies	State agencies	State agencies	AVHRR	AVHRR
1998 +	State agencies	AVHRR	AVHRR	AVHRR	AVHRR

Fuel Loads and Burning Efficiencies

Fuel loads and burning efficiencies are assumed to be constant from year to year and throughout the year, that is, for early, mid and late fire seasons.

When fuel loads are relatively constant throughout the year, emissions can be estimated from annual total areas of savanna and grassland burnt. However in tropical savannas, there are large changes in the amount of grass fuels as the grasses develop, mature and dry through the growing season and the following dry season. Hence the best method for calculating annual emissions from tropical savannas, is to sum the emissions from each fire, taking into account the actual fuel load at the time of the fire. Currently, there is insufficient data to permit this approach.

Temperate grasslands such as the Buttongrass moorland communities in Tasmania and the grasslands in semi-arid Australia do not have such large seasonal variation in standing biomass and fine fuels, and hence emission estimates calculated using the total area of grassland burnt annually are appropriate.

Estimates of fuel loads for the savanna woodlands and grasslands in the NT and WA are drawn from the field measurements in a number of recent studies on grasslands and rangelands (Carter *et al.*, 2000, Dyer *et al.*, 2001, and Carter and Henry, 2003) and from a recently completed study on woodlands in Arnhem Land in the NT (Russell-Smith *et al.*, 2004). Fuel loads in the savanna regions and temperate grassland regions in the remaining States and Territories (ACT, NSW, Queensland, South Australia, Tasmania and Victoria) are based on values in the scientific literature or the experience of State government experts (see Table 6.35).

The burning efficiency is the product of two parameters, the pyrolysis efficiency, i.e. the fraction of fuel exposed to flame that is volatilised, and the fraction of fire scar area that is burnt, i.e. the fire patchiness. The inventory uses separate burning efficiency estimates for savanna grasslands and for woodlands derived from Russell-Smith *et al.* (2004) and Price *et al.* (2003). The inventory estimates are shown in Table 6.36. Meyer (2004) describes the specific derivations of these estimates from the results of Russell-Smith *et al.* (2004) and Price *et al.* (2003). The burning efficiency for the temperate grasslands of 0.72 was provided by Tolhurst (1994) based on measurements of pyrolysis efficiency in grassland in Victoria and the NT and the average patchiness of grass fires.

Table 6.35: Savanna and temperate grassland fuel loads

State	ACT ^(a)	NSW ^(b)	NT ^(c)	Qld ^(d)	SA ^(a)	Tas ^(e)	Vic ^(b)	WA ^(c)
FL _{ij} (Mg/ha)								
Temperate Grassland	11.1	6.9	NO	NO	3.0	9.0	11.7	NO
Savanna Grassland	NO	NO	3.0	3.0	NO	NO	NO	3.0
Savanna Woodland	NO	NO	12.7	IE	NO	NO	NO	12.7

(a) Tolhurst (1994). (b) Chatto (1997). (c) Russell-Smith *et al.* (2004), Meyer (2004). (d) Carter *et al.* (2000), Dyer *et al.* (2001), Carter and Henry (2003). Although not estimated separately the effective fuel load of QLD savanna woodlands is 5.7 (Mg/ha). (e) Marsden-Smedley and Catchpole (1995a, b)

Table 6.36: Savanna and temperate grassland burning efficiencies

Burning Efficiency Z _{ij}	
Temperate Grassland ^(a)	0.72
Savanna Grassland ^(b)	0.76
Savanna Woodland ^(b)	0.40

(a) Tolhurst (1994). (b) Meyer (2004)

Composition of Biomass and Emission Factors

Composition of Biomass

Inventory estimates of the composition of the biomass, i.e. the carbon mass fraction and the nitrogen to carbon ratio, are based on measurements from fires in Australia (Hurst et al. 1994a,b, Henry and Carter 2003, and Russell-Smith et al. 2004).

Meyer (2004) assessed work on Queensland pastures (Henry and Carter 2003) and data for grass fuels in the Northern Territory (Hurst et al., 1994a, and Russell-Smith et al., 2004) and recommended a N:C ratio of 0.012 for temperate and savanna grasslands. Based on studies of woody fuels (Hurst et al. 1994a,b, and Russell-Smith et al. 2004) Meyer (2004) recommended a N:C ratio of 0.011 for savanna woodlands.

Table 6.37: C mass fraction and N:C ratios for savanna and temperate grassland

Carbon mass fraction in dry residue ^(a)	Nitrogen to carbon ratio ^(b)	
	NC _{ij}	
CC _{ij}	Savanna and Temperate Grasslands	Savanna Woodlands
0.46	0.012	0.011

(a) Hurst et al. (1994 a, b). (b) Meyer (2004)

Emission Factors

The emission factors used in the Australian methodology are derived from direct measurement of fire plumes from experimental savanna fires in the Northern Territory (Hurst et al., 1994a,b). The Australian factors are encompassed by those reported for other continents in the international literature (Andreae et al., 1996) although the factor for NO_x is outside the current range specified in the IPCC guidelines (IPCC, 1997).

Table 6.38: Savanna and temperate grassland burning – emission factors

Gas species	Emission factor ^(a)	
	EF _{ij} (Gg element in species/Gg element in fuel burnt)	Elemental to molecular mass conversion factor(C _g)
1. CH ₄	0.0035	16/12
2. N ₂ O	0.0076	44/28
3. NO _x	0.21	46/14
4. CO	0.078	28/12
5. NMVOC	0.0091	14/12

(a) Hurst et al. (1994 a, b)

Table 6.39: Symbols used in algorithms for savanna and temperate grassland burning

State (i)	Activity (j)
1 = NSW	1 = savanna woodland (NT, WA)
2 = Tas	2 = savanna grassland (NT, QLD, WA)
3 = WA	3 = temperate grassland (ACT, NSW, TAS, SA, Vic)
4 = SA	
5 = Vic	
6 = Qld	
7 = NT	
8 = ACT	

For each State, calculate the total mass burnt as:

$$M_{ij} = A_{ij} \times FL_{ij} \times Z_{ij} \times 10^{-3} \dots\dots\dots(4E_1)$$

Where: M_{ij} = mass of fuel burnt in fires (Gg)

A_{ij} = annual area burnt (ha)

FL_{ij} = fuel load (dry weight) (Mg/ha)

Z_{ij} = burning efficiency of fires

then for CH₄, CO and NMVOCs calculate annual emissions:

$$E_{ij} = M_{ij} \times CC_{ij} \times EF_{ij} \times C_g \dots\dots\dots(4E_2)$$

Where: E_{ij} = annual emission from fires

M_{ij} = mass of fuel burnt in fires (Gg)

CC_{ij} = carbon mass fraction in fuel burnt in fires

EF_{ij} = emission factor (Table 6.38)

C_g = factor to convert from elemental mass of gas to molecular mass

and annual emissions for NO_x and N₂O:

$$E_{ij} = M_{ij} \times CC_{ij} \times NC_{ij} \times EF_{ij} \times C_g \dots\dots\dots(4E_3)$$

Where: E_{ij} = annual emission from fires

M_{ij} = mass of fuel burnt in fires (Gg)

CC_{ij} = carbon mass fraction in fires

NC_{ij} = nitrogen to carbon ratio in fuel burnt in fires

EF_{ij} = emission factor (Table 6.38)

C_g = factor to convert from elemental mass of gas to molecular mass

6.7.3 Uncertainties and Time Series Consistency

A quantitative assessment of uncertainty was undertaken and uncertainties for prescribed burning of savannas were estimated to be in the order of –50 to +110%. Further details on the analysis are provided in Annex 7. Time series consistency is ensured by the use of consistent methods and full time series recalculations for all refinements to methodology. Significant effort has been made to ensure time-series consistency in the activity data for this source category given the changes in data sources over time (Meyer 2004).

6.7.4 Source Specific QA/QC

This source category is covered by the general QA/QC procedures detailed in Section 1.6. Activity data and input parameters for this source category were updated in 2005 following a comprehensive review of current literature and experimental data (Meyer 2004).

6.7.5 Recalculations Since The 2008 Inventory

There were no recalculations for this category.

Table 6.40: 4.E Savanna Burning: recalculation of total CO₂-e emissions, 1990-2008

Year	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1990	6,608	6,608	0	0.0
2000	13,202	13,202	0	0.0
2001	15,772	15,772	0	0.0
2002	15,165	15,165	0	0.0
2003	11,262	11,262	0	0.0
2004	11,018	11,018	0	0.0
2005	8,650	8,650	0	0.0
2006	11,478	11,478	0	0.0
2007	11,581	11,581	0	0.0
2008	13,615	13,615	0	0.0

6.7.6 Planned Improvements

The Australian Government is supporting research to improve understanding of emissions outcomes from implementation of fire management practices on northern tropical savannas, including the impact of seasonality on emissions.

Data from this research has become available to underpin new methods for estimating emissions from savanna burning, including the emission factors, vegetation classes, fuel loads and burning efficiencies.

6.8 SOURCE CATEGORY 4.F FIELD BURNING OF AGRICULTURAL RESIDUES

6.8.1 Source Category Description

The burning of residual crop material also releases CH₄, N₂O, CO, NO_x and NMVOCs into the atmosphere. These gases are formed from carbon and nitrogen in the plant material during the combustion process. As per the IPCC Guidelines (IPCC, 1997) the CO₂ emissions from burning of agricultural residues are not included in the inventory total since it is assumed that an equivalent amount of CO₂ is removed by regrowing vegetation in the following year.

Traditionally, burning of agricultural residues in Australia consists of stubble burning (notably for wheat crops), and burning of the sugar cane crop immediately before harvest.

Stubble burning involves firing the standing stalks in either late autumn or spring. Increasingly, as a form of land management, it is being replaced by stubble retention, which reduces erosion and conserves nutrients. In this latter practice the stubble is grazed some weeks after harvest and the next crop is sown by drilling through the remaining vegetation. Firing of sugar cane has also become less common with the rapid introduction of green cane mechanical harvesting. Sugar cane crops are now burnt once every three or four years at the end of the sowing/ratoon cycle.

The amount of crop residue at the time of burning is in most cases less than that at the time of harvest. This applies particularly to crops where there is a long interval between harvest and burning. Vegetation decay and grazing by animals can, over several months, reduce the amount of residue per unit area by one half (R. Jarvis pers. comm., Mulholland et al., 1976). This loss is allowed for in the algorithm.

The amount of material burnt is calculated from the harvested crop yield via information on the residue to crop ratio and the fraction of crop residue remaining at the time of burning etc as presented in Appendix 6.I. This information is derived from Robinson and Kirby (2002), Milthorpe (1982), Saffigna et al. (1982), Galbally et al. (1992), Australian Canegrowers (Milford, pers. comm.) and other agricultural industry sources.

The burning efficiency is estimated to be 96% (Hurst et al., 1994 a, b). The emission factors appropriate to high temperature stubble fires are assumed to be the same as observed in savanna burning (Table 6.41).

Table 6.41: Burning of agricultural residues – emission factors

Gas species	Emission factor ^(a) EF _{ijk} (Gg element in species/Gg element in fuel burnt)	Elemental to molecular mass conversion factor (C _g)
1. CH ₄	0.0035	16/12
2. N ₂ O	0.0076	44/28
3. NO _x	0.2100	46/14
4. CO	0.0780	28/12
5. NMVOC	0.0091	14/12

(a) Hurst et al. (1994 a, b)

Table 6.42: Symbols used in algorithms for burning of agricultural residues

State (i)	Activity (j)	Subset (k)
1 = NSW	1 = burning of agricultural residues	1=Wheat
2 = Tas		2=Barley
3 = WA		3=Maize
4 = SA		4=Oats
5 = VIC		5=Rye
6 = Qld		6=Rice
7 = NT		7=Millet
8 = ACT		8=Sorghum
		9=Triticale
		10=Pulse
		11=Tuber and Roots
		12=Sugar Cane
		13=Peanuts
		14=Other Crops

The mass of fuel burnt is calculated as:

$$M_{ijk} = P_{ik} \times R_k \times S_k \times DM_k \times Z_j \times F_{ik} \dots\dots\dots (4F_1)$$

Where: M_{ijk} = mass of residue burnt from crop (Gg)

P_{ik} = annual production of crop (Gg)

R_k = residue to crop ratio (kg crop residue/kg crop) (Appendix 6.I)

S_k = fraction of crop residue remaining at burning (Appendix 6.I)

DM_k = dry matter content (kg dry weight/kg crop residue) (Appendix 6.I)

Z_j = burning efficiency (fuel burnt/fuel load) for residue from crop

F_{ik} = fraction of the annual production of crop that is burnt (ha burnt/ ha harvested) (Appendix 6.I)

The mass of fuel burnt is converted to an emission of CH₄, CO or NMVOC by multiplying by the carbon content of the fuel, and an emission factor. That is:

$$E_{ijk} = M_{ijk} \times CC_{jk} \times EF_j \times C_g \dots\dots\dots (4F_2)$$

Where: E_{ijk} = annual emission from burning crop residue (Gg)
 CC_{jk} = carbon mass fraction in crop residue
 EF_j = emission factor (Gg element /Gg burnt) (Table 6.41)
 C_g = factor to convert from elemental mass of gas to molecular mass

For N₂O and NO_x an additional term in the algorithm, the nitrogen to carbon ratio (NC_j), is required in order to calculate the fuel nitrogen content. Hence:

$$E_{ijk} = M_{ijk} \times CC_{jk} \times NC_{jk} \times EF_j \times C_g \dots\dots\dots (4F_3)$$

Where: E_{ijk} = annual emission from burning crop residue (Gg)
 CC_{jk} = mass fraction of carbon in crop residue
 NC_{jk} = nitrogen to carbon ratio in crop residue
 EF_j = emission factor (Gg element /Gg burnt) (Table 6.41)
 C_g = factor to convert from elemental mass of gas to molecular mass

6.8.3 Uncertainties and Time Series Consistency

A quantitative assessment of uncertainty was undertaken and uncertainties for the burning of agricultural residues were estimated to be in the order of –40 to +55%. Further details on the analysis are provided in Annex 7. Time series consistency is ensured by the use of consistent methods and full time series recalculations for all refinements to methodology.

6.8.4 Source Specific QA/QC

ABS the principal data supplier has in place a range of quality assurance-quality control procedures associated with survey design, data input and consistency checks on the survey results and the aggregated values. Sampling errors are also evaluated. Data quality used in the inventory is also kept under review by the DCCEE. This source category is also covered by the general QA/QC procedures detailed in Section 1.6.

6.8.5 Recalculations Since The 2008 Inventory

Emissions for the most recent year in the *agriculture* sector are recalculated each year due to the recalculation of the three-year average of emissions once the third year of data becomes available. These changes led to a 8 Gg CO₂-e decrease in the 2008 estimate.

Table 6.43: 4.F Field Burning of Agricultural Residues: recalculation of total CO₂-e emissions (Gg),1990-2008

Year	2010 submission	2011 submission	Change	
	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(Gg CO ₂ -e)	(%)
1990	292	292	0	0.0
2000	410	410	0	0.0
2001	422	422	0	0.0
2002	357	357	0	0.0
2003	365	365	0	0.0
2004	335	335	0	0.0
2005	407	407	0	0.0
2006	339	339	0	0.0
2007	313	313	0	0.0
2008	282	274	-8	-2.7

6.8.6 Planned Improvements

All data and methodologies are kept under review.

APPENDIX 6.A: DAIRY CATTLE

Table 6.A.1: Dairy cattle – liveweight (kg)

State	Milking Cows	Heifers >1	Heifers <1	House Cows Milk and Dry	Dairy Bulls >1	Dairy Bulls <1
NSW/ACT	550	425	240	450	650	300
Tasmania	500	350	220	400	600	250
Western Australia	550	350	180	450	550	250
South Australia	550	450	260	500	500	350
Victoria	550	450	250	450	600	250
Queensland	580	400	150	500	650	200
Northern Territory	500	350	220	400	550	250

Table 6.A.2: Dairy cattle – liveweight gain (kg/day)

State	Milking Cows	Heifers >1	Heifers <1	House Cows Milk and Dry	Dairy Bulls >1	Dairy Bulls <1
NSW/ACT	0.04	0.6	0.6	0.04	0.2	0.9
Tasmania	0.04	0.5	0.8	0.04	0.1	1
Western Australia	0.06	0.8	0.8	0.06	0.1	1
South Australia	0.06	0.5	0.8	0.06	0.1	1
Victoria	0.04	0.5	0.6	0.04	0.1	1
Queensland	0.06	0.7	0.7	0.06	0.1	0.7
Northern Territory	0.06	0.5	0.8	0.06	0.1	1

Table 6.A.3: Dairy cattle – dry matter digestibility of feed intake (%)

State	Milking Cows (%)	Heifers >1 (%)	Heifers <1 (%)	House Cows Milk and Dry (%)	Dairy Bulls >1 (%)	Dairy Bulls <1 (%)
NSW/ACT	75	75	75	75	75	75
Tasmania Spring	75	75	75	75	75	75
Summer	65	65	65	65	65	65
Autumn	65	65	65	65	65	65
Winter	75	75	75	75	75	75
Western Australia	75	75	75	75	75	75
South Australia	75	75	75	75	75	75
Victoria	78	78	78	78	78	78
Queensland	70	65	65	60	65	65
Northern Territory	75	75	75	75	75	75

Table 6.A.4: Dairy cattle – allocation of waste to manure management systems

State	Milking Cows				Other Dairy Cattle
	Lagoon (%)	Liquid/slurry (%)	Daily Spread (%)	Voided at Pasture (%)	Voided at Pasture (%)
NSW/ACT	6.0	0.5	1.5	92.0	100
Tasmania	6.0	0.5	1.5	92.0	100
Western Australia	2.0	0.0	6.0	92.0	100
South Australia	10.0	0.5	1.0	88.5	100
Victoria	6.0	0.5	1.5	92.0	100
Queensland	3.0	0.0	7.0	90.0	100
Northern Territory	3.0	0.0	7.0	90.0	100

Table 6.A.5: Dairy cattle – Methane Conversion Factors (MCF)

State	Manure Management System – Methane Conversion Factors (MCF)				Integrated MCF	
	Lagoon (%)	Liquid/slurry (%)	Daily Spread (%)	Voided at Pasture (%)	Milking Cows (%)	Other Dairy Cattle (%)
NSW/ACT	90	35	0.5	1.0	6.50	1.0
Tasmania	90	35	0.5	1.0	6.50	1.0
Western Australia	90	35	0.5	1.0	2.75	1.0
South Australia	90	35	0.5	1.0	10.07	1.0
Victoria	90	35	0.5	1.0	6.50	1.0
Queensland	90	65	1.0	2.0	4.57	2.0
Northern Territory	90	65	1.0	2.0	4.57	2.0

Table 6.A.6: Dairy cattle – crude protein content of feed intake (%)

State	Milking Cows	Heifers >1	Heifers <1	House Cows Milk and Dry	Dairy Bulls >1	Dairy Bulls <1
NSW/ACT	20	16	16	15	16	16
Tasmania Spring	22	18	18	17	18	18
Summer	18	16	16	15	16	16
Autumn	12	10	10	9	10	10
Winter	20	18	18	17	18	18
Western Australia	20	15	15	14	15	15
South Australia	15	15	15	15	15	15
Victoria	20	16	16	15	16	16
Queensland	18	16	16	15	16	16
Northern Territory	15	15	15	15	15	15

Table 6.A.7: Dairy cattle – standard reference weights (kg)

State	Milking Cows (%)	Heifers >1 (%)	Heifers <1 (%)	House Cows Milk and Dry (%)	Dairy Bulls >1 (%)	Dairy Bulls <1 (%)
NSW/ACT	580	580	580	450	770	770
Tasmania	550	550	550	400	770	770
Western Australia	550	550	550	450	770	770
South Australia	550	550	550	550	770	770
Victoria	550	550	550	450	770	770
Queensland	580	580	580	400	770	770
Northern Territory	550	550	550	400	770	770

Based on SCA 1990

Table 6.A.8: Dairy cattle – Average milk production (kg/head/year)

State	1990	2007	2008
NSW/ACT	3603	5151	5001
Tasmania	3775	4696	4905
Western Australia	4202	5235	5935
South Australia	3934	6417	5813
Victoria	3920	5261	5340
Queensland	3123	4033	4119
Northern Territory	3123	4033	4119

Source: Dairy Australia

APPENDIX 6.B: BEEF CATTLE

Table 6.B.1: Beef cattle – liveweight (kg)

State	Region	Season	Bulls >1 (kg)	Bulls <1 (kg)	Steers <1 (kg)	Cows 1 to 2 (kg)	Cows >2 (kg)	Cows <1 (kg)	Steers >1 (kg)
NSW/ACT		Spring	480	80	75	300	440	75	380
		Summer	520	170	160	360	470	160	420
		Autumn	550	240	220	390	490	220	450
		Winter	560	280	260	410	500	260	460
Tasmania		Spring	700	105	90	300	490	85	480
		Summer	750	480	160	350	530	150	460
		Autumn	725	250	215	360	500	200	490
		Winter	700	260	230	380	460	210	470
Western Australia	South West	Spring	800	340	300	420	550	260	480
		Summer	780	380	340	450	530	300	470
		Autumn	680	100	100	320	480	80	340
		Winter	700	190	170	330	490	150	360
	Pilbara	Spring	450	80	80	260	340	70	370
		Summer	500	150	150	310	360	140	400
		Autumn	550	230	230	330	380	220	420
		Winter	500	250	250	340	360	240	390
	Kimberley	Spring	500	220	210	300	320	180	340
		Summer	550	110	100	220	380	90	390
		Autumn	600	170	160	270	390	140	430
		Winter	550	200	190	280	350	150	400
	South Australia	Spring	800	250	230	400	500	220	420
		Summer	800	320	290	420	500	280	420
		Autumn	700	80	75	300	450	70	400
		Winter	700	160	150	350	450	140	400
Victoria		Spring	820	250	240	410	560	240	510
		Summer	850	280	270	440	550	260	520
		Autumn	700	100	95	300	450	95	410
		Winter	720	150	140	320	470	140	440
Queensland		Spring	750	80	75	300	480	65	350
		Summer	800	170	160	350	500	150	420
		Autumn	760	250	240	390	470	225	470
		Winter	700	320	295	420	450	280	480
Northern Territory		Spring	620	220	210	275	360	190	280
		Summer	650	110	100	240	380	90	260
		Autumn	670	170	160	280	400	140	300
		Winter	660	200	190	290	380	170	285

Table 6.B.2: Beef cattle – liveweight gain (kg/head/day)

State	Region	Season	Bulls >1	Bulls <1	Steers <1	Cows 1 to 2	Cows >2	Cows <1	Steers >1
			(kg/day)	(kg/day)	(kg/day)	(kg/day)	(kg/day)	(kg/day)	(kg/day)
NSW/ACT		Spring	0.2	0.5	0.5	0.4	0.3	0.5	0.4
		Summer	0.4	1.0	0.9	0.7	0.3	0.9	0.4
		Autumn	0.3	0.8	0.7	0.3	0.2	0.7	0.3
		Winter	0.1	0.4	0.4	0.2	0.1	0.4	0.1
Tasmania		Spring	0.50	1.0	1.0	1.0	0.5	1.0	0.50
		Summer	0.55	0.82	0.77	0.55	0.44	0.71	0.5
		Autumn	0.50	0.77	0.6	0.11	-0.33	0.55	0.33
		Winter	-0.27	0.11	0.16	0.22	-0.44	0.11	-0.22
Western Australia	South West	Spring	1.10	1.64	1.42	0.99	0.66	1.21	1.10
Australia	West	Summer	-0.22	0.44	0.44	0.33	-0.22	0.44	-0.11
		Autumn	0.00	0.60	0.60	0.22	-0.55	0.60	0.00
		Winter	0.22	0.99	0.77	0.11	0.11	0.77	0.44
	Pilbara	Spring	-0.55	0.70	0.70	0.22	-0.22	0.70	-0.22
		Summer	0.55	0.77	0.77	0.66	0.55	0.77	0.33
		Autumn	0.55	0.88	0.88	0.22	0.22	0.88	0.22
		Winter	-0.55	0.22	0.22	0.11	-0.22	0.22	-0.33
	Kimberley	Spring	-0.55	0.22	0.22	0.22	-0.33	0.33	-0.66
		Summer	0.55	0.80	0.80	0.44	0.66	0.70	0.55
		Autumn	0.55	0.66	0.66	0.55	0.11	0.55	0.44
		Winter	-0.55	0.33	0.33	0.11	-0.44	0.11	-0.33
South Australia		Spring	1.10	0.99	0.88	0.55	0.55	0.88	0.22
		Summer	0.00	0.77	0.66	0.22	0.00	0.66	0.00
		Autumn	-1.10	0.90	0.80	0.22	-0.55	0.70	-0.22
		Winter	0.00	0.88	0.82	0.55	0.00	0.77	0.00
Victoria		Spring	1.10	1.10	1.10	0.99	0.99	1.10	0.77
		Summer	0.33	0.33	0.33	0.33	-0.10	0.22	0.11
		Autumn	0.20	0.50	0.55	0.44	0.20	0.55	0.20
		Winter	0.22	0.55	0.49	0.22	0.22	0.49	0.33
Queensland		Spring	0.55	1.00	0.90	0.22	0.33	0.80	0.60
		Summer	0.55	0.99	0.93	0.55	0.22	0.93	0.77
		Autumn	-0.44	0.88	0.88	0.44	-0.33	0.82	0.55
		Winter	-0.66	0.77	0.60	0.33	-0.22	0.60	0.11
Northern Territory		Spring	-0.44	0.22	0.22	-0.16	-0.22	0.22	-0.05
		Summer	0.33	0.8	0.8	0.55	0.22	0.8	0.55
		Autumn	0.22	0.66	0.66	0.44	0.22	0.55	0.44
		Winter	-0.11	0.33	0.33	0.11	-0.22	0.33	-0.16

Table 6.B.3: Beef cattle – dry matter digestibility of feed intake (%)

State	Region	Season	Bulls >1	Bulls <1	Steers <1	Cows1 to 2	Cows >2	Cows <1	Steers >1
			(%)	(%)	(%)	(%)	(%)	(%)	(%)
NSW/ACT		Spring	55	55	55	55	55	55	55
		Summer	65	65	65	65	65	65	65
		Autumn	60	60	60	60	60	60	60
		Winter	50	50	50	50	50	50	50
Tasmania		Spring	75	75	75	75	75	75	75
		Summer	60	60	60	60	60	60	60
		Autumn	70	70	70	70	70	70	70
		Winter	75	75	75	75	75	75	75
Western	South	Spring	80	80	80	80	80	80	80
Australia	West	Summer	58	58	58	58	58	58	58
		Autumn	50	50	50	50	50	50	50
		Winter	75	75	75	75	75	75	75
	Pilbara	Spring	40	40	40	40	40	40	40
		Summer	65	65	65	65	65	65	65
		Autumn	55	55	55	55	55	55	55
		Winter	45	45	45	45	45	45	45
	Kimberley	Spring	40	40	40	40	40	40	40
		Summer	65	65	65	65	65	65	65
		Autumn	55	55	55	55	55	55	55
		Winter	45	45	45	45	45	45	45
South		Spring	70	70	70	70	70	70	70
Australia		Summer	55	55	55	55	55	55	55
		Autumn	55	55	55	55	55	55	55
		Winter	75	75	75	75	75	75	75
Victoria		Spring	80	80	80	80	80	80	80
		Summer	55	55	55	55	55	55	55
		Autumn	60	60	60	60	60	60	60
		Winter	76	76	76	76	76	76	76
Queensland		Spring	55	55	55	55	55	55	55
		Summer	65	65	65	65	65	65	65
		Autumn	58	58	58	58	58	58	58
		Winter	52	52	52	52	52	52	52
Northern		Spring	40	40	40	40	40	40	40
Territory		Summer	65	65	65	65	65	65	65
		Autumn	55	55	55	55	55	55	55
		Winter	45	45	45	45	45	45	45

Table 6.B.4: Beef cattle – crude protein content of feed intake (%)

State	Region	Season	Bulls >1	Bulls <1	Steers <1	Cows 1 to 2	Cows >2	Cows <1	Steers >1
			(%)	(%)	(%)	(%)	(%)	(%)	(%)
NSW/ACT		Spring	7	7	7	7	7	7	7
		Summer	13	13	13	13	13	13	13
		Autumn	10	10	10	10	10	10	10
		Winter	6	6	6	6	6	6	6
Tasmania		Spring	20	20	20	20	20	20	20
		Summer	10	10	10	10	10	10	10
		Autumn	16	16	16	16	16	16	16
		Winter	20	20	20	20	20	20	20
Western Australia	South West	Spring	20	20	20	20	20	20	20
		Summer	9	9	9	9	9	9	9
		Autumn	6	6	6	6	6	6	6
		Winter	20	20	20	20	20	20	20
	Pilbara	Spring	4	4	4	4	4	4	4
		Summer	12	12	12	12	12	12	12
		Autumn	9	9	9	9	9	9	9
		Winter	6	6	6	6	6	6	6
	Kimberley	Spring	4	4	4	4	4	4	4
		Summer	12	12	12	12	12	12	12
		Autumn	9	9	9	9	9	9	9
		Winter	6	6	6	6	6	6	6
South Australia		Spring	16	16	16	16	16	16	16
		Summer	7	7	7	7	7	7	7
		Autumn	9	9	9	9	9	9	9
		Winter	20	20	20	20	20	20	20
Victoria		Spring	25	25	25	25	25	25	25
		Summer	7	7	7	7	7	7	7
		Autumn	10	10	10	10	10	10	10
		Winter	21	21	21	21	21	21	21
Queensland		Spring	5	5	5	5	5	5	5
		Summer	15	15	15	15	15	15	15
		Autumn	11	11	11	11	11	11	11
		Winter	6	6	6	6	6	6	6
Northern Territory		Spring	4	4	4	4	4	4	4
		Summer	12	12	12	12	12	12	12
		Autumn	7	7	7	7	7	7	7
		Winter	5	5	5	5	5	5	5

Table 6.B.5: Beef Cattle – feed intake adjustment and milk production and intake

State	Region	Season	Feed adjustment	Milk intake / production (kg/day)
NSW/ACT		Spring	1.3	6
		Summer	1.1	4
		Autumn	0	0
		Winter	0	0
Tasmania		Spring	1.3	6
		Summer	1.1	4
		Autumn	0	0
		Winter	0	0
Western Australia	South West	Spring	0	0
		Summer	0	0
		Autumn	1.3	6
		Winter	1.1	4
	Pilbara	Spring	1.3	4
		Summer	1.1	3
		Autumn	0	0
		Winter	0	0
	Kimberley	Spring	0	0
		Summer	1.3	4
		Autumn	1.1	3
		Winter	0	0
		Spring	0	0
		Summer	0	0
		Autumn	1.3	6
		Winter	1.1	4
Victoria		Spring	0	0
		Summer	0	0
		Autumn	1.3	6
		Winter	1.1	4
Queensland		Spring	1.3	4
		Summer	1.1	3
		Autumn	0	0
		Winter	0	0
Northern Territory		Spring	0	0
		Summer	1.3	4
		Autumn	1.1	3
		Winter	0	0

Table 6.B.6: Beef cattle – standard reference weights

	Bulls >1	Bulls <1	Steers <1	Cows 1 to 2	Cows >2	Cows <1	Steers >1
State	(kg)	(kg)	(kg)	(kg)	(kg)	(kg)	(kg)
NSW/ACT	700	700	600	500	500	500	600
Tasmania	770	770	660	550	550	550	660
Western Australia	770	770	660	550	550	550	660
South Australia	770	770	660	550	550	550	660
Victoria	770	770	660	550	550	550	660
Queensland	770	770	660	550	550	550	660
Northern Territory	770	770	660	550	550	550	660

Based on SCA 1990

APPENDIX 6.C FEEDLOT CATTLE

Table 6.C.1: Feedlot cattle – mean liveweight (kg)

Feedlot Cattle Class/Average time on Feed	1990–1995 ^a	1996+ ^b
domestic/ 75 days	300	360
export/ 140 days	385	490
japan ox/ 250 days	575	565

(a) Working Group estimates, (b) van Slidregt et al. (2000)

Table 6.C.2: Feedlot cattle – intake (kg/day)

Feedlot Cattle Class/Average time on Feed	1990–1995 ^a	1996+ ^b
domestic/ 75 days	7.20	9.8
export/ 140 days	8.47	11.7
japan ox/ 250 days	11.50	11.0

(a) Working group estimates based on the assumption that intake is 2.4%, 2.2% and 2.0% of liveweight for domestic, export and Japanese ox markets respectively. (b) van Slidregt et al. (2000)

Table 6.C.3: Feedlot cattle – proportion of feed components

State	Total grain (incl. molasses)	Other Concentrates	Grasses	Legumes
1990–1995 ^a	0.708	0.072	0.118	0.102
1996+ ^b	0.779	0.048	0.138	0.035

(a) based on working group estimates, (b) van Slidregt et al. (2000).

Table 6.C.4: Feedlot cattle – composition of feed components

Diet Components	Concentrates		Roughage	
	grain	other conc.	grass	legume
Cellulose	0.07	0.19	0.31	0.36
Hemicellulose	0.04	0.11	0.31	0.20
Soluble residue	0.68	0.19	0.21	0.21
Nitrogen	0.02	0.05	0.026	0.032

Based on AFIC 1987 and working group estimates

Table 6.C.5: Feedlot cattle – liveweight gain (kg/day)

Feedlot Cattle Class/Average time on Feed	1990–1995 ^a	1996+ ^b
domestic/ 75 days	1.2	1.7
export/ 140 days	1.2	1.5
japan ox/ 250 days	1.2	1.2

(a) based on working group estimates, (b) van Sliedregt et al. (2000).

APPENDIX 6.D SHEEP

Table 6.D.1: Sheep – liveweight (kg)

State	Season	Sheep > 1				Sheep < 1	
		Rams	Wethers	Maiden Ewes (intended for breeding)	Breeding Ewes	Other Ewes	Lambs & Hoggets
		(kg)	(kg)	(kg)	(kg)	(kg)	(kg)
NSW/ACT	Spring	75	62	44	54	56	20
	Summer	75	55	42	49	51	27
	Autumn	69	55	43	50	50	32
	Winter	69	55	45	50	51	34
Tasmania	Spring	90	55	45	50	50	14
	Summer	90	55	45	50	50	24
	Autumn	75	50	45	50	50	36
	Winter	75	45	50	55	50	42
Western	Spring	75	60	50	55	55	30
Australia	Summer	65	55	45	50	50	30
	Autumn	65	48	40	45	45	10
	Winter	65	48	45	50	50	20
South	Spring	80	70	52	55	55	40
Australia	Summer	70	65	52	55	55	45
	Autumn	70	60	52	55	55	20
	Winter	70	60	52	55	55	30
Victoria	Spring	70	60	50	55	50	22
	Summer	65	55	45	50	50	28
	Autumn	65	52	43	48	50	33
	Winter	60	50	40	45	50	35
Queensland	Spring	58	50	35	40	45	20
	Summer	61	55	40	45	50	25
	Autumn	63	55	40	45	50	20
	Winter	60	50	35	42	48	25

Table 6.D.2: Sheep – dry matter digestibility of feed intake (%)

State	Season	Sheep > 1					Sheep < 1
		Rams	Wethers	Maiden Ewes (intended for breeding)	Breeding Ewes	Other Ewes	Lambs & Hoggets
		(%)	(%)	(%)	(%)	(%)	(%)
NSW/ACT	Spring	75	75	75	75	75	75
	Summer	61	61	61	61	61	61
	Autumn	64	64	64	64	64	64
	Winter	72	72	72	72	72	72
Tasmania	Spring	75	75	75	75	75	75
	Summer	55	55	55	55	55	55
	Autumn	67	67	67	67	67	67
	Winter	70	70	70	70	70	70
Western Australia	Spring	73	73	73	73	73	73
Australia	Summer	55	55	55	55	55	55
	Autumn	50	50	70	70	50	70
	Winter	76	76	76	76	76	76
South Australia	Spring	70	70	70	70	70	70
Australia	Summer	55	55	55	55	55	55
	Autumn	55	55	55	55	55	55
	Winter	75	75	75	75	75	75
Victoria	Spring	70	70	70	70	70	70
	Summer	55	55	55	55	55	55
	Autumn	65	65	65	65	65	65
	Winter	60	60	60	60	60	60
Queensland	Spring	51	51	51	51	51	51
	Summer	55	55	55	55	55	55
	Autumn	59	59	59	59	59	59
	Winter	58	58	58	58	58	58

Table 6.D.3: Sheep – feed availability (t/ha)

State	Season	Sheep > 1					Sheep < 1
		Rams	Wethers	Maiden Ewes (intended for breeding)	Breeding Ewes	Other Ewes	Lambs & Hoggets
		(t/ha)	(t/ha)	(t/ha)	(t/ha)	(t/ha)	(t/ha)
NSW/ACT	Spring	2.90	2.90	2.90	2.90	2.90	2.90
	Summer	2.50	2.50	2.50	2.50	2.50	2.50
	Autumn	1.60	1.60	1.60	1.60	1.60	1.60
	Winter	1.70	1.70	1.70	1.70	1.70	1.70
Tasmania	Spring	2.50	2.50	2.50	2.50	2.50	2.50
	Summer	2.50	2.50	2.50	2.50	2.50	2.50
	Autumn	1.30	1.30	1.30	1.30	1.30	1.30
	Winter	0.80	0.80	0.80	0.80	0.80	0.80
Western	Spring	3.50	3.50	3.50	3.50	3.50	3.50
Australia	Summer	1.50	1.50	1.50	1.50	1.50	1.50
	Autumn	0.70	0.70	0.70	0.70	0.70	0.70
	Winter	1.20	1.20	1.20	1.20	1.20	1.20
South	Spring	4.00	4.00	4.00	4.00	4.00	4.00
Australia	Summer	2.50	2.50	2.50	2.50	2.50	2.50
	Autumn	0.70	0.70	0.70	0.70	0.70	0.70
	Winter	0.90	0.90	0.90	0.90	0.90	0.90
Victoria	Spring	3.20	3.20	3.20	3.20	3.20	3.20
	Summer	3.00	3.00	3.00	3.00	3.00	3.00
	Autumn	1.80	1.80	1.80	1.80	1.80	1.80
	Winter	1.00	1.00	1.00	1.00	1.00	1.00
Queensland	Spring	1.50	1.50	1.50	1.50	1.50	1.50
	Summer	2.00	2.00	2.00	2.00	2.00	2.00
	Autumn	2.20	2.20	2.20	2.20	2.20	2.20
	Winter	1.70	1.70	1.70	1.70	1.70	1.70

Table 6.D.4: Sheep – crude protein content of feed intake (%)

State	Season	Sheep > 1					Sheep < 1
		Rams	Wethers	Maiden Ewes (intended for breeding)	Breeding Ewes	Other Ewes	Lambs & Hoggets
		(%)	(%)	(%)	(%)	(%)	(%)
NSW/ACT	Spring	20	20	20	20	20	20
	Summer	10	10	10	10	10	10
	Autumn	12	12	12	12	12	12
	Winter	18	18	18	18	18	18
Tasmania	Spring	20	20	20	20	20	20
	Summer	7	7	7	7	7	7
	Autumn	14	14	14	14	14	14
	Winter	16	16	16	16	16	16
Western Australia	Spring	18	18	18	18	18	18
Australia	Summer	6	6	6	6	6	6
	Autumn	6	6	16	16	6	16
	Winter	21	21	21	21	21	21
South Australia	Spring	16	16	16	16	16	16
Australia	Summer	7	7	7	7	7	7
	Autumn	9	9	9	9	9	9
	Winter	20	20	20	20	20	20
Victoria	Spring	16	16	16	16	16	16
	Summer	7	7	7	7	7	7
	Autumn	13	13	13	13	13	13
	Winter	10	10	10	10	10	10
Queensland	Spring	8	8	8	8	8	8
	Summer	10	10	10	10	10	10
	Autumn	9	9	9	9	9	9
	Winter	7	7	7	7	7	7

Table 6.D.5: Sheep – liveweight gain (kg/day)

State	Season	Sheep > 1					Sheep < 1
		Rams	Wethers	Maiden Ewes (intended for breeding)	Breeding Ewes	Other Ewes	Lambs & Hoggets
		(kg/day)	(kg/day)	(kg/day)	(kg/day)	(kg/day)	(kg/day)
NSW/ACT	Spring	0.07	0.08	0.07	0.04	0.05	0.16
	Summer	0	-0.08	0.00	-0.05	-0.05	0.08
	Autumn	-0.07	0.00	0.00	0.01	-0.01	0.05
	Winter	0.00	0.00	0.02	0.00	0.01	0.04
Tasmania	Spring	0.16	0.11	0.03	-0.02	0.00	0.15
	Summer	0.00	0.00	0.00	0.00	0.00	0.11
	Autumn	-0.20	-0.10	0.00	0.00	0.00	0.13
	Winter	0	-0.10	0.5	0.02	0.00	0.07
Western	Spring	0.11	0.13	0.05	0.05	0.05	0.11
Australia	Summer	-0.11	-0.05	-0.05	-0.05	-0.05	0.00
	Autumn	0.00	-0.08	0.11	-0.05	-0.05	0.11
	Winter	0.00	0.00	0.05	0.05	0.05	0.11
South	Spring	0.11	0.11	0.00	0.00	0.00	0.11
Australia	Summer	-0.10	-0.10	0.00	0.00	0.00	0.05
	Autumn	0.00	-0.10	0.00	0.00	0.00	0.16
	Winter	0.00	0.00	0.00	0.00	0.00	0.16
Victoria	Spring	0.11	0.11	0.16	0.11	0.00	0.15
	Summer	-0.05	-0.05	-0.05	-0.05	0.00	0.07
	Autumn	0.00	-0.03	-0.02	-0.02	0.00	0.05
	Winter	-0.05	-0.02	-0.03	-0.03	0.00	0.02
Queensland	Spring	-0.02	0.00	0.00	-0.02	-0.03	0.20
	Summer	0.03	0.05	0.05	0.05	0.05	0.05
	Autumn	0.02	0.00	0.00	0.00	0.00	0.20
	Winter	-0.03	-0.05	-0.05	-0.03	-0.02	0.05

Table 6.D.6: Sheep – proportion of lambs receiving milk in each season

State	Spring	Summer	Autumn	Winter
NSW/ACT	0.4	0.1	0.2	0.3
Tasmania	0.6	0	0.1	0.3
Western Australia	0.15	0.1	0.15	0.6
South Australia	0.15	0.05	0.3	0.5
Victoria	0.3	0.1	0.25	0.35
Queensland	0.5	0	0.5	0

Source: Based on breed weighted season of joining (+ 2 seasons) as reported in the MLA 2002 Lamb Survey. Queensland and Tasmania estimates based on information provided by State experts.

Table 6.D.7: Sheep – standard reference weights (kg)

State	Sheep > 1					Sheep < 1
	Rams	Wethers	Maiden Ewes (intended for breeding)	Breeding Ewes	Other Ewes	Lambs & Hoggets
NSW/ACT	78	62	57	57	57	60
Tasmania	77	66	55	55	55	60
Western Australia	84	72	60	60	60	66
South Australia	84	72	60	60	60	66
Victoria	70	60	50	50	50	55
Queensland	70	60	50	50	50	55

Based on SCA 1990

APPENDIX 6.E PIGS

Table 6.E.1: Pigs – feed intake (kg DM/head/day)

Year	Boars	Gilts-intended for breeding	Breeding Sow	Other pigs
1990	2.20	2.20	2.97	1.20
1991	2.20	2.26	2.98	1.23
1992	2.20	2.32	2.98	1.26
1993	2.20	2.38	2.99	1.29
1994	2.20	2.44	2.99	1.32
1995	2.20	2.50	2.99	1.35
1996	2.20	2.56	3.00	1.37
1997	2.20	2.62	3.00	1.40
1998	2.20	2.68	3.01	1.43
1999	2.20	2.74	3.01	1.46
2000+	2.20	2.80	3.02	1.49

Table 6.E.2: Pigs – volatile solids (kg/head/day) entering manure management system

Year	Volatile Solids (kg/head/day)			
	Boars	Gilts-intended for breeding	Breeding Sow	Other pigs
1990	0.38	0.40	0.47	0.22
1991	0.38	0.42	0.47	0.23
1992	0.38	0.43	0.47	0.23
1993	0.39	0.44	0.48	0.24
1994	0.39	0.46	0.48	0.24
1995	0.39	0.47	0.48	0.25
1996	0.39	0.48	0.48	0.26
1997	0.39	0.50	0.49	0.26
1998	0.39	0.51	0.49	0.27
1999	0.39	0.52	0.49	0.28
2000+	0.39	0.54	0.49	0.28

Estimate includes volatile solids from waste feed

Table 6.E.3: Pigs – allocation of waste to manure management systems

State	Manure Management Systems				
	Lagoon	Liquid/slurry	Drylot	Daily Spread	Digester
	(%)	(%)	(%)	(%)	(%)
NSW/ACT	73	0	0	27	0
Tasmania	0	0	0	100	0
Western Australia	65	0	30	5	0
South Australia	73	0	0	27	0
Victoria	71	0	0	27	2
Queensland	92	1.5	5	1.5	0
Northern territory	100	0	0	0	0

Table 6.E.4: Pigs – Methane Conversion Factor (MCF)

State	Manure Management Systems					Integrated MCF
	Lagoon	Liquid/slurry	Drylot	Daily Spread	Digester	
	(%)	(%)	(%)	(%)	(%)	(%)
NSW/ACT	90	35	1.5	0.5	10	65.84
Tasmania	90	35	1.5	0.5	10	0.50
Western Australia	90	35	1.5	0.5	10	58.98
South Australia	90	35	1.5	0.5	10	65.84
Victoria	90	35	1.5	0.5	10	64.24
Queensland	90	65	5.0	1.0	10	84.04
Northern Territory	90	65	5.0	1.0	10	90.00

Table 6.E.5: Pigs – nitrogen (kg /head/year) entering the manure management system

Year	Boars	Gilts-intended for breeding	Breeding Sow	Other pigs
1990	15.29	16.36	20.76	10.37
1991	15.22	15.95	20.75	10.44
1992	15.15	15.53	20.74	10.51
1993	15.08	15.12	20.74	10.57
1994	15.01	14.71	20.73	10.64
1995	14.94	14.30	20.73	10.71
1996	14.87	13.88	20.72	10.77
1997	14.80	13.47	20.71	10.84
1998	14.73	13.06	20.71	10.91
1999	14.66	12.65	20.70	10.98
2000+	14.59	12.23	20.70	11.04

Estimate includes nitrogen in waste feed

APPENDIX 6.F OTHER LIVESTOCK

Table 6.F.1: ‘Other livestock’ – enteric fermentation emission factors (kg CH₄/head/year)

State	Livestock Type								
	Goats	Horses	Deer	Buffalo	Donkeys / Mules	Emus/ Ostriches	Alpacas	Camels	Poultry
All States	5	18	10.7	55	10	5	10	46	NE

Table 6.F.2: ‘Other livestock’ – nitrogen excretion factors (kg N/head/year)

State	Livestock Type									
	Goats	Horses	Deer	Buffalo	Donkeys / Mules	Emus/ Ostriches	Alpacas	Camels	Poultry Meat	Poultry Layers
All States	7.0	39.5	13.2	39.5	13.2	7.0	7.0	39.5	0.7	0.6

Table 6.F.3: ‘Other livestock’ – manure production (kg DM/head/year)

State	Livestock Type							
	Goats	Horses	Deer	Buffalo	Donkeys/ Mules	Emus/ Ostriches	Alpacas	Camels
All States	114	957	319	957	319	114	114	957

APPENDIX 6.G PIGBAL ANALYSIS

The estimation of volatile solids and nitrogen in the waste stream of piggeries was calculated using the nutrient balance model PIGBAL (Casey et al. 1996). Pig industry experts provided information on average intakes and other relevant details for a typical herd in 1990 and 2000. The waste produced for each pig class was then calculated in PIGBAL based on a 1,000 sow herd.

Changes in the herd and feed characteristics are assumed to have occurred gradually from 1990 to 2000. Intake and volatile solids for the years 1991–1999 were, therefore, estimated using a linear regression analysis.

As population statistics are only available for Boars, Gilts, Sow and Other Pigs, weighted averages of intake, volatile solids and nitrogen (as reported in Appendix 6.E) were developed for Sows and Other Pigs based on the herd modelled by PIGBAL.

Table 6.G.1: 1990 feed specification for pigs

Class	Intake (kg/day/pig)	Feed wasted (%)	Crude Protein (%)
Boars	2.20	5.00	13
Gilts	2.20	10.00	18
Gestating sows	2.66	5.00	13
Lactating sows	4.50	5.00	18
Suckers	0.10	20.00	24
Weaner pigs	0.90	15.00	22
Grower pigs	1.50	10.00	21
Finisher pigs	2.20	10.00	18

Table 6.G.2: 2000 feed specifications for pigs

Class	Intake (kg/day/pig)	Feed wasted (%)	Crude Protein (%)
Boars	2.20	5.00	13
Gilts	2.80	10.00	14
Gestating sows	2.66	5.00	13
Lactating sows	5.50	5.00	17
Suckers	0.10	20.00	23
Weaner pigs	1.10	15.00	21
Grower pigs	1.80	10.00	18
Finisher pigs	2.20	10.00	14

Table 6.G.3: Herd characteristics used in PIGBAL

Herd Details		1990	2000
Average lactation length	(days)	28	21
Age out of farrowing accommodation	(days)	28	21
Age out of weaner accommodation	(weeks)	10	10
Age out of grower accommodation	(weeks)	16	16
Age at turnoff	(weeks)	21	24
Age at mating	(weeks)	30	30
Average live pig birth mass	(kg)	1.4	1.4
Average pig weaning mass	(kg)	8.0	6.5
Average pig mass out weaner accom.	(kg)	25	27
Average pig mass out grower accom.	(kg)	55	60
Average pig mass out finisher accom.	(kg)	100	110
Average LW at mating	(kg)	160	170
Average LW at farrowing	(kg)	215	225
Average LW boars at turnoff	(kg)	300	300
Average dressed weight backfatters	(kg)	130	140
Dressing percentage	(%)	76	78
Average dressed weight finishers sold	(kg)	76	86

APPENDIX 6.H SYNTHETIC FERTILISERS

Table 6.H.1: Fraction of fertiliser N applied to each production system, 1990-2008

State	Production System	1990-1995	1996-2000	2001	2002	2003	2004	2005	2006	2007	2008	2009
NSW	Non-irrigated crops	0.3815	0.3815	0.3815	0.3421	0.3143	0.3797	0.3919	0.2610	0.2812	0.4161	0.4161
	Non-irrigated pasture	0.2537	0.2537	0.2537	0.2383	0.2299	0.2227	0.2140	0.1627	0.1885	0.2419	0.2419
	Irrigated crops	0.0892	0.0892	0.0892	0.1178	0.1506	0.1286	0.1161	0.1667	0.1415	0.1234	0.1234
	Irrigated Pasture	0.0352	0.0352	0.0352	0.0422	0.0485	0.0544	0.0555	0.0736	0.0643	0.0448	0.0448
	Sugar	0.0177	0.0177	0.0177	0.0255	0.0354	0.0480	0.0296	0.0368	0.0529	0.0314	0.0314
	Cotton	0.1860	0.1860	0.1860	0.1860	0.1606	0.1081	0.1354	0.2133	0.1671	0.0525	0.0525
	Horticultural Vegetable crops	0.0368	0.0368	0.0368	0.0481	0.0607	0.0584	0.0576	0.0859	0.1045	0.0899	0.0899
	NT Non-irrigated crops	0.0105	0.0105	0.0105	0.0427	0.0537	0.0543	0.0547	0.0021	0.0941	0.0634	0.0634
	Non-irrigated pasture	0.2085	0.2085	0.2085	0.6860	0.7256	0.7340	0.5540	0.0845	0.2935	0.2762	0.2762
	Irrigated crops	0.0256	0.0256	0.0256	0.0058	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
QLD	Irrigated Pasture	0.0208	0.0208	0.0208	0.0047	0.0000	0.0000	0.0000	0.0000	0.0000	0.0130	0.0130
	Sugar	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
	Cotton	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
	Horticultural Vegetable crops	0.7345	0.7345	0.7345	0.2607	0.2208	0.2117	0.3913	0.9134	0.6124	0.6474	0.6474
	Non-irrigated crops	0.0094	0.0094	0.0094	0.0653	0.0558	0.1092	0.0822	0.0754	0.0684	0.1305	0.1305
	Non-irrigated pasture	0.0140	0.0140	0.0140	0.1136	0.1170	0.1751	0.1357	0.1356	0.1686	0.2548	0.2548
	Irrigated crops	0.0287	0.0287	0.0287	0.0312	0.0395	0.0379	0.0274	0.0316	0.0342	0.0634	0.0634
	Irrigated Pasture	0.0193	0.0193	0.0193	0.0165	0.0169	0.0169	0.0148	0.0197	0.0212	0.0222	0.0222
	Sugar	0.6204	0.6204	0.6204	0.5674	0.6230	0.5018	0.5289	0.5174	0.5811	0.4263	0.4263
	Cotton	0.2315	0.2315	0.2315	0.1407	0.0809	0.1018	0.1542	0.1488	0.0562	0.0346	0.0346
NT	Horticultural Vegetable crops	0.0766	0.0766	0.0766	0.0654	0.0669	0.0573	0.0569	0.0715	0.0702	0.0681	0.0681

State	Production System	1990-1995	1996-2000	2001	2002	2003	2004	2005	2006	2007	2008	2009
SA	Non-irrigated crops	0.5443	0.5443	0.5443	0.5285	0.5396	0.5534	0.5241	0.4952	0.5147	0.5147	0.5147
	Non-irrigated pasture	0.3349	0.3349	0.3349	0.3110	0.3042	0.3069	0.2916	0.2925	0.2911	0.2946	0.2946
	Irrigated crops	0.0093	0.0093	0.0093	0.0103	0.0081	0.0053	0.0078	0.0091	0.0066	0.0137	0.0137
	Irrigated Pasture	0.0184	0.0184	0.0184	0.0258	0.0264	0.0266	0.0329	0.0355	0.0324	0.0359	0.0359
	Sugar	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
	Cotton	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
Tas	Horticultural Vegetable crops	0.0931	0.0931	0.0931	0.1244	0.1216	0.1078	0.1436	0.1678	0.1552	0.1411	0.1411
	Non-irrigated crops	0.2410	0.2410	0.2410	0.1457	0.0856	0.1076	0.1038	0.1269	0.0665	0.1170	0.1170
	Non-irrigated pasture	0.3358	0.3358	0.3358	0.5251	0.6594	0.6066	0.6292	0.6105	0.7411	0.6074	0.6074
	Irrigated crops	0.1005	0.1005	0.1005	0.0846	0.0697	0.0668	0.0485	0.0411	0.0272	0.0514	0.0514
	Irrigated Pasture	0.0936	0.0936	0.0936	0.0799	0.0665	0.0798	0.0779	0.0933	0.0801	0.1079	0.1079
	Sugar	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
Vic	Cotton	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
	Horticultural Vegetable crops	0.2290	0.2290	0.2290	0.1647	0.1188	0.1392	0.1406	0.1283	0.0850	0.1163	0.1163
	Non-irrigated crops	0.2896	0.2896	0.2896	0.3153	0.3285	0.3378	0.3463	0.3087	0.3596	0.3634	0.3634
	Non-irrigated pasture	0.3976	0.3976	0.3976	0.4164	0.4178	0.4045	0.4026	0.3980	0.4573	0.3965	0.3965
	Irrigated crops	0.0276	0.0276	0.0276	0.0276	0.0301	0.0312	0.0229	0.0296	0.0132	0.0217	0.0217
	Irrigated Pasture	0.1707	0.1707	0.1707	0.1372	0.1207	0.1290	0.1354	0.1487	0.0773	0.0945	0.0945
WA	Sugar	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
	Cotton	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
	Horticultural Vegetable crops	0.1145	0.1145	0.1145	0.1034	0.1029	0.0975	0.0928	0.1150	0.0926	0.1239	0.1239
	Non-irrigated crops	0.5830	0.5884	0.5884	0.5862	0.5855	0.6013	0.6101	0.5819	0.5514	0.5772	0.5772
	Non-irrigated pasture	0.3845	0.3882	0.3882	0.3909	0.3948	0.3796	0.3732	0.3897	0.4264	0.3982	0.3982
	Irrigated crops	0.0022	0.0013	0.0013	0.0007	0.0000	0.0002	0.0007	0.0026	0.0003	0.0015	0.0015
	Irrigated Pasture	0.0031	0.0019	0.0019	0.0010	0.0000	0.0014	0.0011	0.0024	0.0020	0.0025	0.0025
	Sugar	0.0000	0.0030	0.0030	0.0032	0.0030	0.0030	0.0025	0.0040	0.0030	0.0017	0.0017
	Cotton	0.0000	0.0005	0.0005	0.0003	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
	Horticultural Vegetable crops	0.0273	0.0167	0.0167	0.0178	0.0168	0.0145	0.0124	0.0192	0.0170	0.0189	0.0189

APPENDIX 6.I CROP AND PASTURE ATTRIBUTES

Table 6.I.1: Crop and pasture attributes

Crop type	Residue: Crop ratio	Fraction of residue remaining at time of burning	Dry matter content	Carbon mass frac- tion in dry matter	Elemental N:C ratio (by mass)	Fraction burnt	Fraction Removed
k	R _k	S _k	DM _k	CC _k	NC _k	F _{ik}	FFOD _{ik}
Wheat	1.50	0.5	0.9	0.40	0.008	0.23	0.09
Barley	1.50	0.5	0.8	0.40	0.008	0.23	0.09
Maize(a)	1.50	1.0	0.8	0.42	0.0179	0.3	0.09
Oats	1.50	0.5	0.8	0.40	0.008	0.23	0.09
Rye	1.50	0.5	0.8	0.40	0.008	0.23	0.09
Rice (a)	1.31	1.0	0.8	0.42	0.0155	0.815	0.06
Millet	1.50	0.5	0.8	0.40	0.008	0.23	0.09
Sorghum	1.50	0.5	0.8	0.40	0.008	0.23	0.09
Triticale	1.50	0.5	0.8	0.40	0.008	0.23	0.09
Pulses	2.1	0.5	0.8	0.4	0.05	0.24	0.09
Tuber and Roots	NA	NA	NA	NA	NA	0	1
Sugar cane	0.25	1.0	0.2	0.40	0.025	Table 6.I.2	(b)
Peanuts	1	0.5	0.8	0.40	0.05	0	0.09
Other Crops	1.5	0.5	0.8	0.40	0.008	0.08	0.05
Legume Pastures	1	NA	0.87	0.40	0.08	NA	NA

(a) Robinson and Kirby 2002; (b) 0.03 for QLD and zero for WA and NSW

Table 6.I.2: Fraction of sugar cane burnt in each State

Year	NSW	WA	QLD	AUSTRALIA
1989	1.000	0.000	0.735	0.750
1990	0.978	0.000	0.686	0.701
1991	0.987	0.000	0.664	0.680
1992	0.987	0.000	0.639	0.662
1993	0.987	0.000	0.641	0.663
1994	0.965	0.000	0.596	0.618
1995	0.949	0.000	0.585	0.606
1996	0.975	1.000	0.505	0.532
1997	0.976	1.000	0.430	0.465
1998	0.951	1.000	0.404	0.439
1999	0.951	1.000	0.307	0.354
2000	0.928	1.000	0.346	0.389
2001	0.920	1.000	0.390	0.434
2002	0.897	1.000	0.357	0.398
2003	0.884	1.000	0.331	0.372
2004	0.915	1.000	0.329	0.374
2005	0.963	1.000	0.306	0.348
2006	0.975	1.000	0.282	0.325
2007	0.947	1.000	0.434	0.472
2008	0.947	1.000	0.271	0.314
2009	0.733	1.000	0.263	

APPENDIX 6.J NITROGEN LEACHING AND RUNOFF

Table 6.J.1: Fraction of fertiliser N available for leaching and runoff (FracWET)

	NSW	NT	Qld	SA	Tas	Vic	WA
Irrigated Pasture	1.000	1.000	1.000	1.000	1.000	1.000	1.000
Irrigated crops	1.000	1.000	1.000	1.000	1.000	1.000	1.000
Non-irrigated pasture	0.334	0.811	0.128	0.708	0.991	0.855	0.508
Non-irrigated crops	0.192	0.777	0.043	0.279	0.985	0.438	0.223
Sugar	0.990		0.656				0.759
Cotton (in rotation with wheat and vetch)	0.246		0.075				0.759
Horticultural Vegetable crops	0.599	0.857	0.293	0.667	0.996	0.702	0.911

Table 6.J.2: Fraction of animal waste available for leaching and runoff (FracWET)

State	Region	Dairy Cattle	Free range beef cattle	Feedlot cattle	Sheep	All other categories
ACT		0.96	0.785	0	0.812	0.665
NSW		0.646	0.365	0.192	0.269	0.335
NT		1	0.237	0	0	0.773
QLD		0.348	0.114	0.043	0.018	0.107
SA		0.721	0.691	0.279	0.516	0.415
TASMANIA		1	0.997	0	0.987	0.995
VIC		0.9	0.914	0.438	0.873	0.768
WA		0.996		0.223	0.51	0.668
	South West		0.823			
	Pilbara		0.089			
	Kimberley		0.381			

